

**BIODIVERSITY CHANGES IN THE DERWENT ESTUARY:
A Temporal Analysis of Intertidal and Subtidal Reefs (2010–2024)**

Charlotte McAneney, Ashlee Bastiaansen and Neville
Barrett

December 2025



Derwent Estuary
Program



The Derwent Estuary is one of Tasmania's most ecologically and economically significant waterways, supporting a wide range of biodiversity, recreational and commercial uses, and critical habitats for threatened species such as the Spotted Handfish. Despite its value, the estuary has experienced longstanding pressures from urbanisation, industrial pollution, introduced species, and climate change.

This project forms part of the Derwent Estuary Program's long-term monitoring initiative. It assessed changes in the abundance and distribution of fish, invertebrates, and algal communities on intertidal to subtidal rocky reefs and adjacent sediment habitats, using data from 27 sites surveyed in both 2010 and 2024. Two complementary underwater visual survey methods were employed to capture temporal and spatial changes across habitats and depths.

Key findings indicate that while species richness fluctuated across taxa, several positive ecological shifts have occurred, particularly in algal communities and cryptic fish populations. Notably, *Ecklonia radiata*, a canopy-forming kelp and indicator of water clarity, expanded both in abundance and depth range, coinciding with the disappearance of widespread tubeworm matting and suggesting a shift towards healthier subtidal habitats.

In the fish community, overall richness declined slightly, but community composition remained broadly consistent. Several species exhibited range expansions, while others, such as *Latridopsis forsteri* and *Trachinops caudimaculatus*, declined in both abundance and spatial extent. The marked decline in *L. forsteri* may be at least in part related to changing social behaviour, related to an increase in the popularity of recreational spearfishing. Cryptic fish richness and abundance increased significantly at several upper estuary sites, potentially reflecting improving environmental conditions and habitat complexity.

Mobile macro-invertebrate richness remained stable overall. However, introduced species such as the seastars *Asterias amurensis* and *Patiriella regularis*, as well as the red seaweed *Grateloupia turuturu* increased sharply in abundance and remain a concern for ecological balance. In contrast, other potential invasive species (*Carcinus maenas*, *Metacarcinus novaezelandiae*, and *Petrolisthes elongatus*) showed minimal range expansion, indicating no further escalation in threat level since 2010.

With the exception of the algae, introduced species were most dominant in the upper to mid estuarine sites surveyed, whereas native biodiversity was higher in the southern, more marine-influenced zones. While introduced species increased in number and distribution, particularly among cryptic fish and invertebrates, their presence did not appear to further displace native species in a measurable way over the decade examined here.

Overall, the findings indicate areas of concern—especially rising invasive species numbers—but also evidence of ecological recovery. The expansion of kelp forests, reduced coverage of tube-worm matting, and increased cryptic fish abundance in previously degraded areas point to improving estuarine conditions. Continued monitoring will be essential to track these trends and guide future management actions.

Extended Summary

The Derwent Estuary has been identified as one of Tasmania's most ecologically and economically important estuaries. Approximately 40% of Tasmania's population resides along its margins, and the estuary supports a wide range of recreational activities, including boating and fishing, as well as industry, including Nyrstar Hobart zinc smelter, Boyer paper mill, Incat shipbuilding, TasPorts, and marine transportation. Ecologically, the estuary provides vital habitat for several threatened and endemic species, including the critically endangered Spotted handfish, highlighting its importance as both biodiversity rich and a focus for conservation efforts. Due to urbanisation and historic industrial uses, many pressures affect the Derwent Estuary, including heavy metal contamination, introduced and invasive species, elevated levels of nutrients and climate change. These pressures, when combined, can significantly deteriorate estuarine habitats and the species found within this dynamic ecosystem. Subsequently, assessing biodiversity through time can provide insight into both the recovery of the estuary and its environmental conditions, and the various ecological elements (e.g., invasive species) that threaten a balanced ecosystem.

The Derwent Estuary Program (DEP) is a regional not-for-profit partnership between the Tasmanian Government, local governments, industry, scientists and the community, underpinned by a comprehensive integrated monitoring program documenting environmental conditions and trends. This project contributes to the broader DEP monitoring of the estuary by examining decadal patterns of biodiversity and abundance of fish, invertebrates and algae across intertidal to subtidal rocky reefs and adjacent sediment habitats in the Derwent Estuary. An earlier ARC and DEP funded study examined the biota at 27 sites in 2010, with this study replication that sampling in 2024, using two quantitative survey methods (underwater visual census [UVC] techniques) to assess patterns of abundance and distribution along the estuarine gradient. This report synthesises the results from this sampling, providing an overview of temporal changes observed in the estuary, and positive signs of recovery within the ecosystem. The key findings from both survey methods are presented below. For survey purposes, sites were grouped into three regions along the estuarine gradient: the upper estuary (Cadbury Point, site 1, to Cornelian Bay, site 6), the mid estuary (200 m north of Tasman Bridge East Shore, site 7, to Battery Point, site 15), and the lower estuary (Tranmere Reef, site 16, to Pearsons Point, site 27).

The first survey method utilised a standard belt transect adopted by the Australian Temperate Reef Collaboration (ATRC). A total of 40 fish, 21 cryptic fish, 37 macro-invertebrates and 71 algal species were recorded along standard subtidal rocky reefs across the Derwent Estuary. Compared to the 2010 baseline surveys, species richness fluctuated across taxa in 2024: fish richness declined, whilst cryptic fish and macro-invertebrate richness increased. Fish species richness generally followed the estuarine gradient, increasing from north to south, with the highest richness recorded at lower estuary sites such as Blackmans Bay South, Lucas Point, and Pearsons Point.

Fish community composition remained broadly consistent with 2010, with seven species overlapping amongst the most abundant in both surveys: *Dinolestes lewini*, *Forsterygion varium*, *Latridopsis forsteri*, *Notolabrus tetricus*, *Pempheris multiradiata*, *Pictilabrus laticlavius*, and *Trachinops caudimaculatus*. Of concern is the marked decline of *L. forsteri*, potentially due to a combination of sequential recruitment failure and increased fishing

pressure, particularly via a marked increase in recreational spearfishing. Several species exhibited minor range shifts. *P. multiradiata* and *P. laticlavus* extended further north into the mid-estuary in 2024, while the introduced *F. varium* expanded into the lower estuary. Most abundant species, except *T. caudimaculatus*, *F. varium*, and *Neodaax balteatus*, showed increasing abundance towards the southern part of the estuary. Cryptic fish species richness and overall abundance both increased in 2024, suggesting improved or less extreme environmental conditions and greater availability of complex habitats, which are preferred by cryptic species. Notably, richness and abundance rose substantially at several upper estuary sites – Bedlam Walls North, South, and Point South of Geilston Bay – contrasting with 2010, when higher richness was concentrated in the mid-estuary. Two species, *F. varium* and *P. multiradiata*, were consistently among the four most abundant cryptic fish species in both surveys. While both were also detected in the standard surveys, cryptic surveys offered a more comprehensive assessment of their distribution. The introduced *F. varium* tripled in abundance and extended its range into both the upper and lower estuary, indicating ongoing range shifts. Likewise, *Forsterygion gymnotum* expanded its range, occurring at nine sites in 2024, compared to only two in 2010.

Species richness of mobile macro-invertebrates was comparable to 2010, with no clear increases or spatial trends observed across the estuary. *Patiriella regularis* remained the most abundant macro-invertebrate species, with considerably higher numbers recorded in 2024 – more than double the 2010 counts at many sites. *Asterias amurensis* also showed a marked increase in abundance. However, both species exhibited minimal range shifts, with distribution largely confined to the upper and mid estuary. In the absence of any major flooding events in recent years, the upper estuary may have entered a state of relative stability, allowing opportunistic species like *A. amurensis* and *P. regularis* to flourish in high abundance. In contrast, other abundant macro-invertebrates such as *Heliocidaris erythrogramma*, *Jasus edwardsii*, *Lunella undulata*, *Cenolia trichoptera*, and *Meridiastra calcar* were predominantly recorded in the southern and mid-estuary regions. The significant numbers recorded for both *A. amurensis* and *P. regularis* are of concern, and highlight the need for ongoing monitoring of invasive species (such as this decadal study) and management interventions where practical (e.g. *A. amurensis* removal in core spotted handfish habitat). In contrast, the introduced *Carcinus maenas*, *Metacarcinus novaezelandiae* and *Petrolisthes elongatus* showed no significant increases in abundance, and exhibited minimal range shifts – a positive outcome considering the concerns raised in 2010 about their potential to become invasive threats.

Overall, the distribution patterns of native and introduced cryptic fish and macro-invertebrate species in 2024 were broadly consistent with those observed in 2010. Introduced species richness remained highest in the upper estuary, while native species were more diverse in the lower estuary, closer towards the marine extreme. However, notable range expansions have occurred for introduced species, with presence now recorded at all sites except White Rock, South Arm. This broader distribution has been accompanied by increased abundances—particularly in the upper estuary, where introduced cryptic fish and invertebrates now average over 300 individuals per site, compared to 100–150 in 2010. Importantly, this rise in introduced species abundance has not been matched by a corresponding decline in native species abundance, suggesting that coexistence, rather than replacement, may be occurring at some sites.

Algal species richness and spatial distribution patterns were broadly similar between surveys, with higher richness consistently observed in the southern estuary in both 2010 and 2024. However, percent cover of algae increased dramatically in 2024 for several upper estuary sites, spanning from Bedlam Walls North to Battery Point – locations that were largely dominated by tubeworm matting in 2010, representing an alternate stable state. Foliose algal cover also increased across a similar spatial extent, particularly from Bedlam Walls North to Tranmere Point. Algal species composition remained relatively stable across surveys, with dominant species in 2024 including *Ecklonia radiata*, *Carpoglossum confluens*, *Sargassum fallax*, *Acrocarpia paniculata*, and *Lessonia corrugata*. Notably, *E. radiata* expanded its distribution northwards into areas where it had not been previously recorded. This shift – alongside the complete disappearance of tubeworm matting between Bedlam Walls North and Battery Point – indicates a substantial ecological transition. Together, these changes suggest a positive trajectory for the estuary’s recovery, marked by increased algal abundance and re-establishment of kelp-dominated communities in areas previously impacted by ecological degradation.

The second survey method recorded the abundance of fish, invertebrates and algae in a 20-minute timed swim (and intertidal search) across multiple depths and habitats, using a ranking system (1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = >1000 individuals). Overall, species richness of fish, invertebrates, and algal species remained relatively consistent between 2010 and 2024, with a slight decline observed in 2024. Richness was generally highest in reef habitats and in the lower estuary, particularly within the intertidal, 0-1 m, and 1-4.9 m depth zones. As in the standard surveys, richness tended to increase from north to south. While site-level variability was present, changes in richness between years were minimal.

Introduced species richness and spatial distribution patterns were broadly similar across surveys, although overall richness increased in 2024, particularly on reef habitats. Species recorded with only a single individual at a site (least abundant) showed high variability in both surveys, suggesting fluctuations in rare or transient taxa. Most observations in sand habitats occurred at 5–8.9 m depth, whereas reef habitats supported a broader range of species from intertidal to 4.9 m.

In total, 60 fish species, 140 macro-invertebrates, and 45 algal species were recorded during the timed surveys. *Forsterygion varium*, *Notolabrus tetricus*, *Forsterygion gymnotum*, *Trachinops caudimaculatus*, *Neoodax balteatus* were among the most abundant fish species recorded in both 2010 and 2024. Notably, *F. varium* and *N. balteatus* expanded their distribution across the entire estuary in 2024, with several new site occurrences and increases in rank abundance by several-fold. Comparatively, *T. caudimaculatus* showed a contraction in range, with its absence recorded at multiple sites where it was previously observed in 2010.

As in the standard surveys, macro-invertebrate species richness generally increased from north to south, particularly in intertidal habitats. Despite some site-level variability, year-to-year changes in richness were minimal. *Asterias amurensis*, *Crassostrea gigas*, *Patiriella regularis*, *Ostrea angasi*, *Petrolisthes elongatus*, *Mytilus galloprovincialis* were among the most abundant macro-invertebrate species recorded in both 2010 and 2024. As *C. gigas* and *O. angasi* can overlap in spatial and depth distribution, and have overlapping shell morphologies in some locations, a follow-up survey was undertaken to (in)validate field

observations. That survey included collection of specimens from multiple depths at a range of locations where they were recorded, with identifications checked by a taxonomic expert, and further validated by tissue testing with a DNA probe previously developed for identification of *O. angasi*. The results suggested that the initial field observations were generally correct, however, in places some individuals of the two species were impossible to separate based on external shell morphology, possibly leading to some mis-identification in field surveys. Despite this, both species were present in the samples collected, with *C. gigas* dominating in the intertidal zone, and with increasing *O. angasi* numbers at depth, from a minor component at 0-1 m depth, to dominating the oyster numbers at depths below 5 m. Overall, *O. angasi* appears to be abundant within the estuary, although not forming the extensive oyster beds historically reported for this estuary. Over the decade between surveys, *A. amurensis* increased in rank abundance at most sites, occurring at several new sites across all habitat types. While *A. amurensis* was most frequently observed in sand habitats, rank abundance on reef and ecotone habitats also increased substantially – by several orders of magnitude in 2024. In contrast, *C. gigas* declined by more than ten-fold at nearly all sites and was absent from all ecotone habitats in 2024. *P. regularis* maintained a similar spatial distribution to 2010, with some fluctuations in rank abundance across sites. Notably, two sites (Bellerive Yacht Box and Rosny Point South) recorded high rank abundances in 2024, with 100-1000 individuals on sediment surveys, despite no observations of *P. regularis* in 2010.

Algal species richness increased from north to south – a spatial pattern observed in 2024 across multiple depths for reef habitats; intertidal, 0-1 m, and 1-4.9 m. Algal community composition across the surveys varied, with only three species remaining among the most abundant in 2024; *Cystophora moniliformis*, *Ecklonia radiata*, and *Ulva spp.* Both *C. moniliformis* and *E. radiata* shifted further north, supporting the ecological shift seen in the standard surveys. *Ecklonia radiata* can be used as an indicator for water clarity, and as such improved environment conditions. We found this species to have increased in both rank abundance and an expanded depth range in 2024, particularly within 0-1 m, 1-4.9 m and 5-8.9 m subtidal zones. Notably, Bellerive Bluff and Bellerive/Howrah Bluff increased from a rank of 1-2 (≤ 10 individuals) in 2010, to near-complete coverage (rank 4-5; 101-1000+ individuals) in 2024. This expansion in kelp coverage at deeper depths, potentially due to improved water quality and greater light penetration, coupled with the disappearance of widespread tube-matting previously recorded between Bedlam Walls and Battery Point, suggests an overall positive ecological shift and signs of recovery in the estuary.

For introduced algal species, *Aodes nitidissima* remained in a tight distribution in proximity to the Tasman Bridge, while both *Undaria pinatiffida* and *Grateloupia turuturu* extending their range further up the estuary. The introduced red algae *Grateloupia turuturu* (first recorded in Tasmania in 2006 and suspected to have originated from Japan) was found to have increased to very high densities in the immediate subtidal zone at many locations, particularly seaward of Bellerive Bluff on the eastern shore.

Overall, the findings of this repeated survey indicate areas of concern—especially the rising numbers of invasive species—but also evidence of ecological recovery within the Derwent Estuary. The expansion of kelp cover, reduced coverage of tube-worm matting, and increased cryptic fish abundance in previously degraded areas point to improving estuarine conditions. Continued monitoring will be essential to track these trends and guide future management actions.

1. Introduction

The Derwent Estuary covers an area of 198 km², extending from the northern-most point of New Norfolk to Iron Pot Light at its mouth. Home to a diverse range of species, including the threatened Spotted Handfish (*Brachionichthys hirsutus*), and the Derwent River seastar (*Marginaster littoralis*), a species now presumed to be extinct (O'hara et al. 2018) but requiring confirmation. This estuary is predominantly made up of sand and silt, alongside other habitats including wetlands, seagrass meadows and near-shore rocky reefs (DEP 2020). As a temperate estuarine system adjacent to a major urban centre, the Derwent serves as the focal point for the Hobart community with numerous recreational and commercial uses. The estuary is an important and productive ecosystem, providing key habitat for numerous species. However, due to its proximity to Hobart and catchment activities including farming, forestry, a paper mill and aquaculture, the estuary continues to face cumulative pressures, including urban run-off, siltation, sewage treatment plant nutrient discharges, legacy heavy metal contamination, climate change and invasive species, all of which add to the degraded conditions of the Derwent (DEP 2020; DEP 2021). Like many temperate estuarine systems globally, the Derwent is also likely responding to broader environmental changes such as warming waters and varying flood levels from the riverine extremes, which may influence salinity gradients, marine species distributions and estuarine connectivity. Recent work has highlighted gaps in understanding the ecological consequences of estuarine outflows, including contaminant and nutrient export, particularly in systems along Australia's southeast coast (Rasmussen et al. 2024).

Historically, the Derwent Estuary has recorded high concentrations of metals, as a result of past industrial activities and proximity to Hobart. Currently, the largest source of metals in the estuary is from contaminated groundwater (Macleod & Coughanowr 2019). The Nyrstar zinc smelter located at Lutana began discharging effluent containing metals when it was established in 1917, and the paper mill in Boyer also discharged metals, particularly mercury, into the estuary in the past (DEP 2020). Once in the water environment, metals and other contaminants pose persistent toxic health risks, often evidenced by the accumulation of metals in fish and shellfish tissues (de Souza Machado et al. 2016; DEP 2021). Management and remediation work is ongoing within the Derwent Estuary, including the implementation of stormwater catchment systems and groundwater interception walls, each allowing for metal removal processes (Nyrstar Hobart 2017). As a result, recent studies show declines in total zinc in ambient waters across surveys from 2007 to 2019 (DEP 2021). However, metals still remain in the tissue of all wild seafood species collected throughout the Derwent in 2019, and the Derwent remains a heavily degraded estuary. Assessing biodiversity changes through time will provide a clearer picture of the nature of the estuary, and any positive recoveries that may be occurring since 2010.

The 2010 biological baseline survey of Derwent Estuary reefs and adjacent sediment habitat documented clear spatial gradients in biodiversity, with fish and algal species richness increasing southward toward the lower estuary and more marine and wave-influenced-influenced sites (Barrett et al. 2012). Macroinvertebrate communities were also strongly structured by habitat type. Several sites exhibited signs of ecological degradation, including high prevalence of introduced species and a lack of native algal cover. Introduced species are a significant issue of concern within the Derwent, with previous work identifying *Asterias amurensis* and *Patiriella regularis* as particularly invasive (Ross et al. 2004; Ross et al.

2006). The baseline surveys firmly established several introduced species as inhabiting the mid-estuary, from Bedlam Walls to Battery Point and Bellerive Bluff, including both introduced sea stars, as well as the introduced crabs *Petrolistes elongatus* and *Carcinus maenus*. Assessing how these introduced populations have changed in the last decade is critical to conservation efforts and sustainable management in the Derwent.

Furthermore, ecosystem changes within habitats are critical to monitor. The Derwent follows a biological gradient typical of other estuaries, with a seagrass/filamentous algal mix on shallow, silty reefs in the upper estuary, to a diverse and complex algal assemblage in the south (Potter et al. 2010; Alves et al. 2020). The baseline survey identified several ecological features indicating a degraded ecosystem, including the 100% cover of tube-worm matting observed in the mid-estuary at Rosny Point, Battery Point and further upstream. Historically, algal assemblages have failed to re-establish at these sites, likely due to the restrictive and all-encompassing nature of the polychaete matting (Barrett et al. 2012). Monitoring sites where this alternative stable state is present could further our understanding of the recovery progress within the estuary.

1.1 Objectives and approach

This study replicates the biological surveys completed in 2010, with the aims of identifying and quantifying biodiversity changes through time, relative to the initial study, in addition to providing a framework for a decadal monitoring program going forward. This ongoing monitoring will inform our knowledge on the native species within the Derwent Estuary and how they are responding to changing water quality and environmental conditions, as well as describing changes in distribution and abundance of introduced pest species found in this system.

The initial study, conducted over a decade ago, aimed to quantify the distribution and abundance of reef-associated biodiversity within the estuary, thoroughly survey for rare species including *M. littoralis*, and quantitatively describe the patterns of introduced species. Here, we present the results of the second survey undertaken in 2023 and 2024; and draw comparisons to the first survey in 2010. This study aimed to:

1. Provide further data on the spatial distribution and abundance of species and benthic communities,
2. Quantitatively describe changes in historically abundant and newly dominant species, and,
3. Assess population changes through time for key introduced species.

The findings of this repeated survey are essential for monitoring changes in reef community dynamics over time in a changing environment, evaluating the effectiveness of current management actions, informing targeted restoration efforts, and guiding future monitoring programs in one of Tasmania's most ecologically and economically important estuaries.

2. Methods

This study replicates the methodology of surveys undertaken in the Derwent Estuary in 2009 and 2010, as part of an ongoing monitoring program to survey reef-associated biota and identify changing patterns in spatial and temporal distribution.

Multiple methods were utilised to describe the distribution of reef-associated biodiversity across 27 sites in the estuary. For survey purposes, sites were grouped into three regions along the estuarine gradient: the upper estuary (Cadbury Point, site 1, to Cornelian Bay, site 6), the mid estuary (200 m north of Tasman Bridge East Shore, site 7, to Battery Point, site 15), and the lower estuary (Tranmere Reef, site 16, to Pearsons Point, site 27).

Transect-based methods were conducted on rocky reefs in April 2023, and March – May 2024, at various depths, typically at 5 m where possible, but shallower at locations where reefs did not reach this depth. Additionally, timed surveys were undertaken during the same period to record the diversity and rank abundance of both common and rare species. This method allowed a greater area to be surveyed for biodiversity values than the transect method, trading off quantitative strength for increased species inventory and an understanding of depth-based trends at many sites. While these were primarily rocky reef and shoreline focussed, they also included ecotone habitat (where reef grades to sediment) and sediment habitat at many locations. Due to site conditions and time constraints, surveys could not be completed at all sites to the same extent of replication. Details of sites, depths and habitats sampled are in Table 1.

2.1 Standard subtidal reef surveys

The standard survey method involves the underwater visual census (UVC) of fish, mobile macro-invertebrates and algae densities, conducted according to the Australian Temperate Reef Collaborating (ATRC) methodology.

Surveys were conducted on SCUBA by a 2-4 personnel dive team at 19 sites. At each site, two 100 m transects were laid out in opposite directions, starting at the boat anchor, and divided into two 50 m segments (totalling 4 x 50 m transect blocks for each depth surveyed). Divers recorded fish observations along 5 m bands on either side of the transect (Figure 1), with a total of 4 x 10 m x 50 m fish transects surveyed at each site. Fish abundance and size of individual fish was recorded.

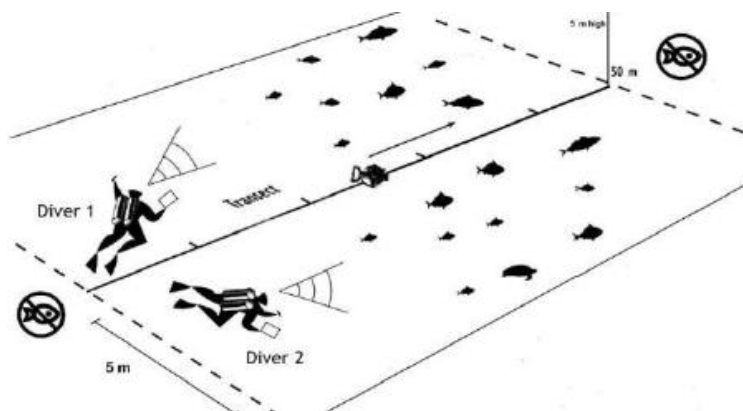


Figure 1: Stylised representation of method 1 survey technique

Cryptic fish and mobile macro-invertebrates were surveyed along the same transect line, within a 1 m band on one side of the transect (Figure 2). Cryptic fish were identified, counted and sized where possible, and mobile macro-invertebrates were identified and counted. A total of 4 x 1 m x 50 m transects were typically surveyed at each site. At some sites, there was insufficient reef for 4 x 50 m transects, so these were restricted to 2 x 50 m (see Table 1).

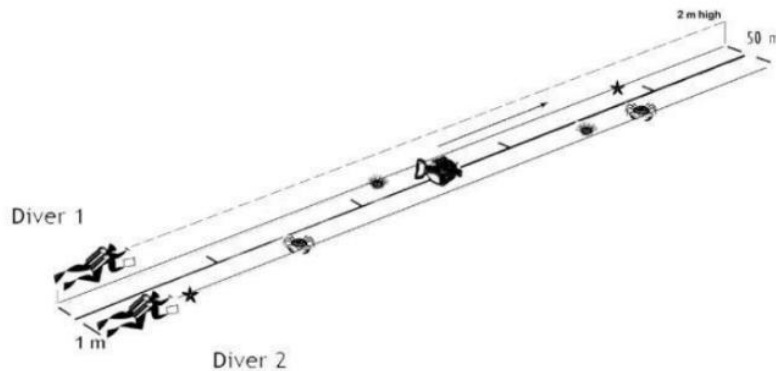


Figure 2: Stylised representation of method 1 survey technique.

The cover of macroalgae, sessile invertebrate species and substrata was quantified by placing a grided 0.25 m² quadrat at 10 m intervals along the transect lines. Covers were assessed by identifying and counting species that occurred directly under 50 intersection grid points. Percentage cover could then be quantified.

2.2 Timed intertidal and subtidal surveys

Timed intertidal and subtidal surveys were undertaken to record the diversity and rank abundances of common and rare species across various depth ranges at 24 sites.

Surveys were conducted on SCUBA by a 2-4 personnel dive team. A survey involved a 20-minute search, with divers recording fish, macro-invertebrates, large brown and green algae species, and introduced or 'key' species of red algae. Surveys were typically conducted along reef systems, but where possible, soft sediment and ecotone searches (comprising of patch reef and open sediments) were completed to capture the habitat of known rare species such as the spotted handfish. Surveys were undertaken between April and June 2024 at different depth ranges (intertidal, 0 – 0.9 m, 1 – 4.9 m, 5 – 8.9 m, 9 – 15 m) in reef, sand, and ecotone habitats depending on what habitat/depth range was available at each site. Species presence was recorded as a ranked abundance on a log₁₀ scale in the following categories: 1 (only one specimen observed), 2 (2 – 10 individuals), 3 (11–100 individuals), 4 (101 – 1000 individuals) and 5 (> 1000 individuals). The first 10 minutes focussed on more common species, followed by targeted searches for rarer species in the last 10 minutes. This involved searching crevices and turning over a minimum of ten boulders (where possible).

Intertidal searches were undertaken at 19 sites and were restricted to rocky shoreline habitat. Any sites not surveyed (Table 1) was due to a lack of appropriate intertidal habitat. Where possible, a minimum of ten boulders/rocks were turned at each site to identify and count the abundance of cryptic species.

2.3 Data analysis

Data from the standard subtidal reef surveys were obtained from (for 2010 survey data), and entered into (for 2024 survey data) the National Reef Monitoring Network database hosted by the AODN (IMOS) (<https://imos.org.au/facility/autonomous-underwater-vehicles/national-reef-monitoring-network>) for both survey periods and analysed using RStudio. Data from the 2024 timed intertidal and subtidal surveys were collated with the 2010 dataset and similarly analysed using RStudio. All data were cleaned, summarised, and visualised using a combination of ‘tidyverse’ packages, with site-level comparisons conducted to assess changes in species richness, abundance, and spatial distribution across depths and habitats. The most abundant species identified for each survey method and survey period were analysed across sites to evaluate spatial and temporal shifts in their abundance and distribution between 2010 and 2024.

2.4 Targeted Oyster Identification

As it was found that in some locations *Ostrea angasi* and *Crassostrea gigas* were not always able to be readily differentiated based on external morphological features during diver-based ecological surveys, a follow-up study was undertaken at a selected range of sites and depth ranges where oyster species had been encountered during the main surveys. At each site and depth combination, approximately 10 individuals were collected where possible and returned for lab-based identification by a taxonomic expert based on characteristic internal structures of each species. To confirm morphological IDs, DNA was extracted from adductor muscle of a sub-sample of oysters using the QIAGEN DNeasy Blood & Tissue kit. Extracted DNA was diluted to 1 ng/uL and an *O. angasi* specific taqman qPCR test was performed in triplicate for each sample following the methods presented in Coutts et al. (2022). A positive result required amplification of all three technical replicates while a negative result required no amplification in all three technical replicates.

Table 1: Details for standard and timed surveys conducted in the Derwent in 2010 and 2024. Positions are recorded in decimal degrees using WGS 84.

<u>SITE DETAILS</u>				<u>STANDARD SURVEYS</u>			<u>TIMED SURVEYS</u>			
Site No.	Site Name	Lat	Long	Depth (m)	No. of transects (2010)	No. of transects (2024)	Depth	Habitat	No. of searches (2010)	No. of searches (2024)
1	Cadbury Point	-42.7885	147.281	1	2	2	Intertidal 1.1 – 4.9	Reef	2 2	2 2
2	Dowsings Point	-42.8184	147.2977				Intertidal 0 – 1	Reef Ecotone Reef	2 1 1	2 1 1
3	Bedlam Walls North	-42.832	147.3273	4	2	2	Intertidal 1.1 – 4.9	Reef Reef	2 2	2 2
4	Bedlam Walls South	-42.8379	147.3312	4	2	2	Intertidal 1.1 – 4.9	Reef Reef	2 2	2 2
5	Point South of Gielston Bay	-42.8443	147.3376	3	2	2	Intertidal 0 – 1 1.1 – 4.9 5 – 8.9	Reef Reef Ecotone Reef Sand	2 3 2 3 1	2 3 2 3 1
6	Cornelian Bay	-42.8518	147.3279	1	2	2	Intertidal 0 – 1 1.1 – 4.9	Reef Reef Reef Sand	2 1 2 1	2 1 2 1
7	200 m North Tasman Bridge East Shore	-42.861	147.3501	2	2	2	Intertidal 0 – 1 1.1 – 4.9 5 – 8.9	Reef Reef Ecotone Reef Sand	2 3 2 2 2	2 3 2 2 1
8	Collegiate Rowing Club	-42.8636	147.3362				Intertidal 0 – 1	Reef Reef	2 2	2 2
9	Rosny Point North	-42.8719	147.3504	2	2	2	Intertidal 0 – 1 1.1 – 4.9 5 – 8.9	Reef Reef Reef Sand	2 3 2 3	2 3 2 2
10	Naval Cadets	-42.8724	147.3355				Intertidal 0 – 1	Reef Reef	2 2	2 2
11	Bellerive Yacht Box	-42.8757	147.3619				5 – 8.9	Sand	2	2
12	Rosny Point South	-42.8762	147.3543				5 – 8.9	Sand	2	2
13	Bellerive Bluff	-42.882	147.3652	4	2	2	Intertidal 0 – 1 1.1 – 4.9 5 – 8.9 9 – 15	Reef Reef Reef Ecotone Ecotone Sand Sand	2 3 1 3 1 2 1	2 3 1 3 1 2 1
14	Bellerive/Howrah Bluff	-42.8825	147.3854	4	2	2	Intertidal 0 – 1	Reef Reef	2 2	2 2

							1.1 – 4.9	Reef	2	2
							5 – 8.9	Sand	1	1
							9 – 15	Ecotone	2	2
15	Battery Point	-42.8897	147.3394	2	2	2	Intertidal	Reef	2	2
							0 – 1	Reef	2	3
							5 – 8.9	Sand	1	1
16	Tranmere Reef	-42.8999	147.4095				5 – 8.9	Reef	1	1
17	Tranmere Point	-42.9304	147.4093	3	4	4	Intertidal	Reef	3	2
							0 – 1	Reef	2	2
							1.1 – 4.9	Ecotone	1	1
								Reef	1	2
								Sand	1	1
							5 – 8.9	Reef	1	1
18	The Grange, Taroona	-42.9326	147.3616	4	4	4	Intertidal	Reef	2	2
							0 – 1	Reef	2	2
							1.1 – 4.9	Reef	2	2
							5 – 8.9	Ecotone	2	2
19	Crayfish Point	-42.9521	147.356	3	4	4	Intertidal	Reef	2	2
							0 – 1	Reef	3	3
							1.1 – 4.9	Reef	2	2
							5 – 8.9	Reef	2	2
								Sand	1	1
20	Alum Cliffs	-42.9666	147.3419	5	4	4	Intertidal	Reef	2	2
							0 – 1	Reef	3	3
							1.1 – 4.9	Reef	2	2
							5 – 8.9	Ecotone	1	1
								Sand	1	1
21	Mary-Ann Bay	-42.9726	147.3996				5 – 8.9	Sand	2	2
22	White Rock, South Arm	-42.9774	147.3923	4	4	4	Intertidal	Reef	3	2
							0 – 1	Reef	2	1
							1.1 – 4.9	Reef	2	3
							5 – 8.9	Ecotone	1	1
								Sand	1	1
23	Opossum Bay	-42.9901	147.4018				5 – 8.9	Sand	2	2
24	Blackmans Bay Blowhole	-42.9984	147.3297	4	4	4	Intertidal	Reef	2	2
							0 – 1	Reef	3	2
							1.1 – 4.9	Reef	2	2
							5 – 8.9	Ecotone	1	1
								Sand	1	1
25	Blackmans Bay South	-43.0118	147.3311	5	4	4				
26	Lucas Point	-43.0383	147.3389	5	4	4				
27	Pearsons Point	-43.0534	147.3421	5	4	4				

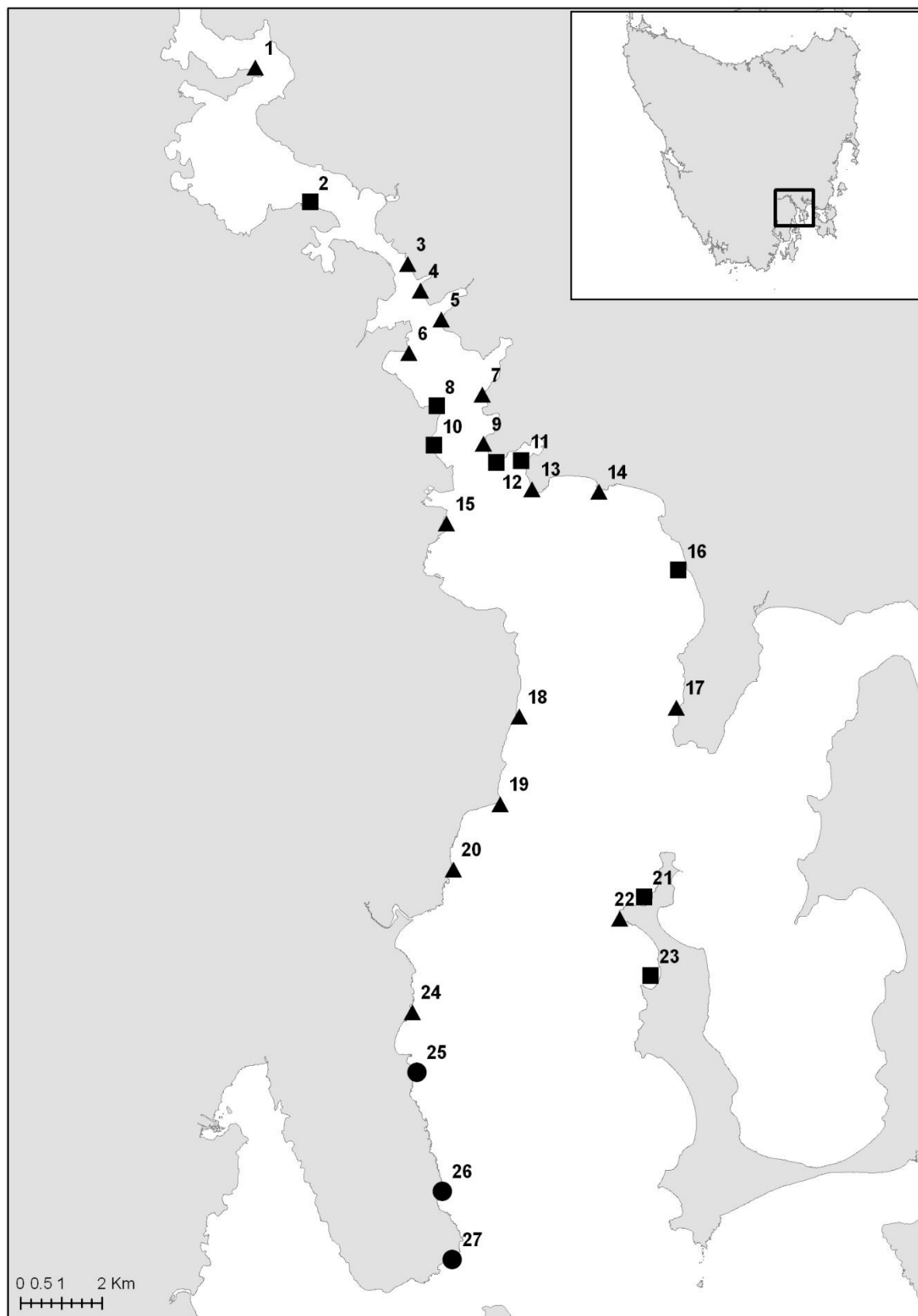


Figure 3: General map showing locations of sites surveyed in the Derwent in February-April in 2010, and March-June 2024. All sites were surveyed using the same methods in both survey periods. ● standard surveys; ■ timed surveys; ▲ standard and timed species surveys.

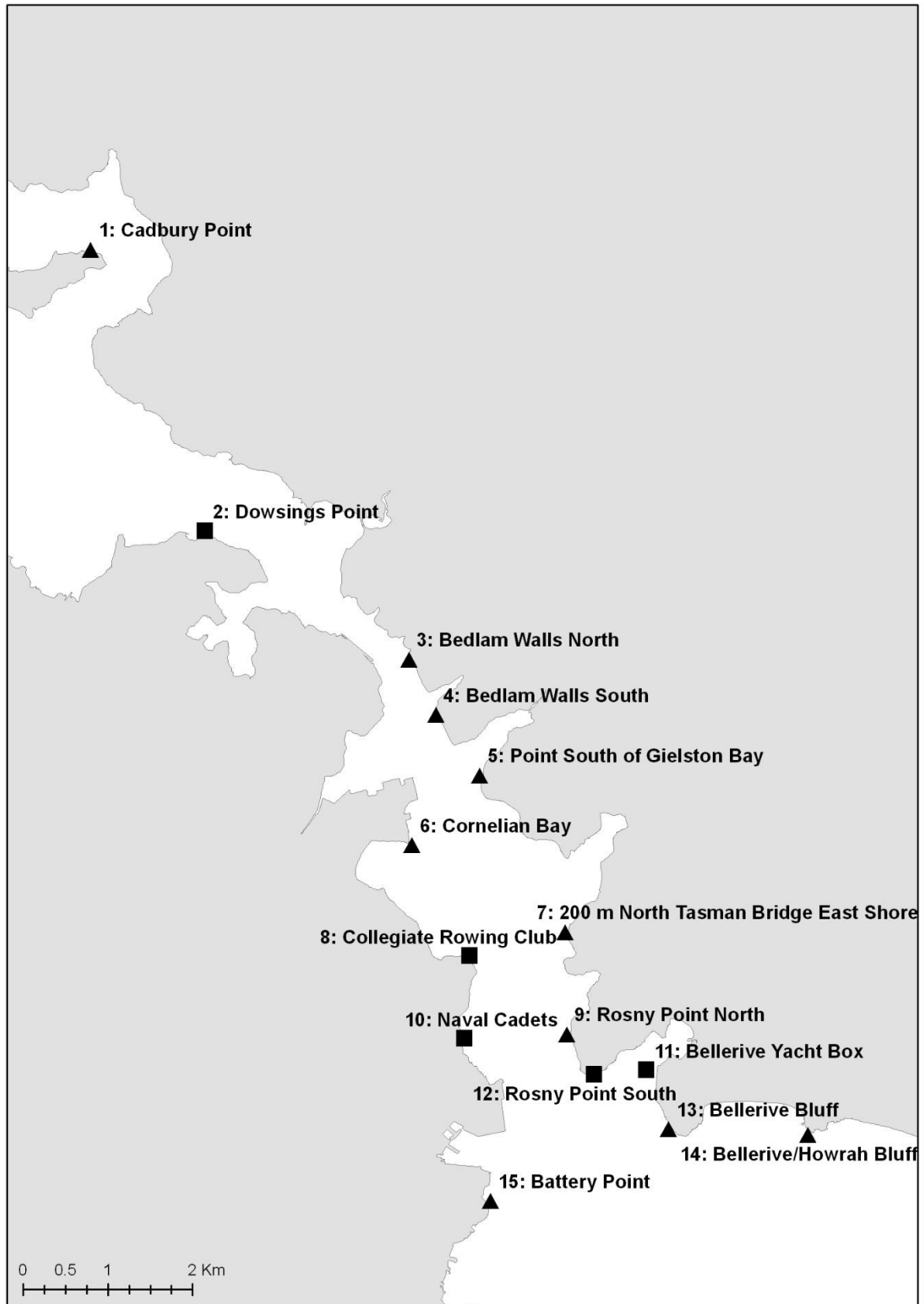


Figure 4: Detailed map showing site positions surveyed in the northern Derwent in February-April in 2010, and March-June 2024. All sites were surveyed using the same methods in both survey periods. ■ timed surveys; ▲ standard and timed species surveys.

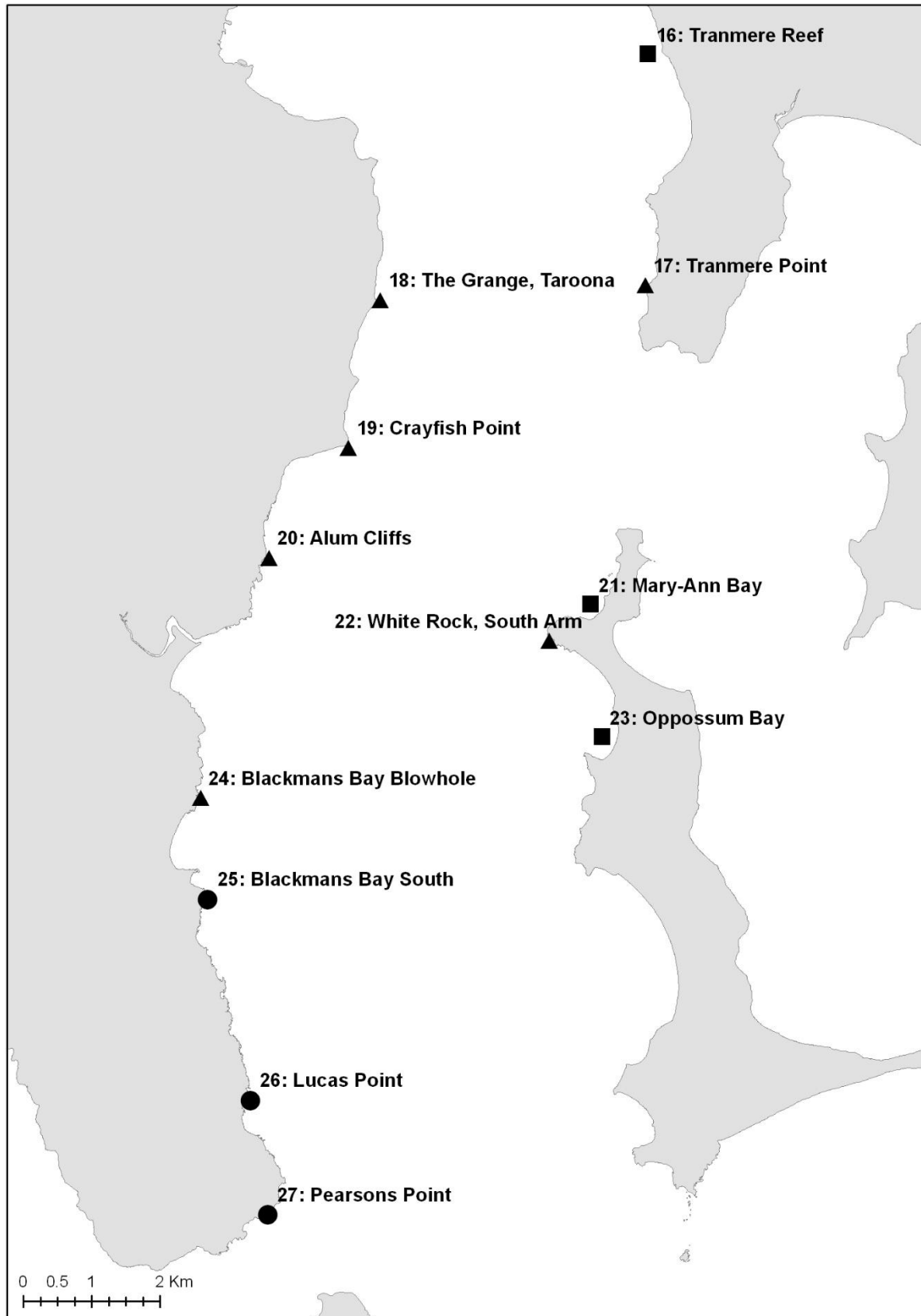


Figure 5: Detailed map showing site positions surveyed in the southern Derwent in February-April in 2010, and March-June 2024. All sites were surveyed using the same methods in both survey periods. ▲ standard and timed surveys; ● standard surveys; ■ timed surveys.

3. RESULTS

A substantial dataset was collected through the 2024 repeat sampling and combined with the baseline dataset from 2010 to allow temporal comparisons. Together, these datasets provide a comprehensive overview of spatial and temporal trends in biodiversity across the Derwent Estuary. The following results summarise observed changes in species richness, abundance, and distribution across fish, invertebrate, and algal communities, assessed across multiple depths and habitat types.

3.1 Standard subtidal surveys

Fish

Forty species of fish were recorded during the 2024 standard survey, a slight decrease from the 54 counted in the 2010 surveys. Species richness was generally consistent across the region, ranging from 4-6 species for most sites (Figure 6). Sites 1 and 3 (Cadbury Point and Bedlam Walls North) in the upper estuary had the lowest recorded species richness, ranging from 1-3 species, whereas sites 24, 25, 27, and 27 (Blackmans Bay Blowhole, Blackmans Bay South, Lucas Point, and Pearson Point) located in the lower estuary had the greatest richness (6-12 species). This pattern of gradually increasing richness from upper to lower is similar to the 2010 surveys, however there is an overall decline in the number of species recorded, with fewer species found in the current survey. A slight increase in richness was observed in the upper and mid- estuary when compared to the 2010 surveys (sites 4 to 9), however the opposite was evident in the lower sites, with a decline in species richness from site 17 (Tranmere Point) onwards.

The most abundant fish species were largely consistent across survey periods, though overall abundances generally decreased in the 2024 survey. Seven species – *Dinolestes lewini*, *Forsterygion varium* (an introduced species), *Latridopsis forsteri*, *Notolabrus tetricus*, *Pempheris multiradiata*, *Pictilabrus laticlavius*, and *Trachinops caudimaculatus* – remained among the top ten most abundant in both 2010 and 2024. *T. caudimaculatus* was the most abundant species across both surveys, although its numbers declined substantially in 2024, with only 100-500 individuals recorded per transect, compared to over 1000 individuals on multiple transects in 2010 (Figure 7g). A similar decline was observed in other species, including *D. lewini* (Figure 7a), and *L. forsteri* (Figure 7c). The decline in *L. forsteri* was over an order of magnitude, and across multiple sites where large resident schools had been previously sighted, including site 18 (The Grange). Minor range shifts are also apparent, with *P. multiradiata* and *P. laticlavius* (Figure 7e and 7f) occurring further north into the mid-estuary in 2024, and the introduced *F. varium* (Figure 7b) extending into the lower estuary. In 2024, *N. tetricus* abundance generally increased further down the estuary, rising steadily from site 13 onwards, although dips were observed at sites 19 and 27 (Figure 7d). In contrast, its distribution in 2010 was more variable, peaking at sites 19, 20 and 25. Notably, 45 individuals were recorded on a single transect at site 9 (Rosny Point North) in 2024, despite the species being entirely absent from that site in 2010.

Several species that were abundant in 2010 were rarely recorded in 2024, if at all.

Acanthaluteres vittiger was only observed at three of the eight sites where it was previously recorded in 2010 and was newly recorded at two upper estuary sites (sites 4 and 6) where it had not been detected in the earlier survey (Figure 8a). In contrast, *Favonigobius lateralis* and *Pseudocaranx georgianus* (Figure 8b and 8c) were not recorded at any sites in 2024.

Conversely, three species that were rarely observed in 2010 became among the top abundant species in 2024. *Acanthaluteres spilomelanurus* showed the highest numbers in the mid-estuary at sites 13 and 14, despite being absent from those sites in 2010 (Figure 8d).

Atypichthys strigatus was recorded for the first time in the estuary, with 61 individuals recorded in 2024 – mostly in the lower estuary (Figure 8e). This was the first time this species was present, having not previously been observed in 2010. *Neoodax balteatus*, while present in 2010 across much of the mid and lower estuary, showed a substantial increase in abundance in 2024 (Figure 8f). This included new observations from upper and mid-estuary sites (sites 1, 6, and 9) where it had not previously been observed.

While not among the most abundant species, the introduced species *Forsterygion gymnotum*, has potential ecological implications and warrants ongoing monitoring (Figure 9). Overall abundance varied little across survey periods, with 40 individuals recorded in 2010 and 39 in the 2024 survey. However, its site distribution shifted noticeably. In 2010, site 3 recorded the highest abundance, averaging 17 individuals, while fewer than 2 individuals were observed at other sites. Comparatively, the 2024 survey showed a more even distribution across five sites, each averaging between 3-6 individuals.

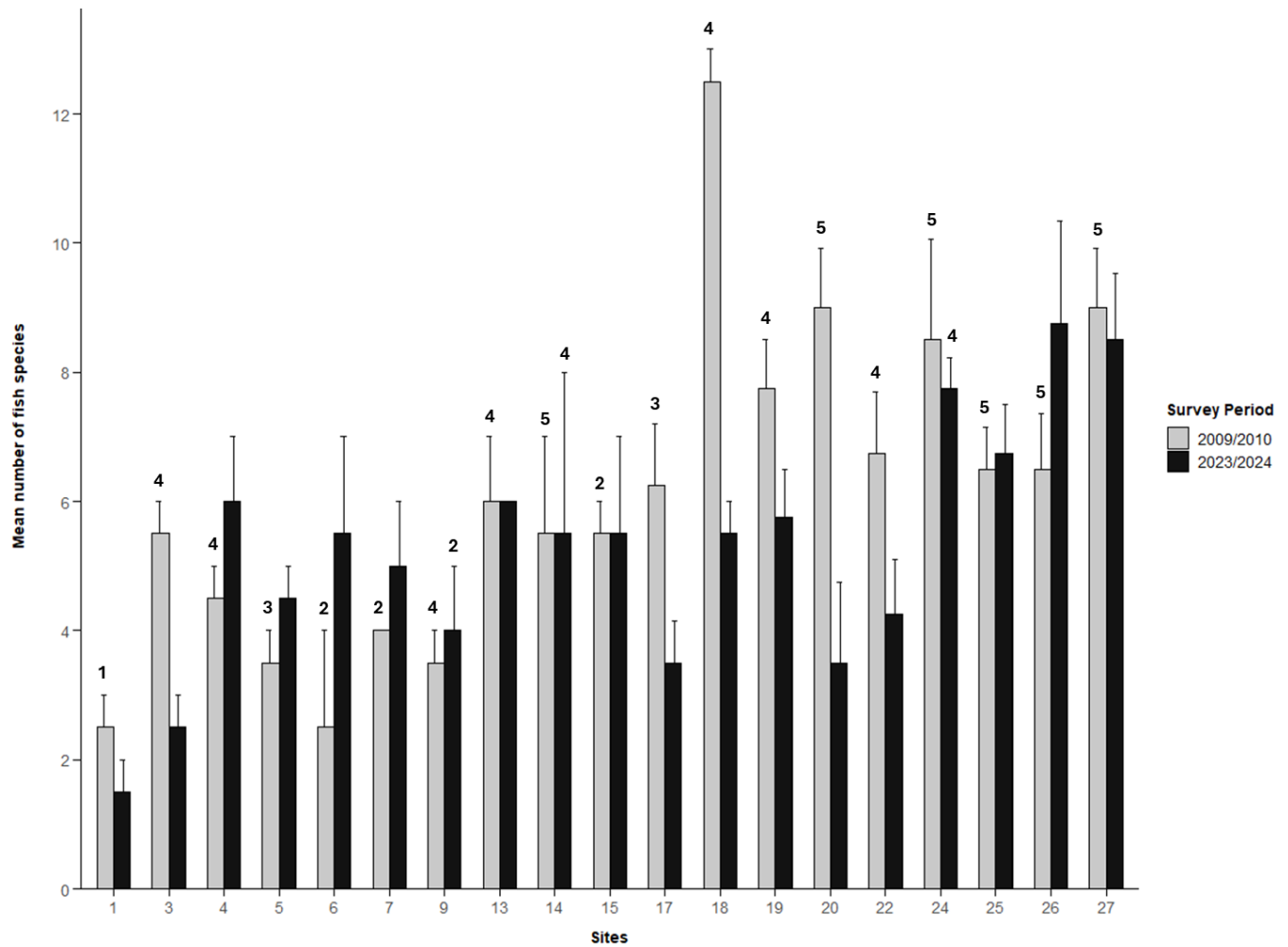


Figure 6: Fish species richness (mean $\pm$ SE) per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black). Transect depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

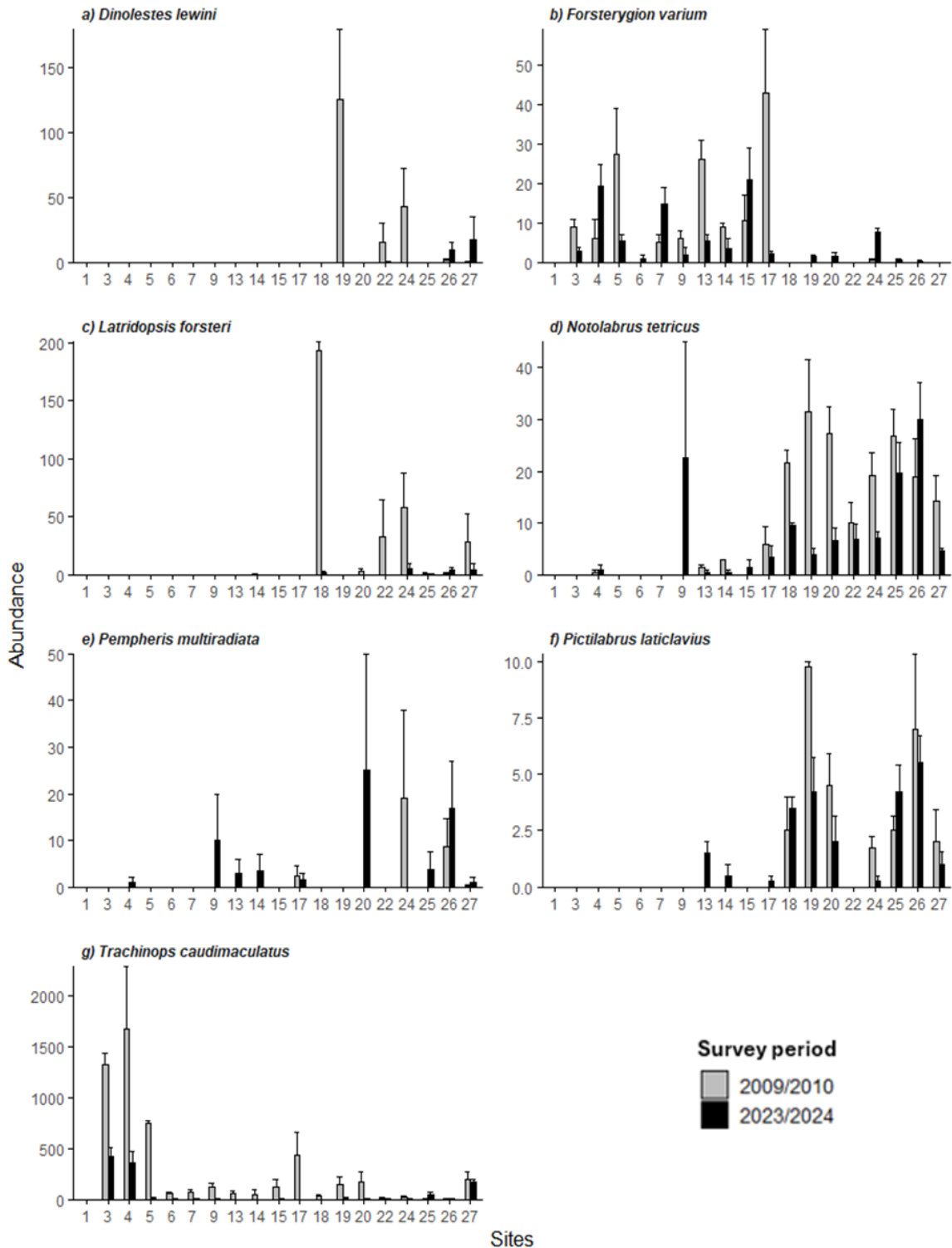


Figure 7: Abundance (mean $\pm$ SE) of the most abundant species per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black). The seven species shown were among the top 10 most abundant in both survey periods, and are presented alphabetically.

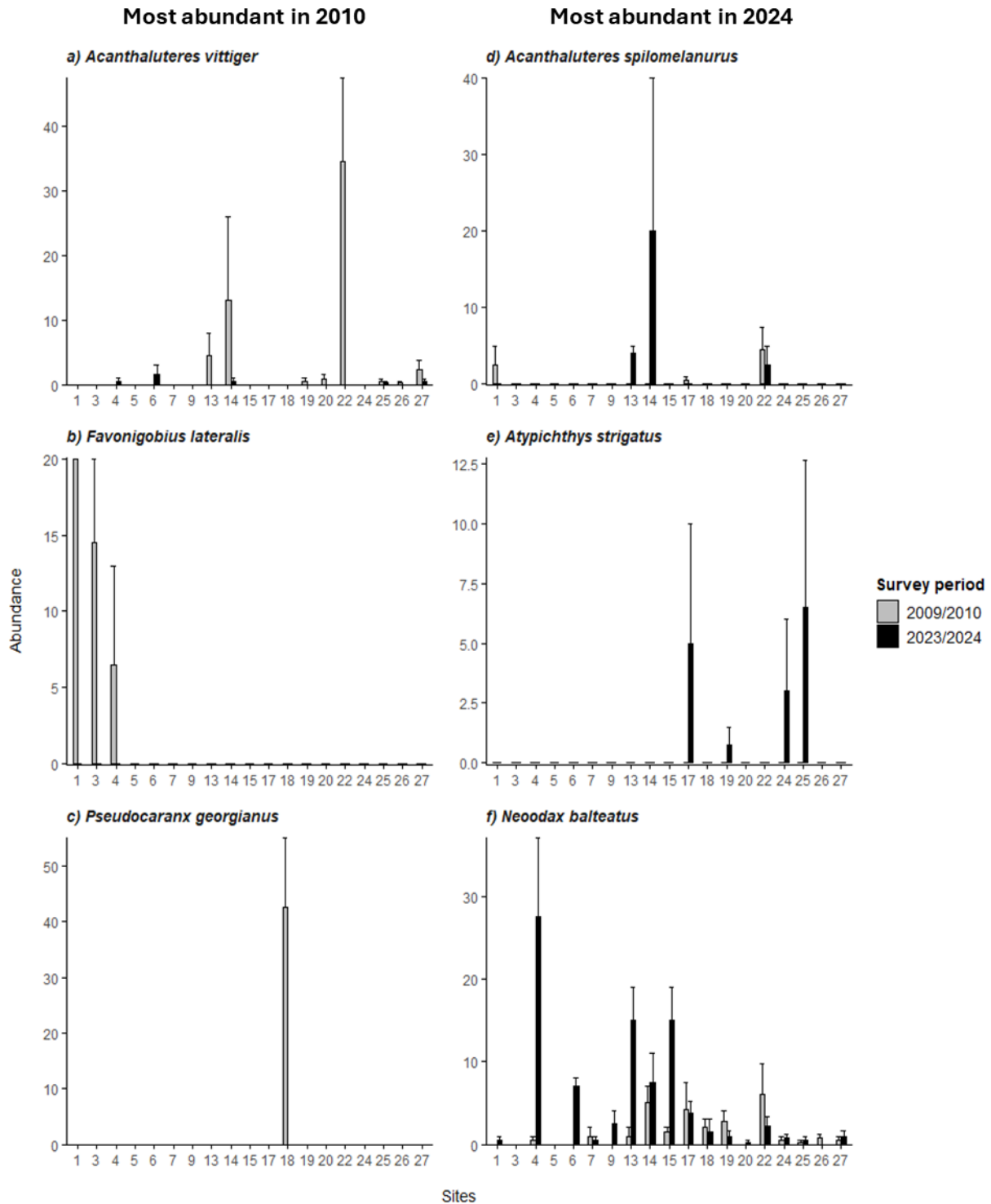


Figure 8: Abundance (mean $\pm$ SE) of fish species uniquely among the top 10 most abundant in 2010 (a, b, c) and 2024 (d, e, f), per transect (500 m²) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods (indicated by grey and black fill) are shown to highlight temporal shifts in occurrence and abundance. Species are presented alphabetically.

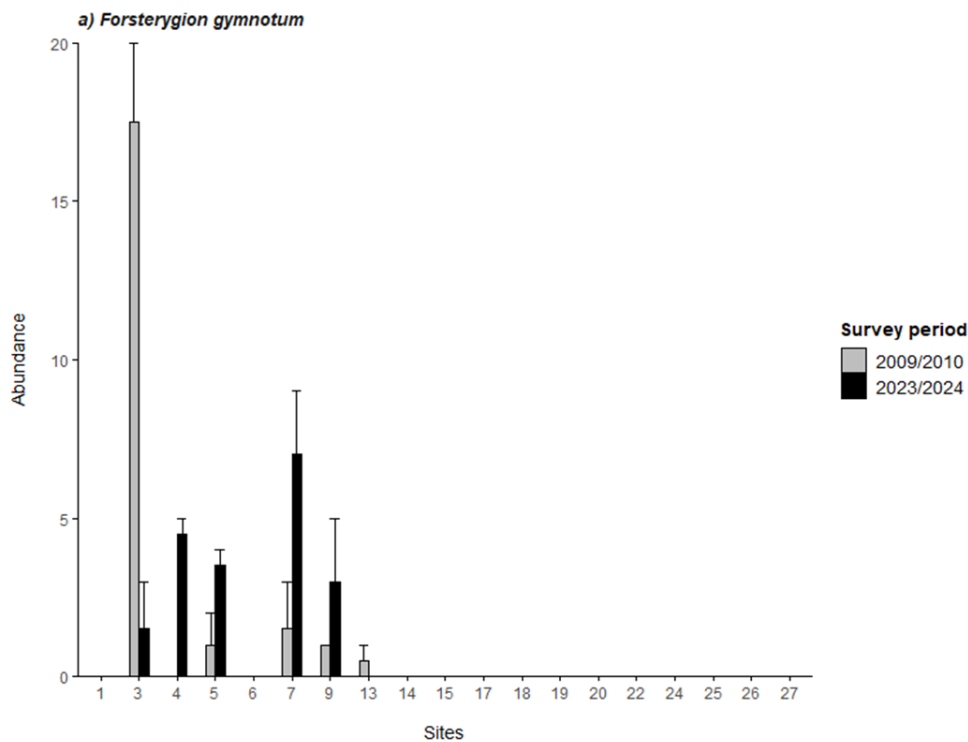


Figure 9: Abundance (mean $\pm$ SE) of the introduced *Forsterygion gymnotum* per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black). Previously recorded as *Grahamina gymnota*.

Cryptic fish and mobile macro-invertebrates

Twenty-one cryptic fish and 37 mobile macro-invertebrates were recorded in the Derwent Estuary in 2024, an increase from the 15 and 35 species, respectively, counted in 2010. Both cryptic fish species richness and overall abundance showed considerable increases in 2024 (Figure 10). Most sites in 2024 recorded 3-6 species per transect, compared to only five sites in 2010 with a mean exceeding 2 species. The increase in species richness was more pronounced in the upper estuary, with fewer cryptic fish species recorded from site 17 southward. This contrasts with the 2010 distribution, where greater species richness was observed in the mid-estuary. Site 1 in the north, along with sites 19 and 22 in the south, had the lowest species richness, with only one species recorded at both site 1 and 19, and no cryptic fish at site 22. In 2024, cryptic fish abundance was more sporadic, with no clear spatial trends. This is in contrast to 2010 when higher numbers were recorded in the mid-estuary. Several sites recorded over 40 individuals on average, compared to much lower means in 2010.

We assessed the top four abundant cryptic fish species across each survey period, identifying two species that remained consistent across surveys. *Forsterygion varium* increased substantially in 2024, with 797 individuals, compared to 221 individuals in 2010 (Figure 11a). This introduced species expanded its range into both the upper and lower estuary, where it was previously more common in the mid estuary. *Pempheris multiradiata* also saw an increase in abundance in 2024, with higher numbers recorded at several mid-estuary sites

(Figure 11b). Site 18 in the lower estuary (The Grange, Tarroona) had the highest abundance, averaging almost 30 individuals.

Gobiid spp. and *Scorpaena papillosa* (Figure 12a and 12b) were both abundant in 2010 but declined in numbers in 2024. A minor range shift was observed for *Gobiid* spp., with decreased numbers in the mid-estuary, and greater presence at upper estuary sites. *S. papillosa*, previously recorded at 11 sites in 2010, was only present at five sites in 2024, with increased numbers observed only at site 27. By comparison, *Forsterygion gymnotum* and *Trinorfolkia clarkei* (Figure 12c and 12d) were most abundant in 2024. A significant range expansion was recorded for the introduced *F. gymnotum*, which was previously found only at sites 9 and 13, but was recorded at nine sites in 2024. Similarly, *T. clarkei* was recorded at several sites in 2024 where it had not been detected in 2010, including sites in the upper, mid and lower estuary.

Species richness of cryptic macro-invertebrates was relatively consistent throughout the estuary, with means of 3-6 species recorded for most sites (Figure 13). In 2010, species richness gradually increased along a southern distribution, with only a single species recorded at site 1, compared to an average of 6-9 species in the lower estuary sites. Site 1 remained the site with lowest species richness in 2024, however no clear spatial trend was observed across the estuary, with site 24 in the lower estuary and site 14 in the mid-estuary both recording the greatest species richness (8.5 and 7.5, respectively). A slight decline in species richness across surveys is evident in the lower estuary, where all sites averaged more than five species in 2010 (with up to 9 species at site 25). In 2024, species richness ranged from 2-4 species at sites 22, 26, 27, while sites 20, 24, and 25 recorded 6-8 species.

The introduced echinoderm *Patiriella regularis* remained the most abundant macro-invertebrate across both surveys, although considerably higher counts were recorded in 2024 (Figure 14g). All sites where the species was present recorded mean abundances exceeding 200 individuals per site, with peaks close to and exceeding 400 at sites 3, 7 and 9 – more than doubling the numbers recorded at most sites in 2010. However, two sites had considerably lower mean abundances of *P. regularis*, with only 0.25 and 7.5 individuals per site (site 24 and 17, respectively), indicating some localized variation in distribution. Its range remained largely unchanged, with most occurrences still concentrated in the upper estuary. Similarly, *Asterias amurensis*, another introduced echinoderm, exhibited substantial increases in abundance with minimal range shifts (Figure 14a). Average abundances were recorded as between 50 and 200 individuals per site in 2024, a stark increase from the 2010 averages of 5-10 individuals per site.

Range shifts were observed for several other macro-invertebrates, all of which were among the most abundant for both surveys. The native *Heliocidaris erythrogramma* increased in abundance in the mid-estuary, and at site 17 (upper limit of the lower estuary) while becoming less common in other lower estuary sites where it was previously recorded in 2010 (Figure 14c). *Jasus edwardsii* which was primarily concentrated in the lower estuary in 2010, showed an overall increase in abundance but was also found further north in the mid-estuary (Figure 14d). Interestingly, *Lunella undulata* was recorded only at two lower estuary sites in 2024 (sites 18 and 19), both of which had low to no presence in 2010, when the species was more abundant at further south (Figure 14e). Other native species like *Cenolia trichoptera* and *Meridiastra calcar* (Figure 14b and 14f) had minimal changes across time, with both

species still concentrated at the same sites in the south in 2024, and relatively similar abundances.

Amblypneustes ovum, *Cabestana spengleri*, and *Maoriclplus roseus* (Figure 15a, 15b, and 15c) were all among the most abundant species in 2010, but were rarely recorded in 2024. In contrast, *Coscinasterias muricata*, *Haliotis rubra*, and *Neothyonidium* spp. (Figure 15d, 15e, and 15f) experienced significant increases in abundance, resulting in their inclusion among the top ten abundant species in 2024. Minimal range shifts occurred for *C. muricata* and *H. rubra*, while recordings of *Neothyonidium* spp. at site 25 appear to represent a new sighting.

While not among the most abundant species, three other introduced species are worth monitoring across surveys, due to their potential ecological implications (Figure 16). Unlike *P. regularis* and *A. amurensis*, the introduced *Carcinus maenas*, *Metacarcinus novaezelandiae* and *Petrolisthes elongatus* did not show significant increases in abundance. All three species exhibited relatively little change from 2010 to 2024, although distribution varied slightly for *C. maenas* and *M. novaezelandiae*, both of which only occurred at site 6 in 2024, where they had not previously been detected.

In general, the patterns of native versus introduced cryptic fish and macro-invertebrate species in 2024 mirrored those observed in 2010. Introduced species richness was greater and more prevalent in the upper estuary, while native species had greater richness in the lower estuary (Figure 17). However, in 2010, introduced species were concentrated between sites 3 and 18. In 2024, this range expanded to include all sites except site 22. With this broader distribution, there has also been an increase in their numbers, particularly in the upper and mid-estuary, where introduced cryptic fish and invertebrates now average 300 or more individuals per site, compared to 100-150 individuals per site in 2010. Sites 7 and 9, for instance, recorded means above 600 introduced individuals, more than doubling the previous averages of 150 – 260 individuals per site. Although richness increased for introduced species in the lower estuary sites, abundance still remained low, suggesting that the range extension for some of these introduced species is still in its initial phase. In contrast, we can see minimal changes in the abundance of native species in the upper estuary. Increases in native abundance can be seen at multiple sites (particularly sites 6, and 14–19). Notably, the rise in introduced species abundance does not correspond with any clear declines in native species abundance across sites in the estuary.

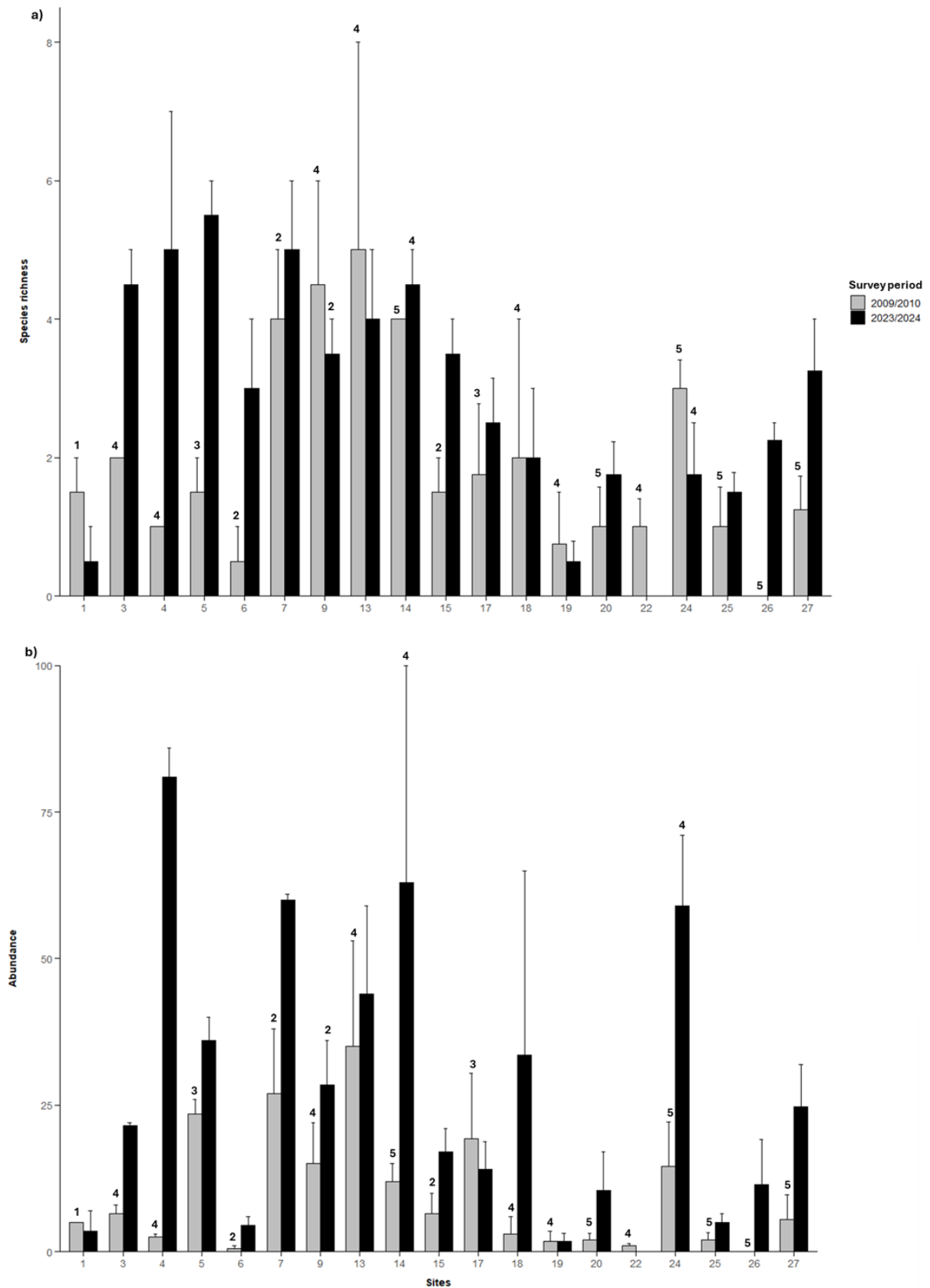


Figure 10: a) Species richness (mean $\pm$ SE) and b) abundance (mean $\pm$ SE) of cryptic fish in 2010 (grey) and 2024 (black). Transect depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths.

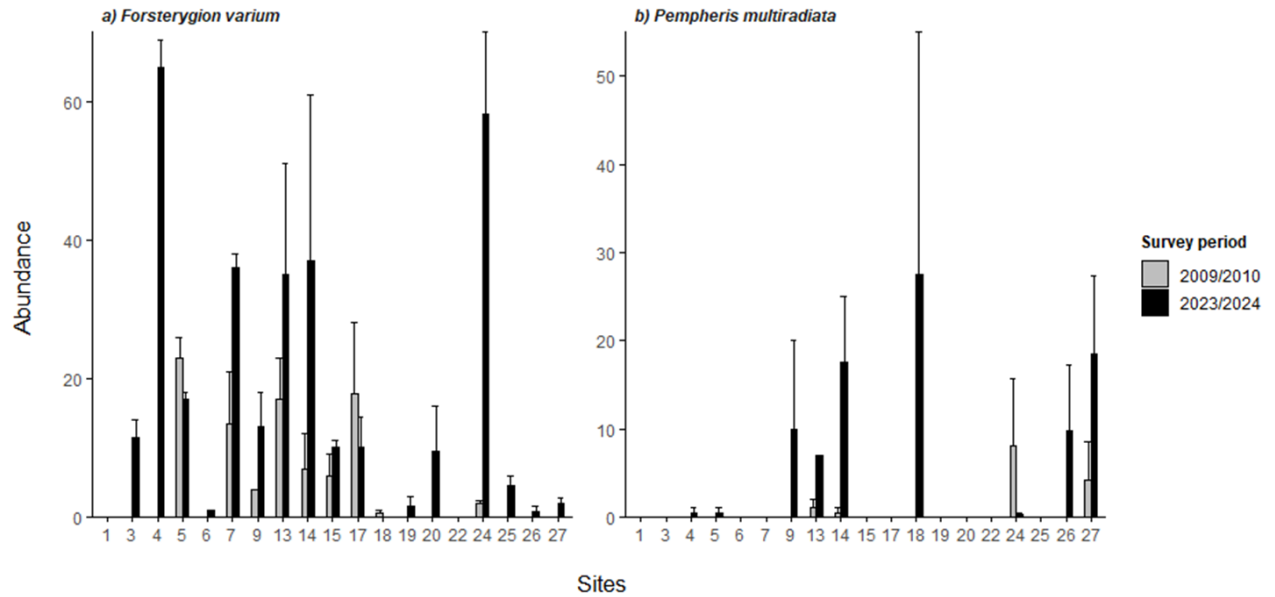


Figure 11: Abundance (mean $\pm$ SE) of the most abundant cryptic fish species per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black). The two species shown were among the top 4 most abundant in both survey periods, and are presented alphabetically.

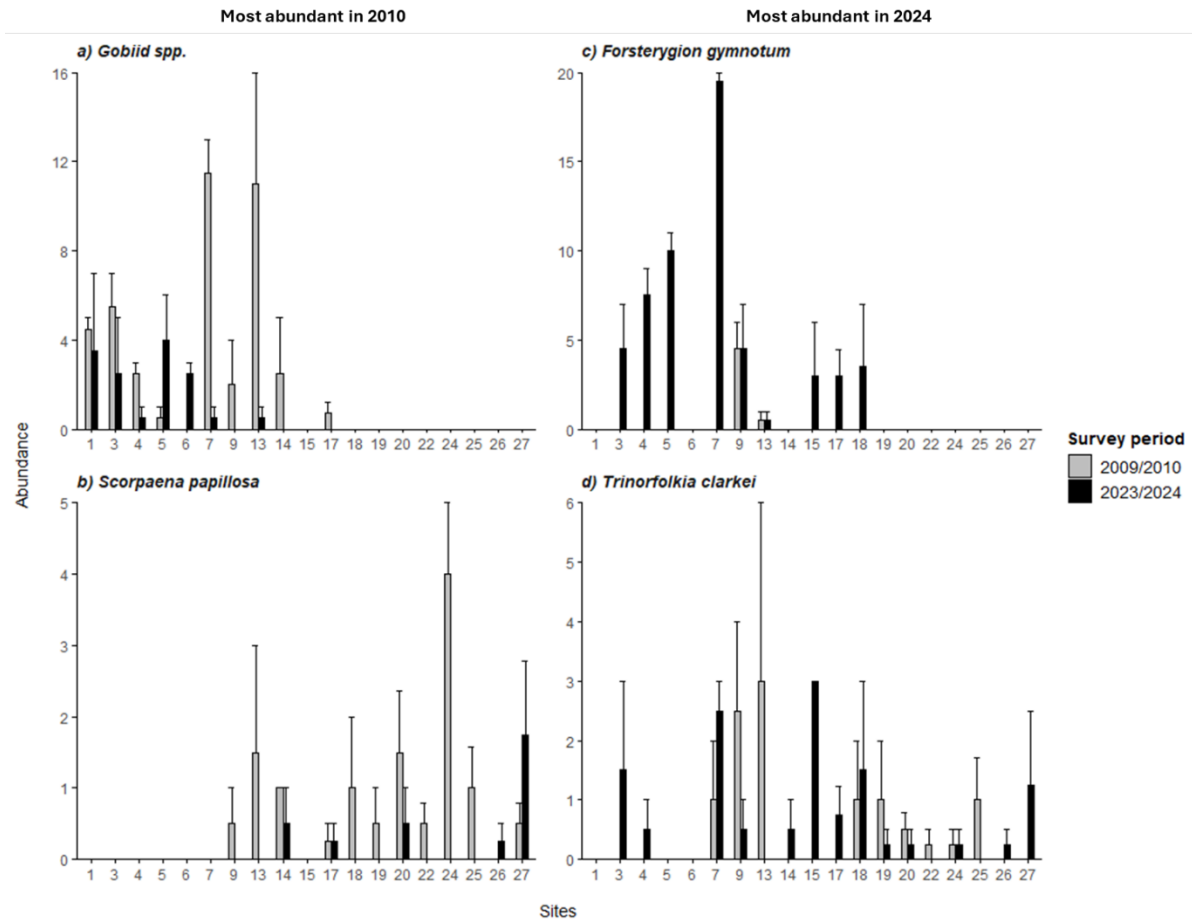


Figure 12: Abundance (mean $\pm$ SE) of cryptic fish species uniquely among the top 4 most abundant in 2010 (a, b) and 2024 (d, e), per transect (500 m²) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

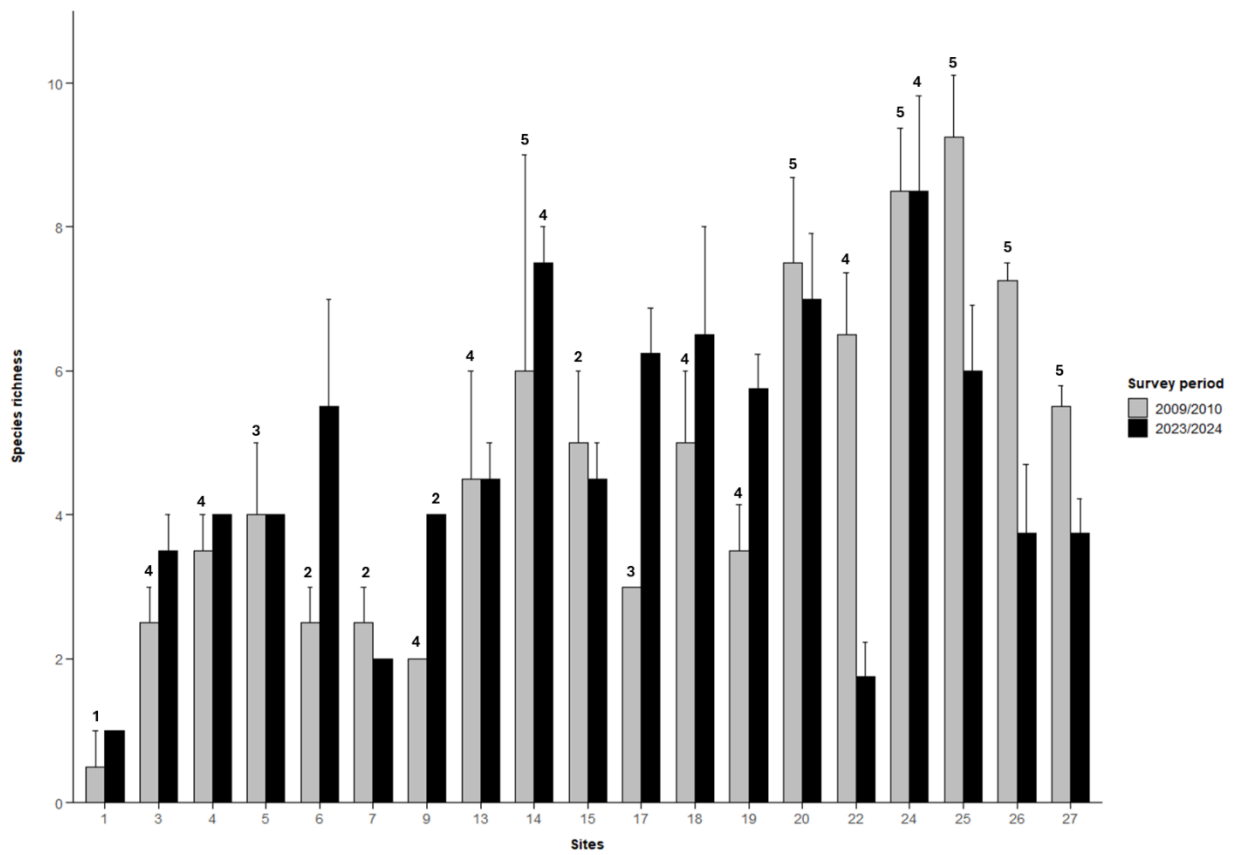


Figure 13: Species richness (mean $\pm$ SE) of mobile macro-invertebrates in 2010 (grey) and 2024 (black). Transect depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

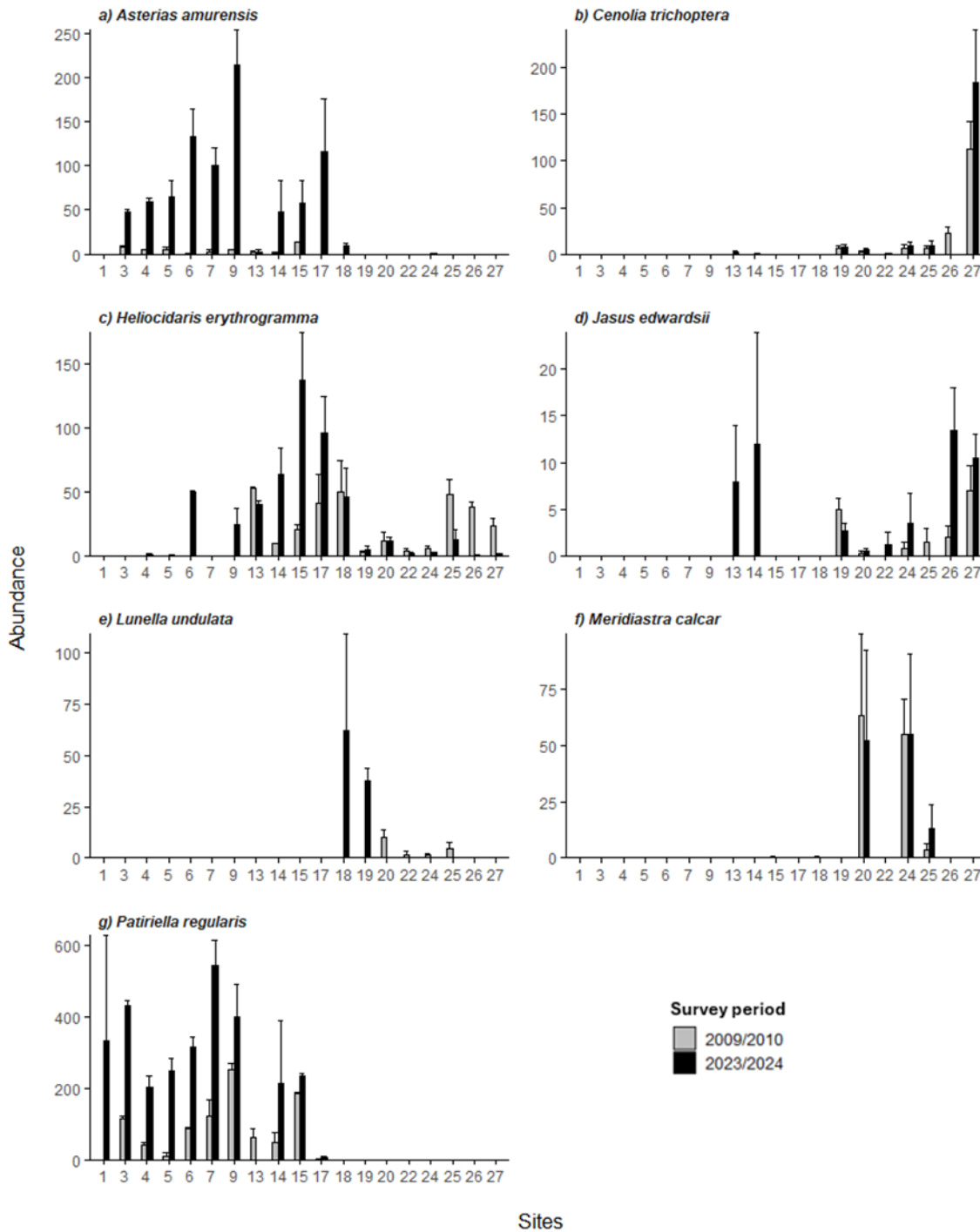


Figure 14: Abundance (mean $\pm$ SE) of the most abundant mobile macro-invertebrate species per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black). The seven species shown were among the top 10 most abundant in both survey periods, and are presented alphabetically.

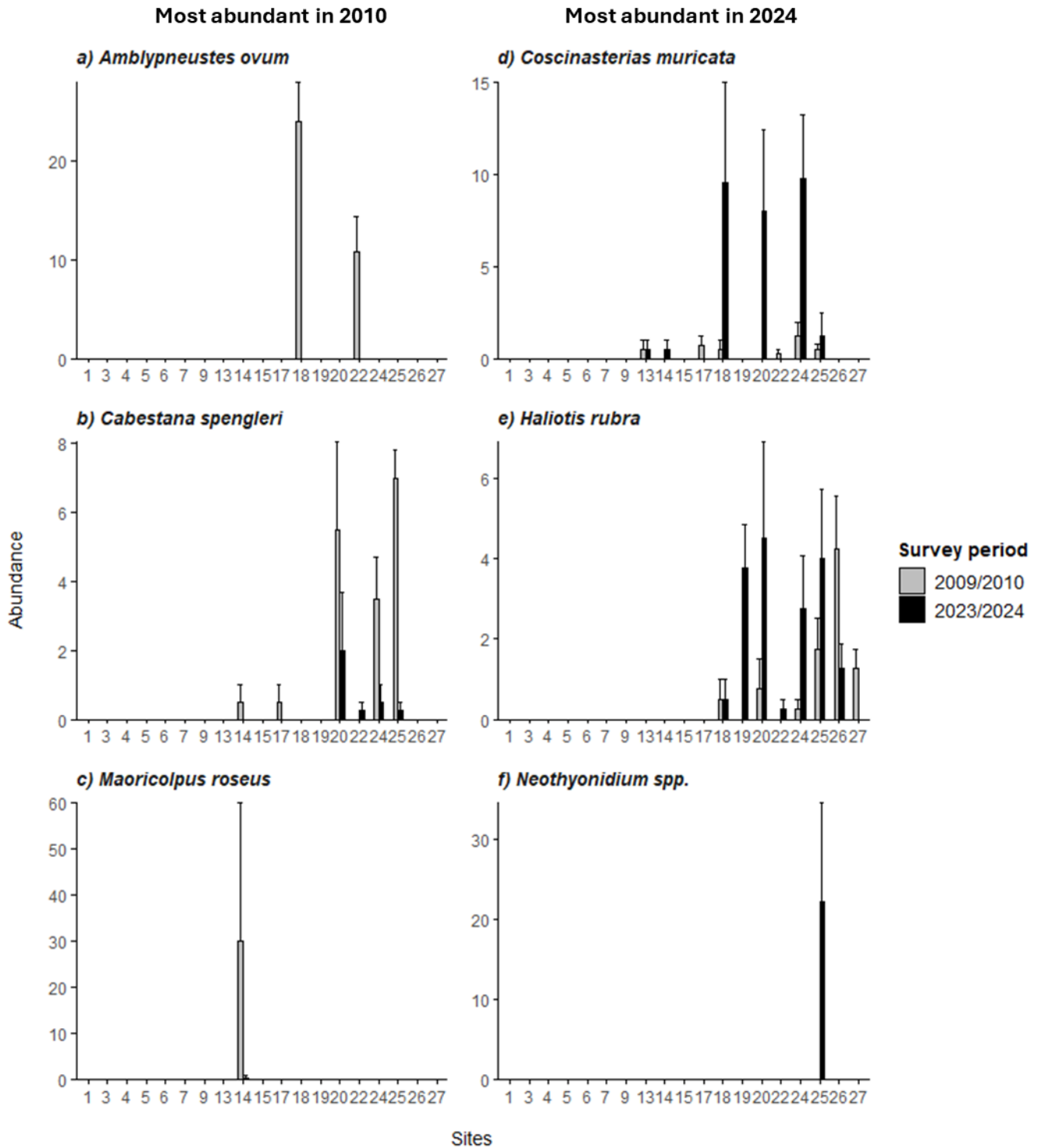


Figure 15: Abundance (mean $\pm$ SE) of mobile macro-invertebrate species uniquely among the top 10 most abundant in 2010 (a, b, c) and 2024 (d, e, f), per transect (500 m²) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

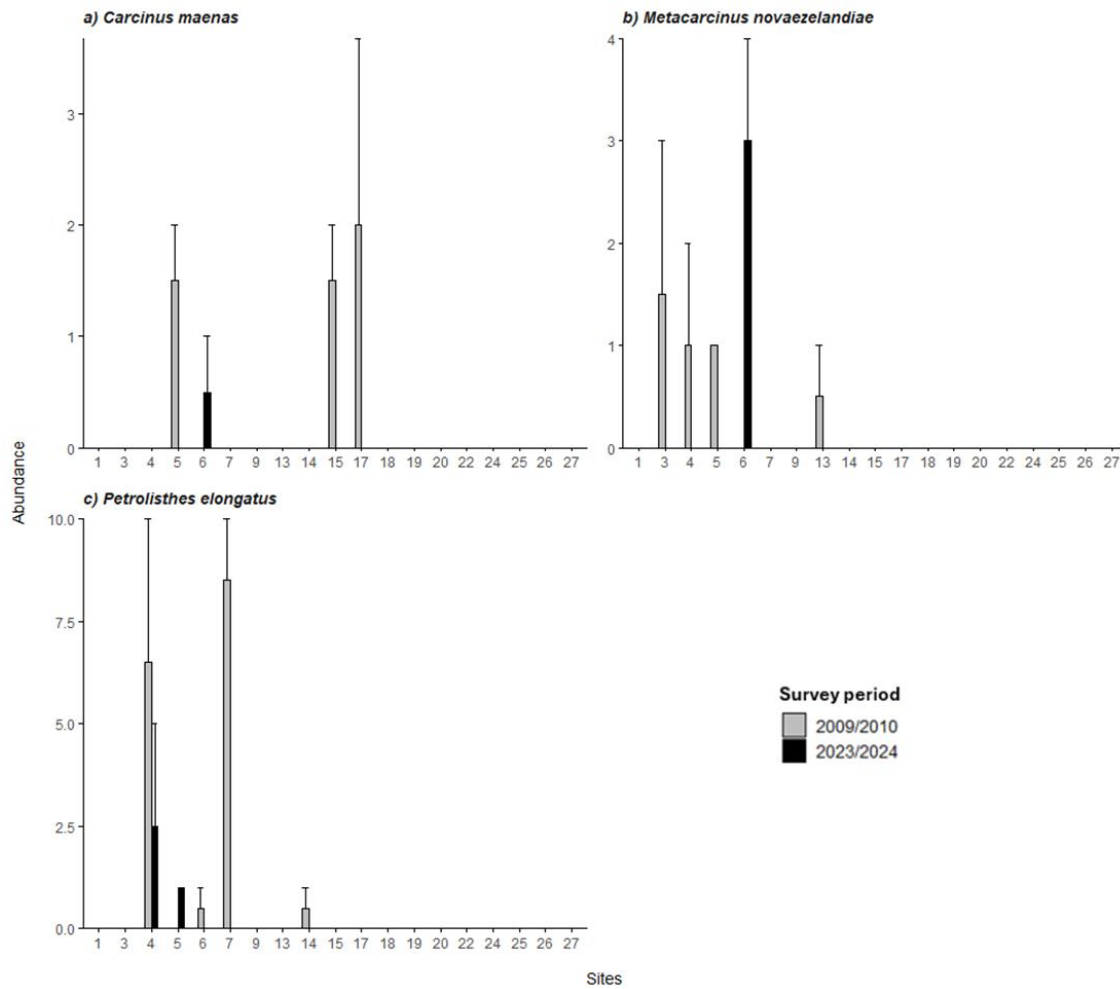


Figure 16: Abundance (mean $\pm$ SE) of three introduced macro-invertebrate species per transect (500 m²) in the Derwent in 2010 (grey) and 2024 (black).

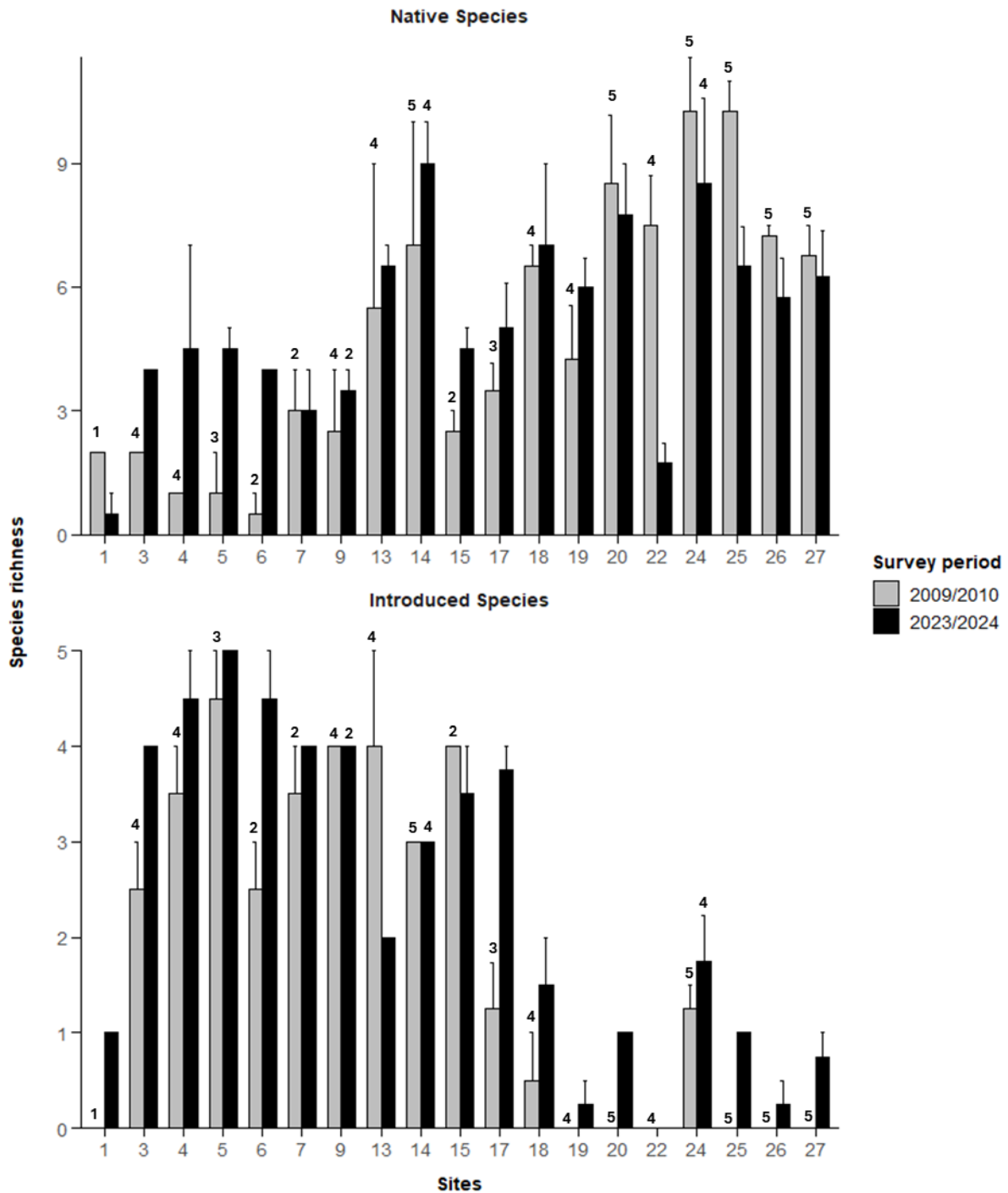


Figure 17: Species richness (mean $\pm$ SE) of native and introduced cryptic fish and cryptic invertebrates in 2010 (grey) and 2024 (black). Transect depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

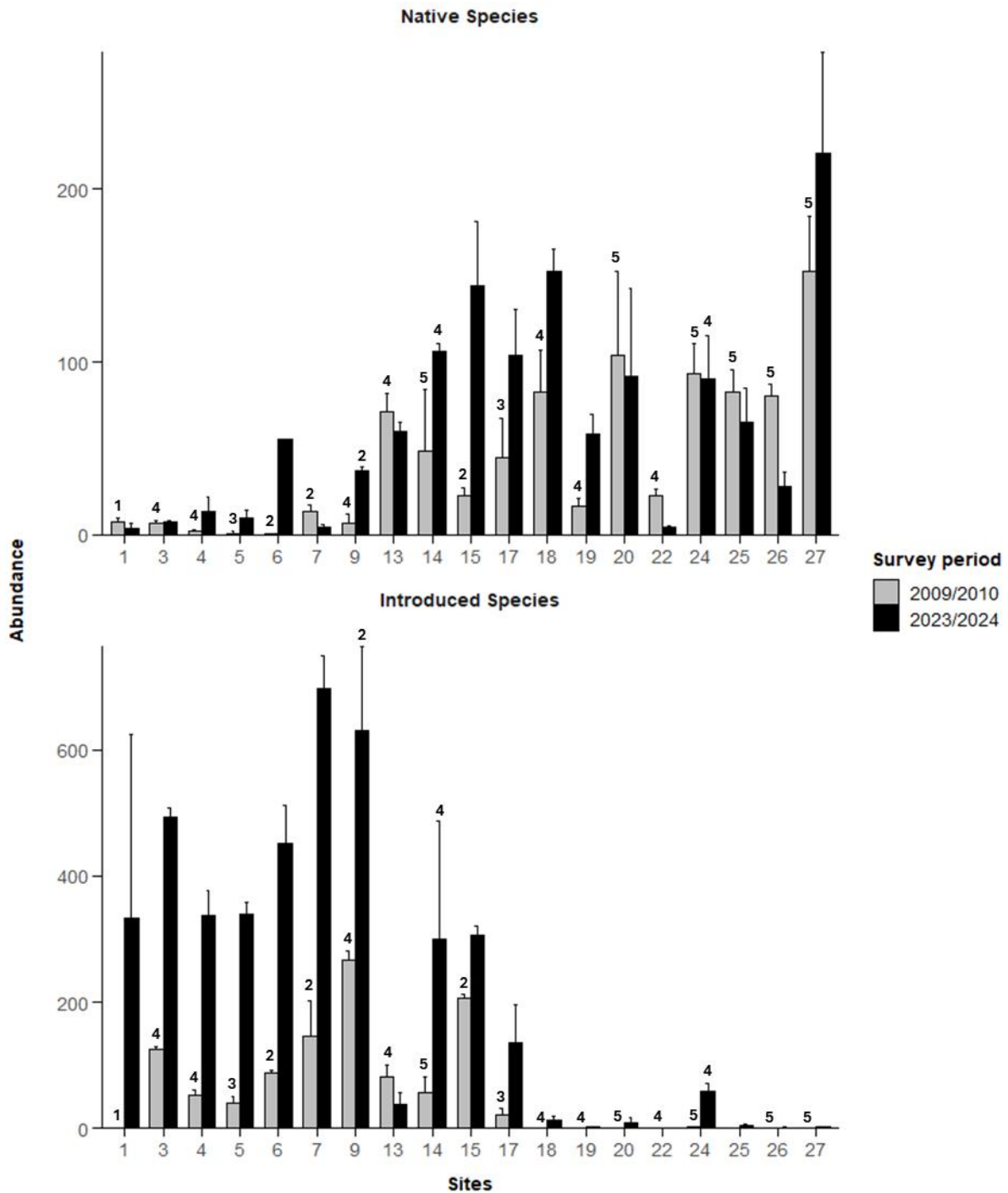


Figure 18: Abundance (mean $\pm$ SE) of native and introduced cryptic fish and cryptic invertebrates in 2010 (grey) and 2024 (black). Transect depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

Algae recorded on transect-based surveys

Sixty-eight algal species were recorded in 2024, with 50 species overlapping with the 2010 survey. Species richness and spatial distributions were generally similar between survey periods, with greater richness observed in the lower estuary (Figure 19). Percent cover in the

upper and mid-estuary (sites 1-16) increased compared to 2010, with almost all sites exceeding 50% coverage, despite lower species richness at these sites. Foliose algae followed a similar pattern, with increased richness in the lower estuary during both surveys (Figure 20). In 2024, foliose species were recorded in higher numbers at sites 3 – 17, including several sites (sites 3 – 6) where they were not previously observed. Percent cover at these upper estuary sites also increased, averaging between 40 and 100%.

The mean percentage cover of brown algae remained relatively consistent between surveys, although an increase was observed in the mid-estuary, particularly from sites 9 to 17 (Figure 21a). Green algae followed a similar trend, with generally minor differences between surveys, except for sites 6 and 15, where considerably greater coverage was recorded in 2024 (Figure 21c). In contrast, foliose red algae exhibited a shift in spatial distribution, with increased presence and coverage in several upper and mid-estuary sites (2 – 14), and decreased coverage further south in the lower estuary (sites 18 – 24; Figure 21b). Sessile invertebrate cover showed the most pronounced change between surveys, with a marked decrease in coverage across the upper and mid-estuary sites, where most averaged less than 20% in 2024, compared to 75% or greater in 2010 (Figure 21d). In the south, sessile invertebrate coverage varied minimally, with only sites 18, 20, and 22 in the lower estuary showing an increase in 2024.

Ecklonia radiata and *Carpoglossum confluens* (Figure 22c and 22b) were both consistently the most abundant brown algae across surveys, with increased coverage in 2024 for *E. radiata* across a majority of sites, and relatively consistent coverage for *C. confluens* from 2010 to 2024. *E. radiata*'s spatial distribution extended further north into sites 9, 13, 14, and site 17, where minimal to no coverage was observed in 2010. Minor differences across surveys were observed for the other three species recorded as most abundant in 2010 and 2024, with some small variations in site distributions. For example, *Sargassum fallax* had fluctuating percent cover across sites, but was observed at sites 13, 15, and 17 in 2024, despite not being present in 2010 (Figure 22e). Conversely, *Lessonia corrugata* was not recorded in 2024 at sites 19, 25, or 26, despite its presence at all three sites in 2010 (Figure 22d). *Acrocarpia paniculata* showed a similar pattern to 2010, with presence concentrated in the lower estuary and most sites recording an average of $\leq 15\%$ cover (Figure 22a).

Dictyopteris muelleri was among the top abundant species in 2010 but showed a decline in 2024, along with a shift in its spatial distribution (Figure 23a). It was recorded at sites 14, 17, 22, 25 and 27 in 2024, where it had not been observed previously. In contrast, *Halopteris paniculata* emerged as one of the top abundant brown algae species in 2024, with new occurrences at sites 7, 18, and 25 (Figure 23b).

We assessed the top six abundant red algal species across each survey period, identifying four species or groups that remained consistent across surveys. Due to the difficult nature of identifying red algal species, many unidentified species (e.g., unidentified algae (crustose coralline) were placed into their orders for reporting purposes (e.g., *Corallinales* spp.).

Peyssonnelia spp. and *Corallinales* spp. (Figure 24a and 24c) remained the two most abundant red algal taxa across both survey periods, with minimal shifts in their overall spatial distributions. Both were more abundant in the lower estuary, although *Peyssonnelia* spp. showed increased percent cover at several mid-estuary sites in 2024, including sites 9 – 15. However, percent cover for *Peyssonnelia* spp declined in general, evident particularly at sites

further south in the lower estuary (sites 17, 18, 20 and 22). In contrast, several lower estuary sites showed increased percent cover of *Corallinales* spp., with sites 19, 25, and 26 averaging over 30% in 2024, compared to 20% or lower in 2010. *Lenormandia marginata* similarly exhibited minimal spatial variation, with occurrences restricted to site 19 and further south for both surveys, and with little change in percent cover (**Error! Reference source not found.b**). *Rhodophyta* spp., which includes Unidentified algae (red turf) and (filamentous red), had considerable variation across survey periods (Figure 24d). No patterns across time were discernible, although most coverage averaged below 25%, regardless of survey period, except for sites 1 and 15.

Plocamium angustum and *Pollexfenia lobata* (

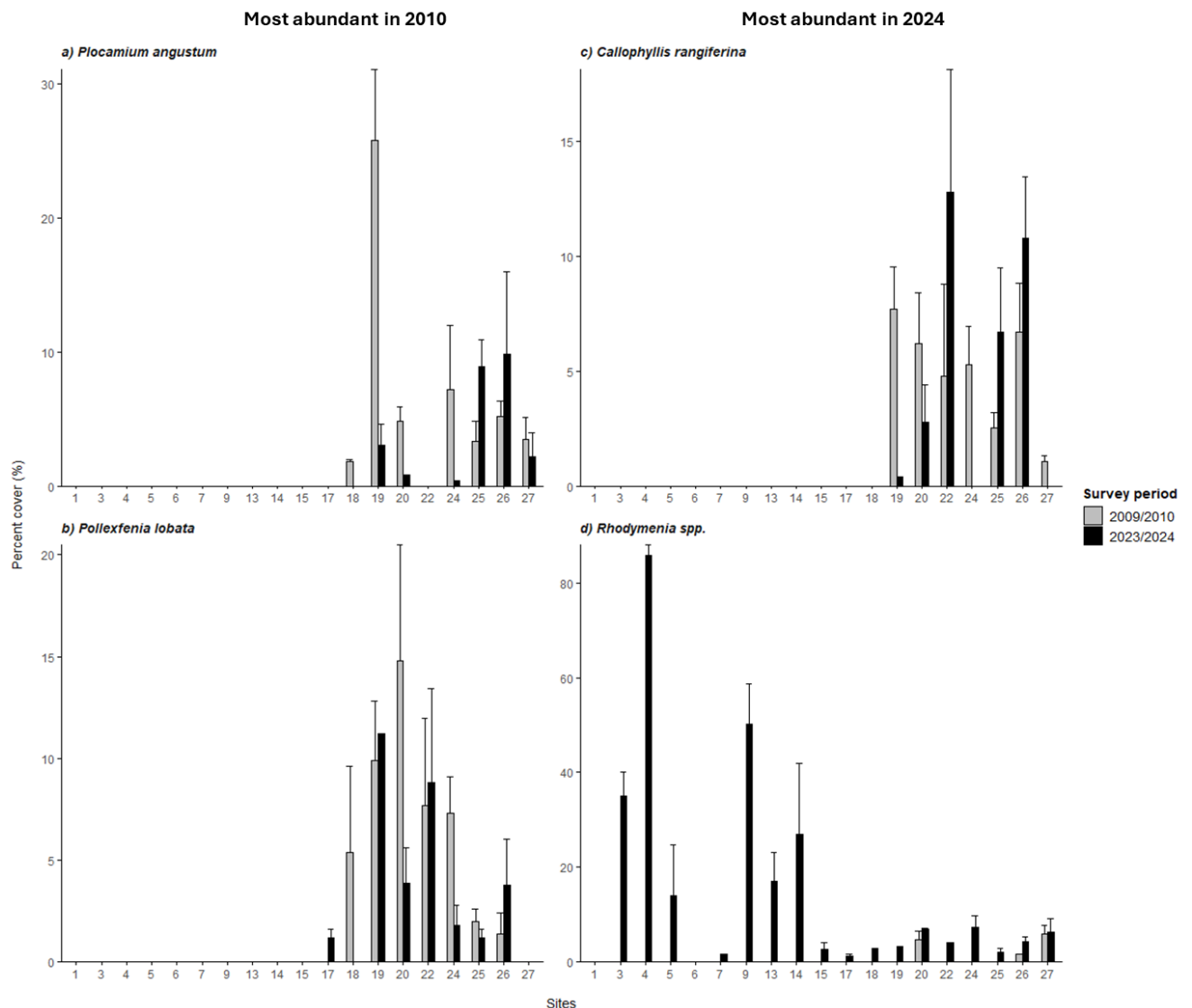


Figure 25: Percent cover (mean $\pm$ SE) of red algae species uniquely among the top 6 most abundant in 2010 (a, b) and 2024 (c, d) in the Derwent. While each species peaked in percent cover during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

a and 25b) were both among the top six most abundant species in 2010, with minimal range shifts observed in 2024. Overall, most sites showed a decline in percent cover for both species. However, minor increases were recorded for *P. angustum* at sites 25 and 26, and for

P. lobata at sites 19, 22, and 26. Mean coverage in 2024 remained at or below 10% across all sites for both species, with the exception of site 19 for *P. lobata* (11%). In comparison, *Callophyllis rangiferina* and *Rhodymenia* spp. (

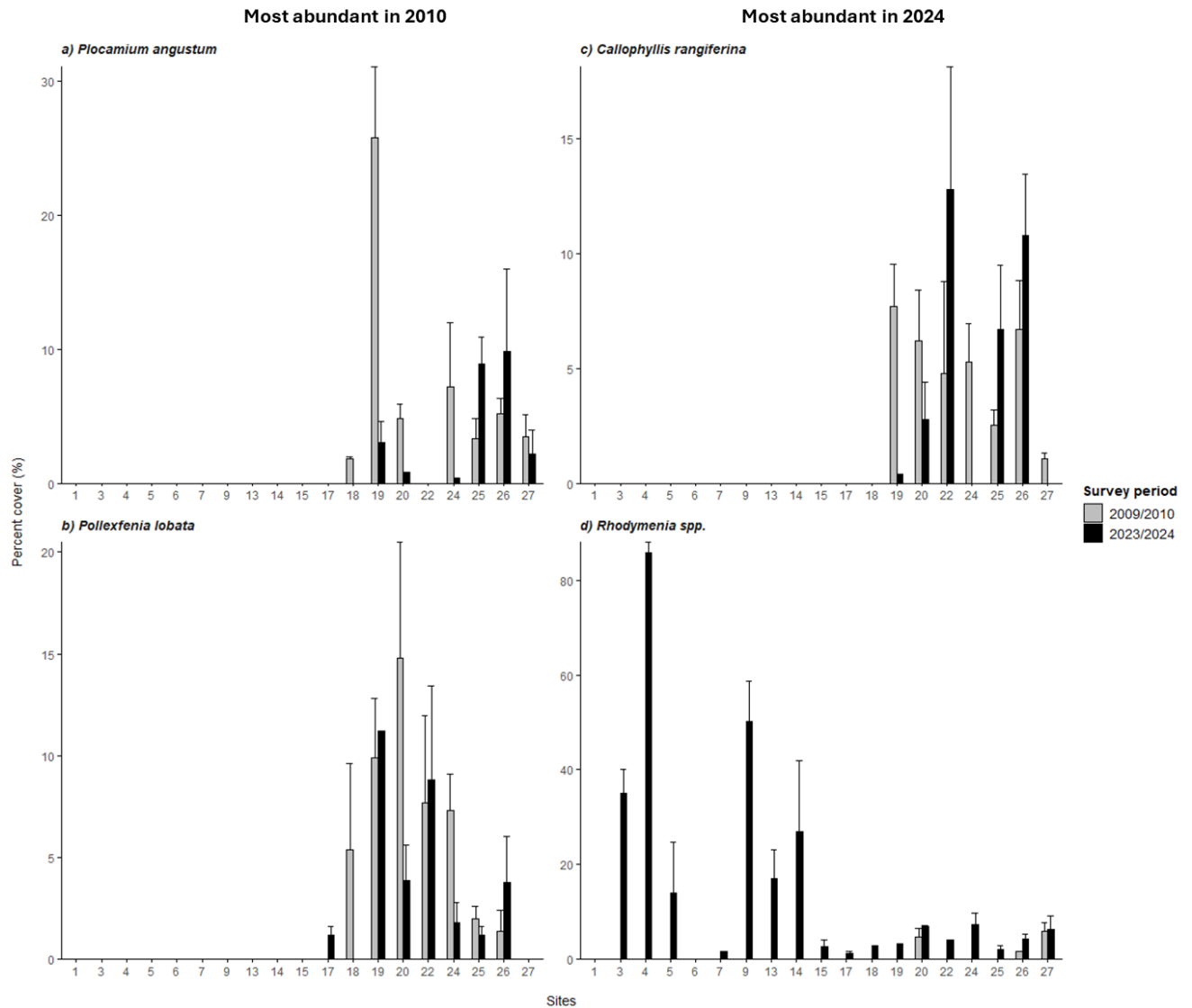


Figure 25: Percent cover (mean $\pm$ SE) of red algae species uniquely among the top 6 most abundant in 2010 (a, b) and 2024 (c, d) in the Derwent. While each species peaked in percent cover during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

c and 25d) were both unique to the top six abundant species in 2024, showing a notable increase since 2010. Coverage of *Rhodymenia* spp. was quite high in 2024, averaging above 20% for four sites (3, 4, 9, 14) and over 80% at site 4. Most of the *Rhodymenia* in the upper to mid-estuary in 2024 was identified as *R. leptophylla*, which despite reaching covers in excess of 80% at site 4, and more than 50% at site 9 in 2024, was absent at many of these locations in 2010. This represents a major system change given the dominance of this species in 2024 relative to 2010.

In contrast, the increase in *Callophyllus rangiferina* was in the lower estuary, with less marked change relative to 2010, with an average cover of around 15% at sites 22 and 26, in 2024.

Of the three introduced seaweeds (two red and one brown), only *Grateloupia turuturu* increased in percent cover and spatial distribution in 2024, occurring at six new sites across the mid and lower estuary (Figure 26b). Both *Aeodes nitidissima* and *Undaria pinnatifida* (Figure 26a and 26c) were rarely observed in 2024, present at only sites 7 and 9 (*A. nitidissima*) and sites 19, 22, and 25 (*U. pinnatifida*).

Sessile invertebrates recorded on transect based surveys

Eighteen species of sessile invertebrates were recorded in 2024, including 11 that overlapped with the 2010 survey, and seven new, unique species.

Mytilus galloprovincialis showed a notable increase in coverage in 2024, occurring at nine sites, including four with 25 – 50% coverage and five with up to 10% (Figure 27b). This marks a substantial increase in both abundance and spatial distribution compared to 2010, when it was present at only three sites. *Porifera* spp. showed an overall decline in percent cover from 2010, with most sites recording under 20% in 2024 (Figure 27d). However, its range expanded into the upper and mid-estuary, with new occurrences from sites 3 – 13, indicating a northward shift in distribution despite the reduced abundance. Similarly, *Ostrea angasi* became more prevalent in the upper and mid-estuary, with new occurrences at sites 3, 5, 7, 13, 14, and 17 – none of which recorded the species in 2010 (Figure 27c). Despite this shift, percent cover remained relatively consistent across surveys, generally ranging between 5 – 10%, and never exceeding 15%. The peak cover was reported at site 9, adjacent to the Tasman bridge on the eastern shore, with species identification validated during a subsequent targeted survey (see later). In contrast, *Bryozoa* spp. showed no notable range extension, remaining concentrated in the lower estuary sites (sites 25 – 27) in 2024, although in greater abundance compared to 2010 (Figure 27a).

Animalia spp., a cover-all category for unidentified sessile invertebrates, primarily composed of Unidentified invertebrate (tube matting) was the most abundant sessile invertebrate in 2010, averaging over 50% cover for all sites where it was present (Figure 28a). However, *Animalia* spp. was not recorded at any sites in 2024, suggesting key ecological changes in the upper estuary. Likewise, *Pyura gibbosa*, which was previously abundant in 2010, declined in cover by 2024 – no longer occurring at site 24, and instead shifting further north into sites 19 and 20 (Figure 28b). In contrast, *Ascidacea* spp. increased in percent cover in 2024, albeit at low levels (1-2% per site; Figure 28c). A shift in spatial distribution was also evident, with greater occurrence at site 19 and sites 25 – 27, and a lack of presence at site 20 where it had previously been recorded. The introduced *Crassostrea gigas* was also among the most abundant sessile invertebrates in 2024, but its spatial range narrowed considerably (Figure 28d). Previously recorded at four sites (each with $\leq 5\%$ cover), it was only found at site 1, where it reached an average cover of 14%.

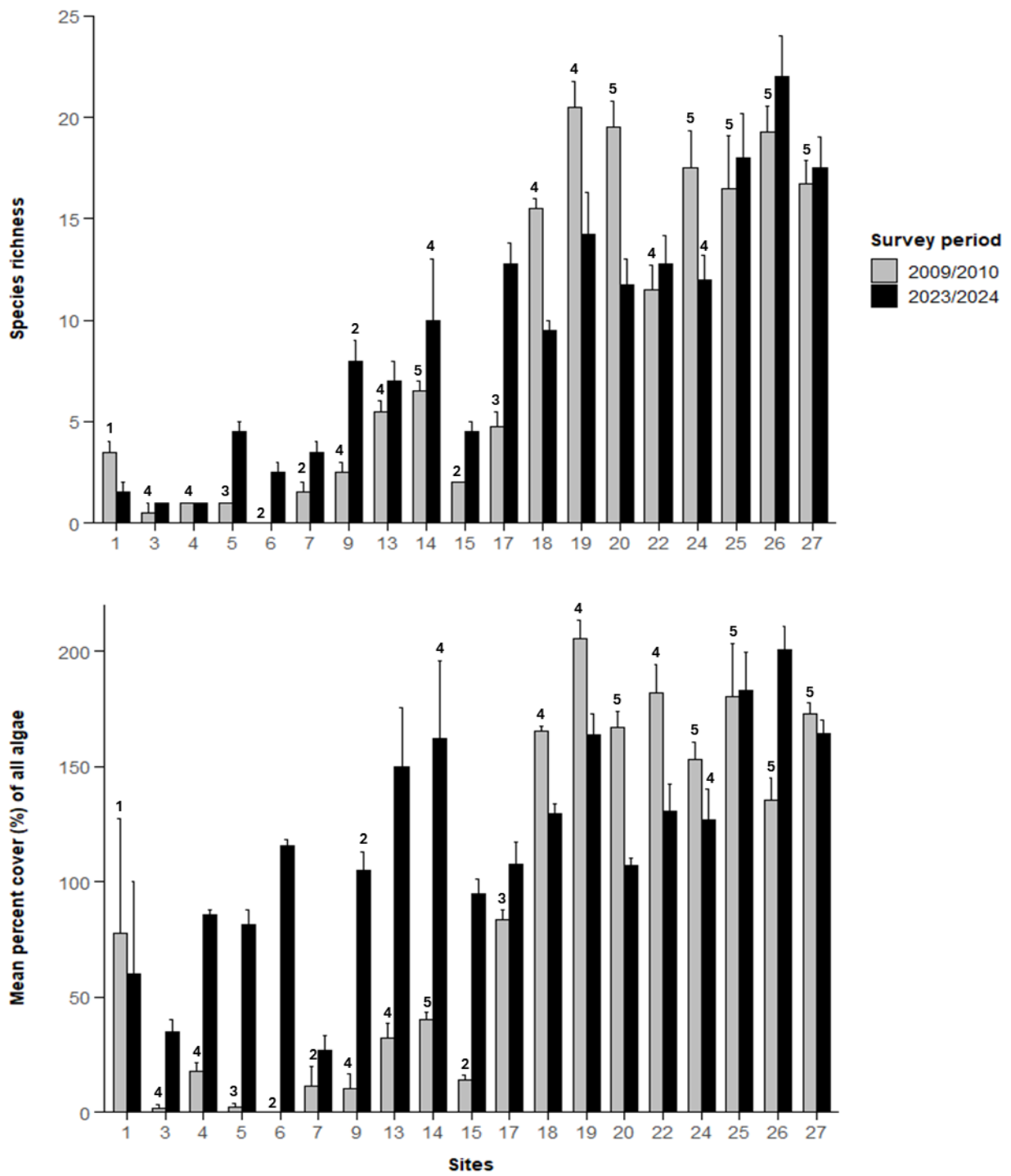


Figure 19: Species richness (mean $\pm$ SE) and percent cover (mean $\pm$ SE) of algae in 2010 (grey) and 2024 (black). Survey depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

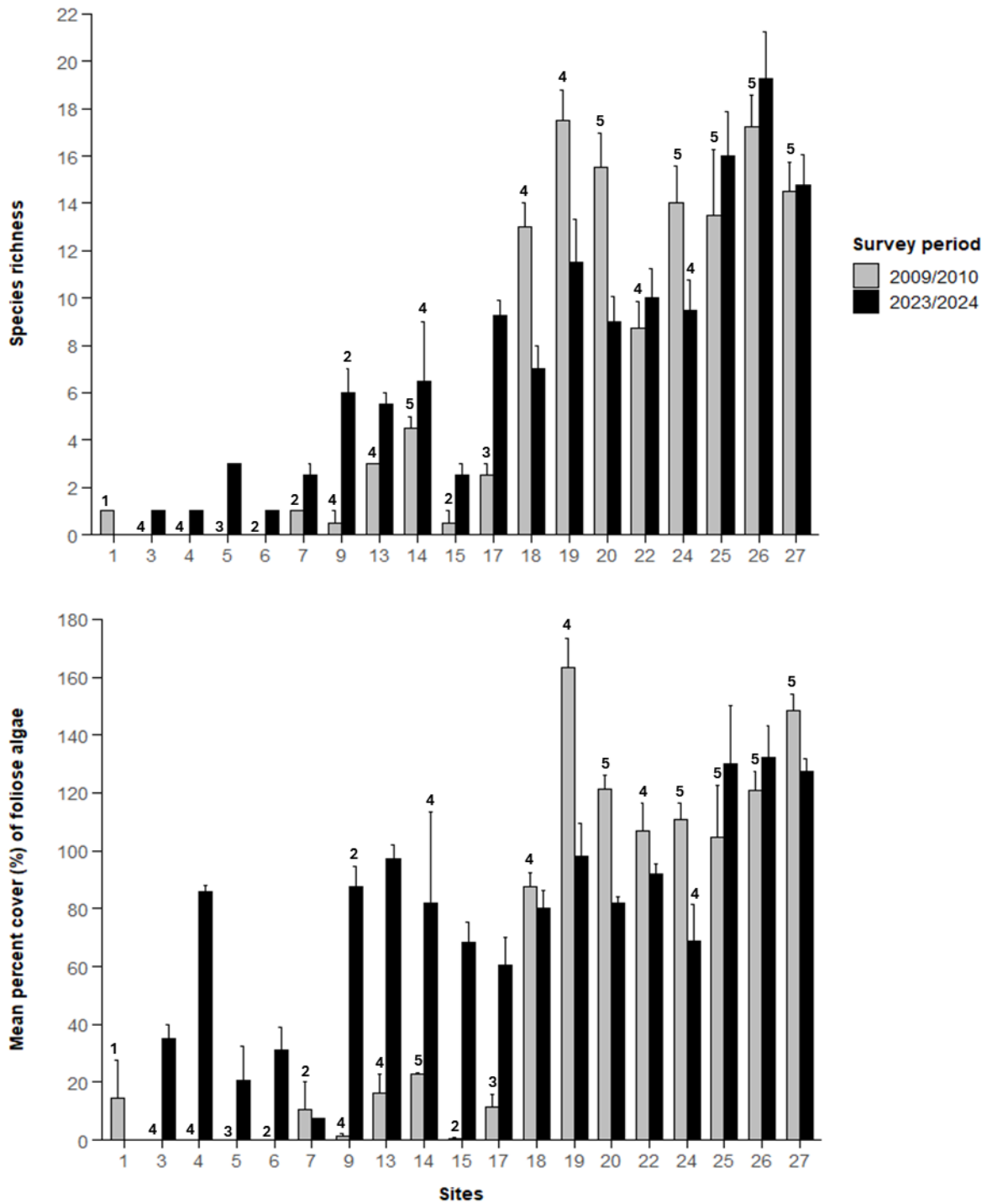


Figure 20: Species richness (mean $\pm$ SE) and percent cover (mean $\pm$ SE) of foliose algae in 2010 (grey) and 2024 (black). Survey depths at each site are indicated by the number above the grey bars, with sites 9, 14, and 24 differing in depths across years.

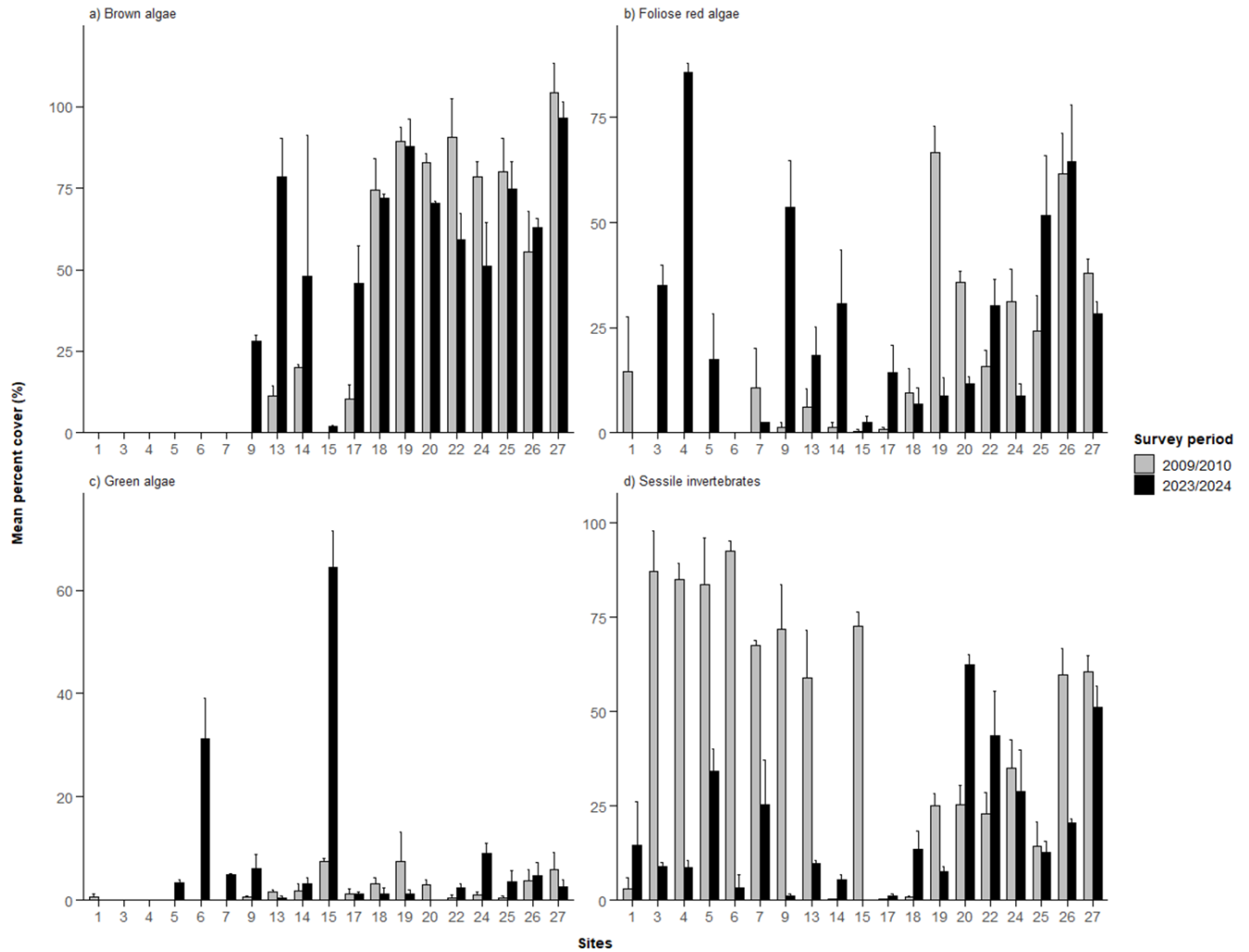


Figure 21: Percent cover (mean $\pm$ SE) of brown (a), foliose red (b), green (c) algae, and sessile invertebrates (d) in the Derwent across survey periods (2010: grey, 2024: black).

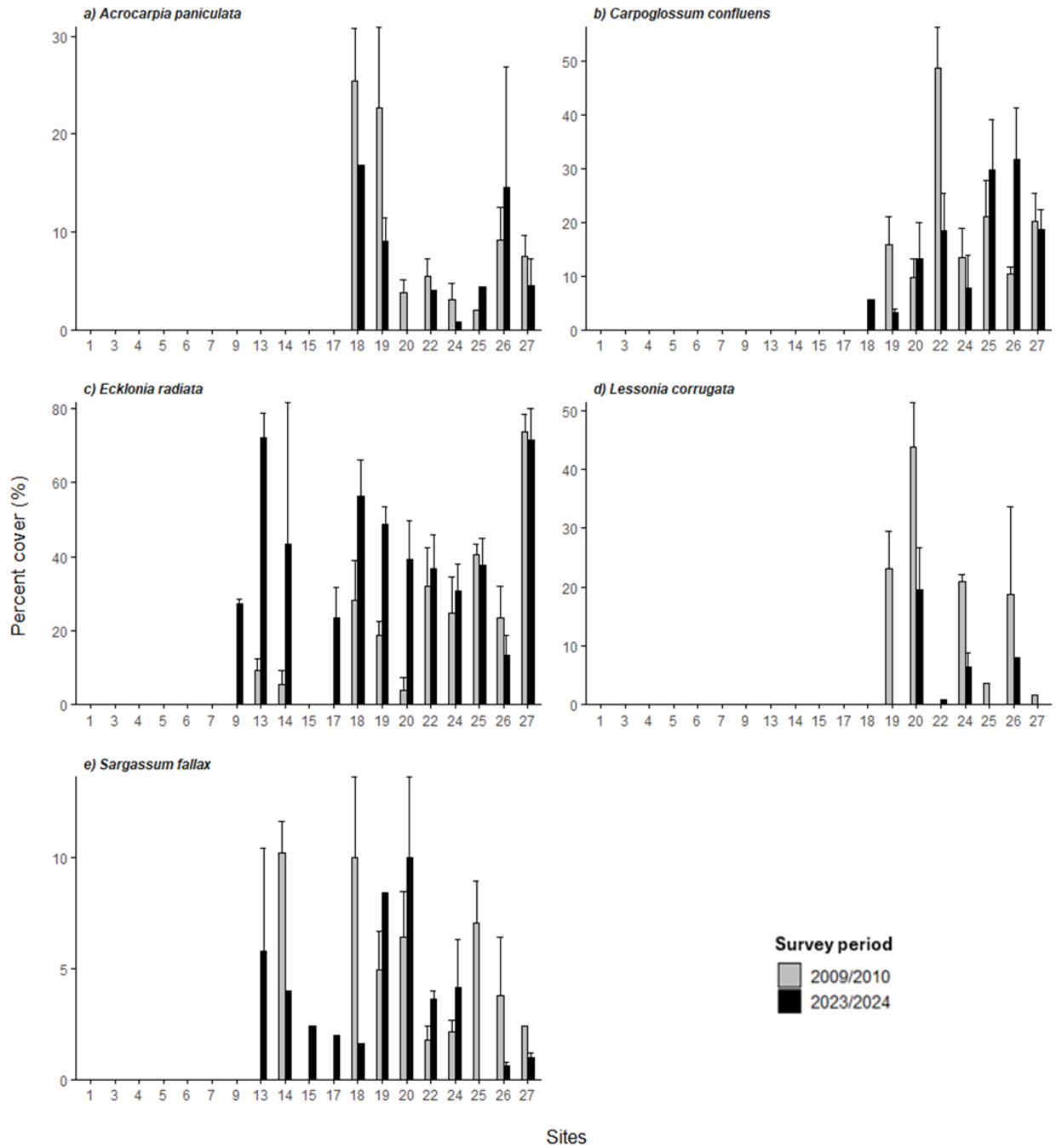


Figure 22: Percent cover (mean $\pm$ SE) of the most abundant brown algae in the Derwent in 2010 (grey) and 2024 (black). The five species shown were among the top 6 most abundant in both survey periods, and are presented alphabetically.

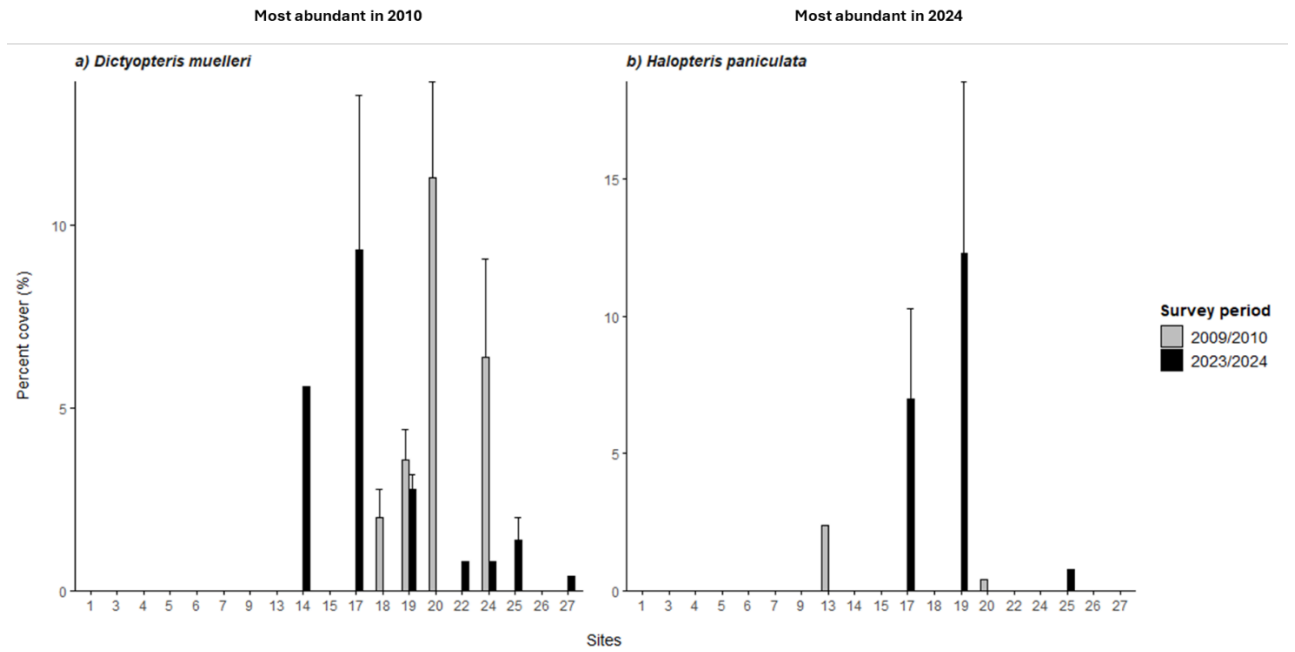


Figure 23: Percent cover (mean $\pm$ SE) of brown algae species uniquely among the top 6 most abundant in 2010 (a) and 2024 (b) in the Derwent. While each species peaked in percent cover during one period, changes across both survey periods are indicated by grey and black fill.

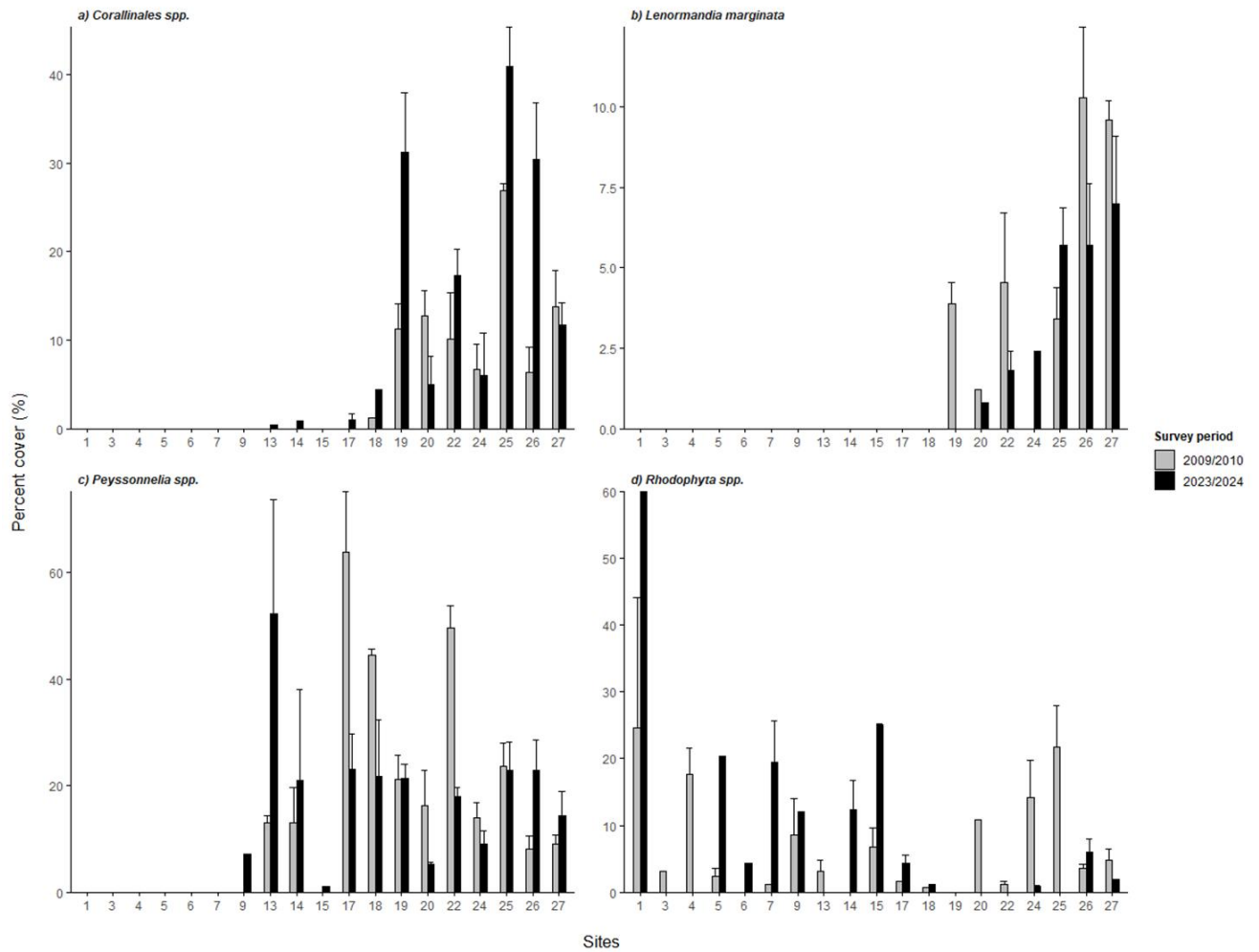


Figure 24: Percent cover (mean $\pm$ SE) of the most abundant red algae in the Derwent in 2010 (grey) and 2024 (black). The four species shown were among the top 6 most abundant in both survey periods, and are presented alphabetically. Here, *Rhodophyta spp.*, comprises unidentified algae (red turf) and filamentous red algae.

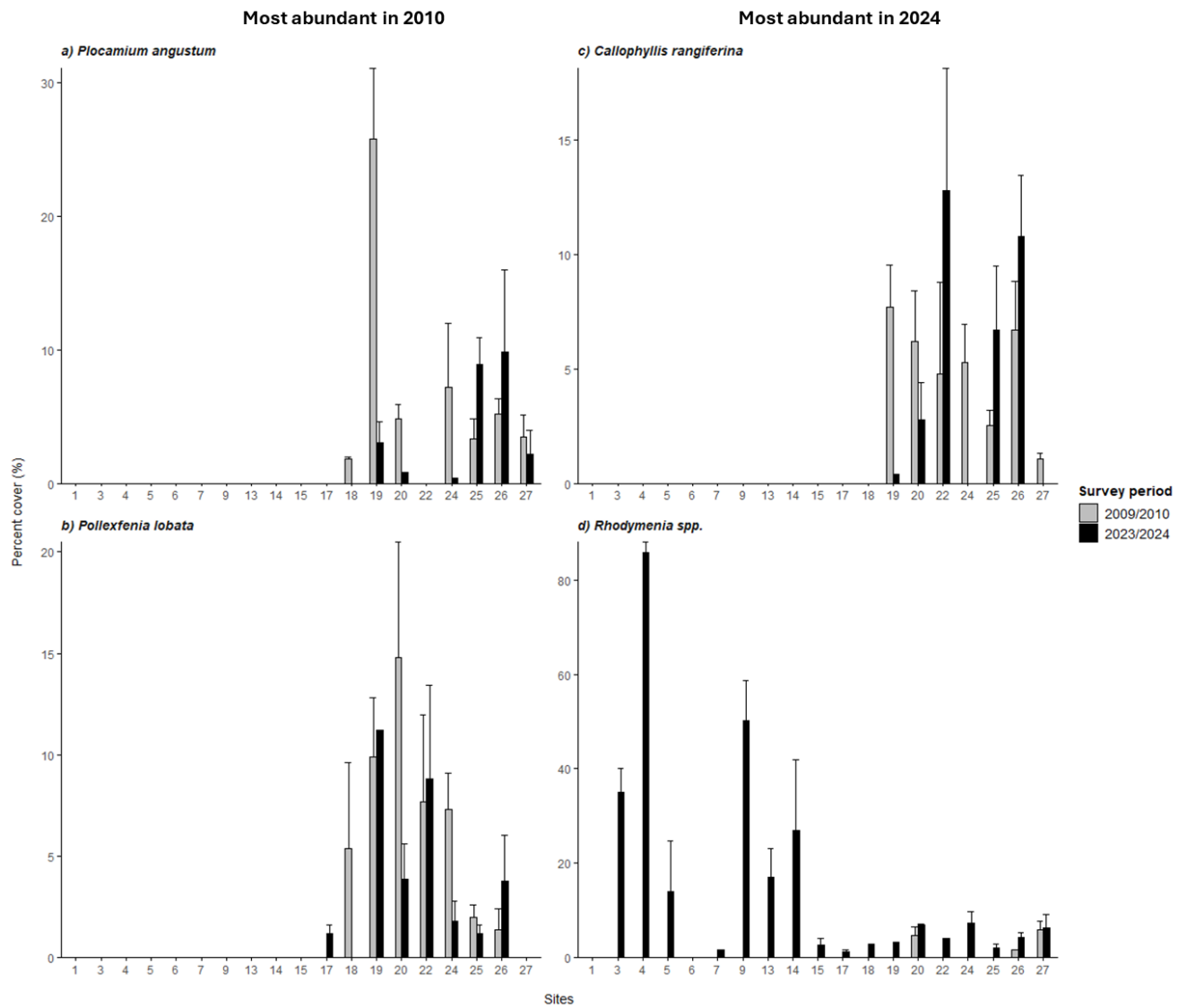


Figure 25: Percent cover (mean $\pm$ SE) of red algae species uniquely among the top 6 most abundant in 2010 (a, b) and 2024 (c, d) in the Derwent. While each species peaked in percent cover during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

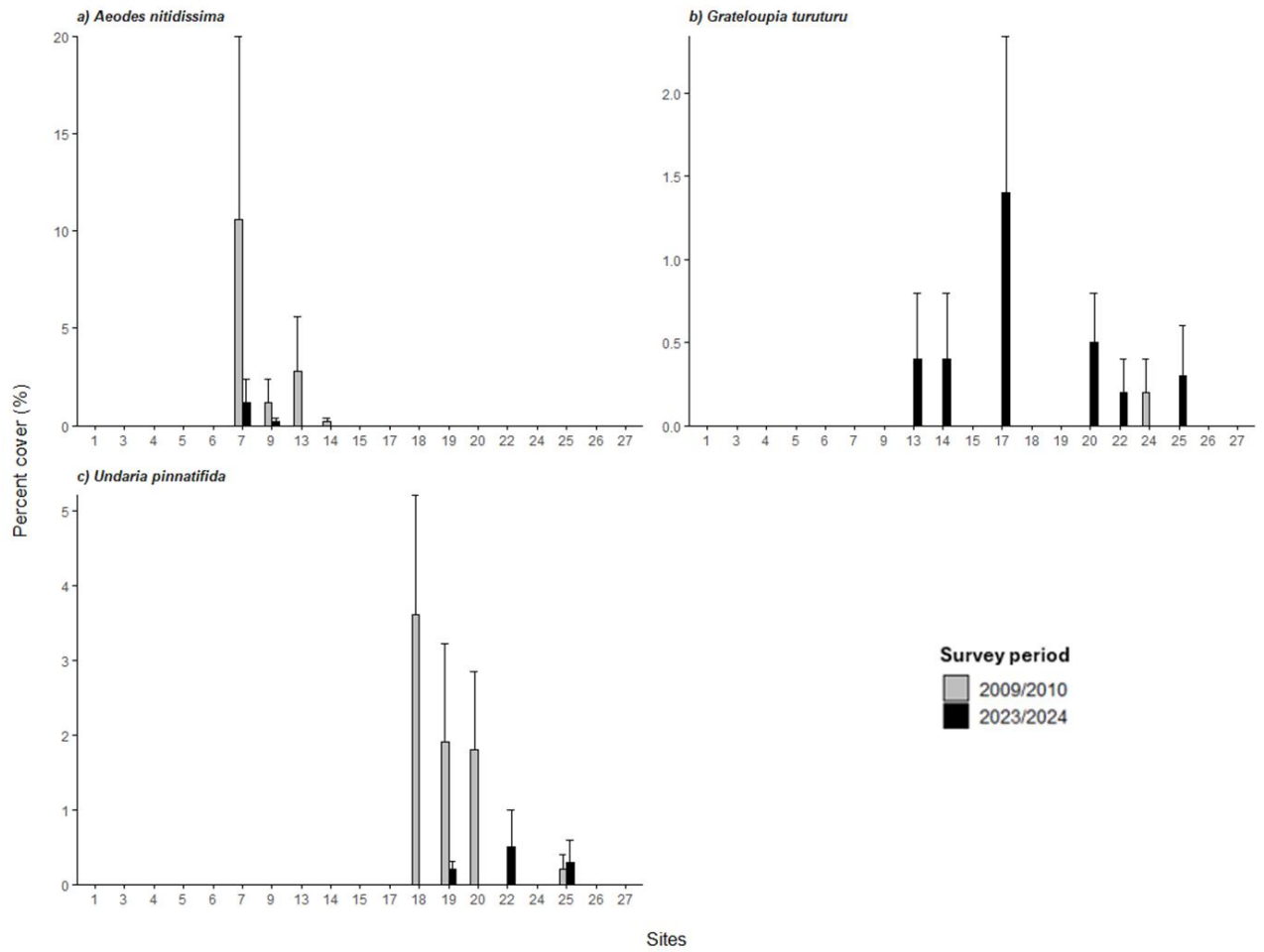


Figure 26: Percent cover (mean $\pm$ SE) of three introduced algae species in the Derwent in 2010 (grey) and 2024 (black). Species are presented alphabetically.

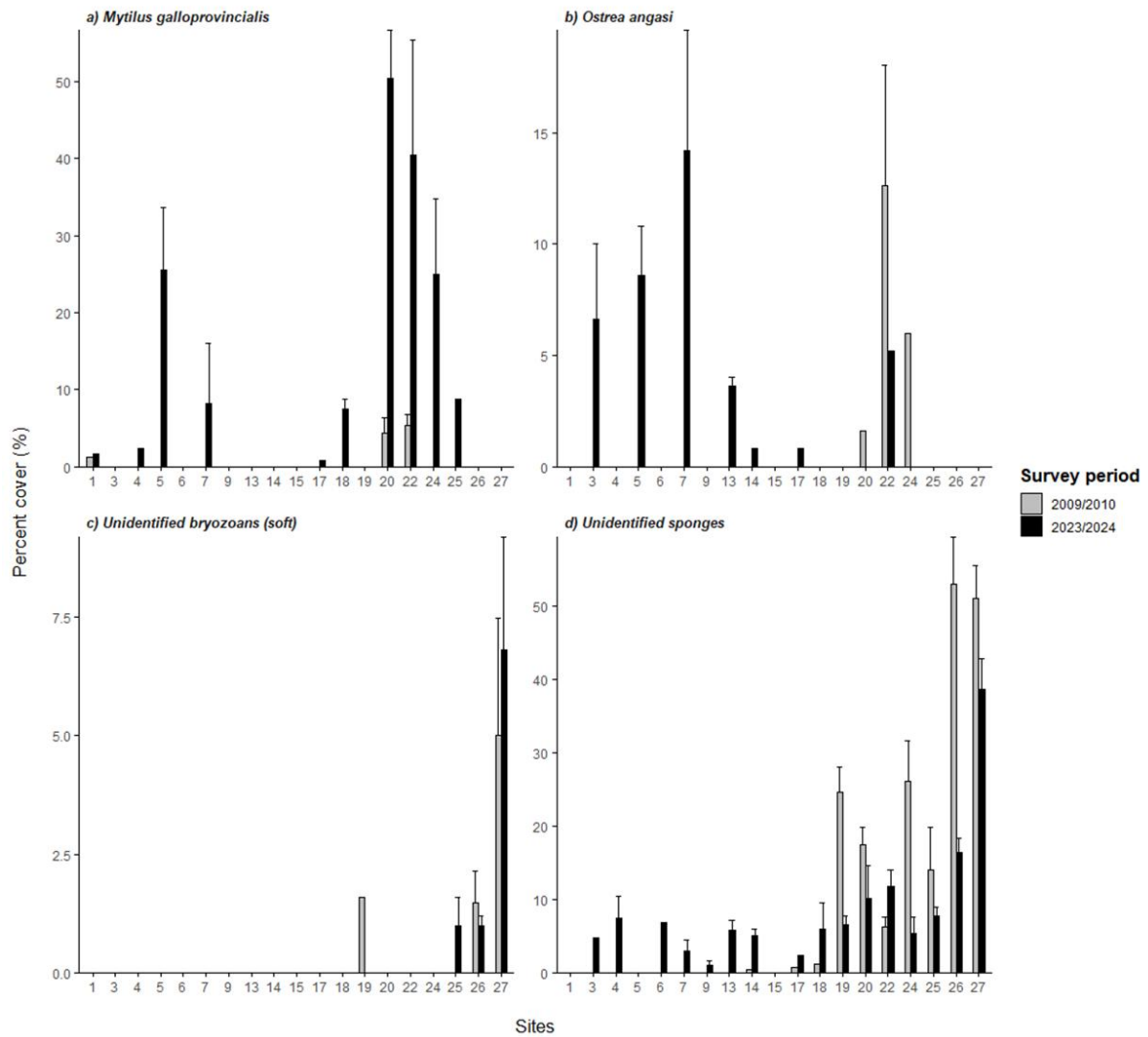


Figure 27: Percent cover (mean $\pm$ SE) of the most abundant sessile invertebrates in the Derwent across survey periods (2010: grey, 2024: black). The four species shown were among the top 6 most abundant in both survey periods, and are presented alphabetically.

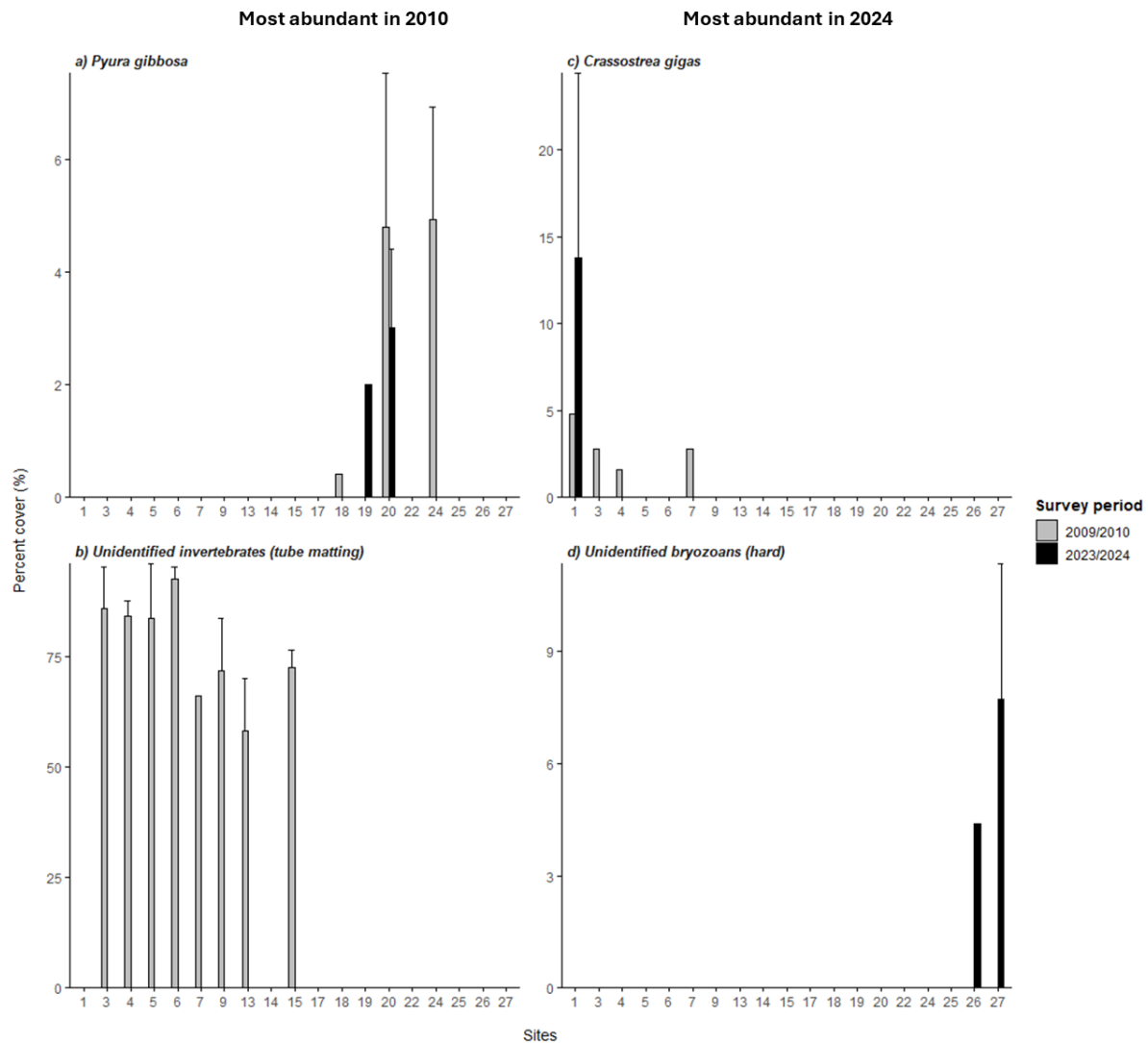


Figure 28: Percent cover (mean $\pm$ SE) of sessile invertebrate species uniquely among the top 6 most abundant in 2010 (a, b) and 2024 (c, d) in the Derwent. While each species peaked in percent cover during one period, changes across both survey periods are indicated by grey and black fill. Species are presented alphabetically.

3.2 Timed intertidal and subtidal surveys

General patterns

Overall, species richness of fish, invertebrates, and algal species remained relatively consistent across survey periods, with a slight decline observed in 2024. Richness was generally highest in reef habitats, and followed a similar pattern across intertidal, 0-1 m, and 1-4.9 m zones (**Error! Reference source not found.**). Species counts tended to increase from north to south, as seen in the standard surveys, with the highest richness recorded in the lower estuary.

Introduced species richness in 2024 mostly mirrored that recorded in 2010, with a slight overall increase, particularly on reef habitats (Figure 30). The highest counts of introduced species were observed in the upper and mid-estuary at 0-1 m and 1-4.9 m. In the intertidal

zone, distribution followed a similar curve, with reduced counts in two upper estuary sites (sites 1 and 2) and in the lower estuary, and relatively stable higher counts throughout the mid-estuary – although some sites varied in this pattern.

Species considered least abundant or rarely encountered (defined as species recorded with only a single individual at a site) had considerable variability across the estuary (**Error! Reference source not found.**). Overall, species richness declined slightly but followed similar spatial patterns to the 2010 baseline. Richness tended to increase in the lower estuary, between 1-8.9 m depth, across both reef, ecotone and sand habitats. Majority of recordings from sand habitats were concentrated in the 5-8.9 m zone, versus reef habitats, where most observations were from intertidal to 4.9 m.

Species richness was assessed across rank abundance scores (1-5) for taxonomic groups in Reef, Ecotone, and Sand habitats (**Error! Reference source not found.**) and at different depth zones (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, and 9-15 m; **Error! Reference source not found.**). No clear trend through time could be determined, nor any obvious pattern for taxonomic groups.

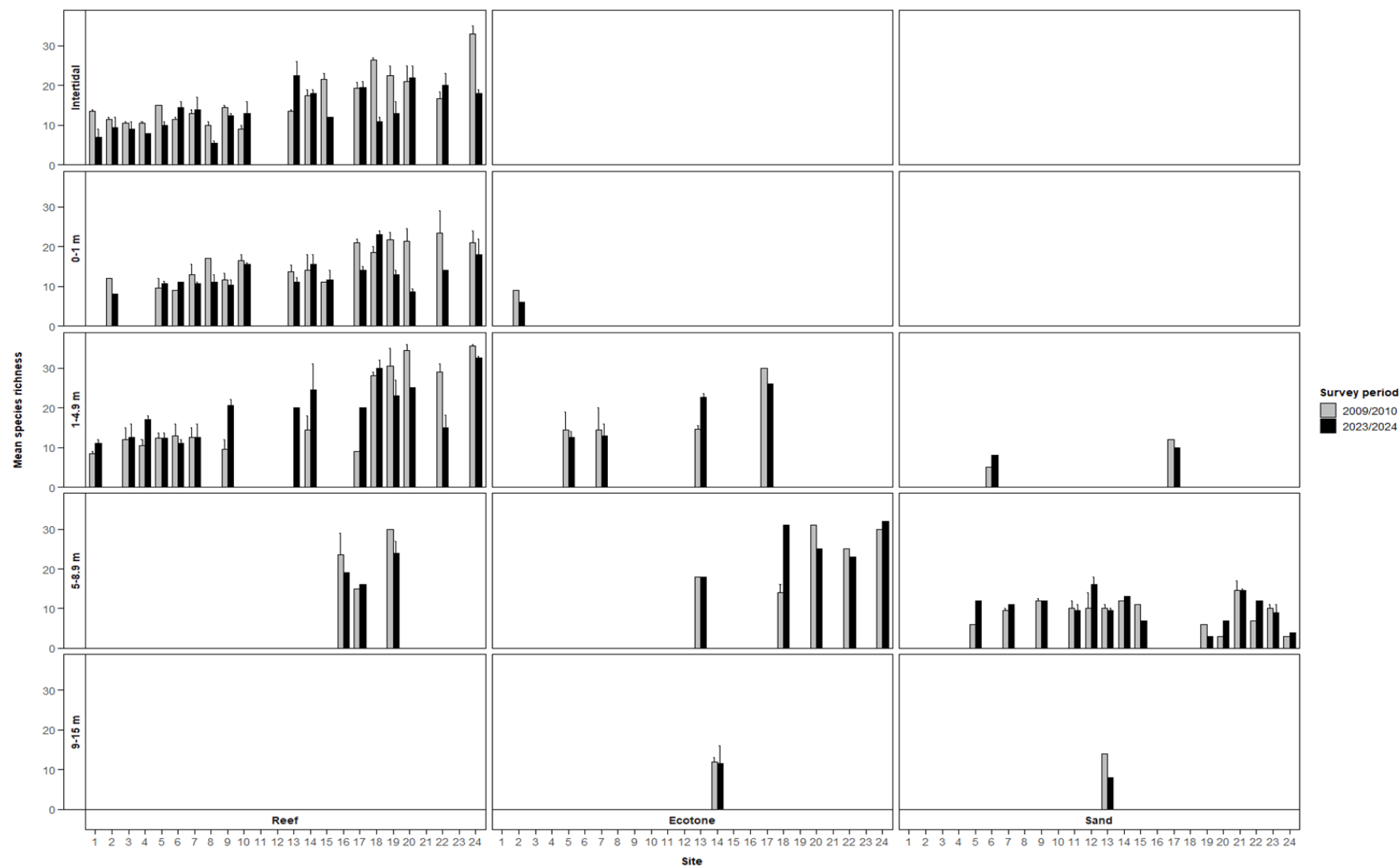


Figure 29: Species richness (mean $\pm$ SE) of all fish, invertebrate and algal species recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

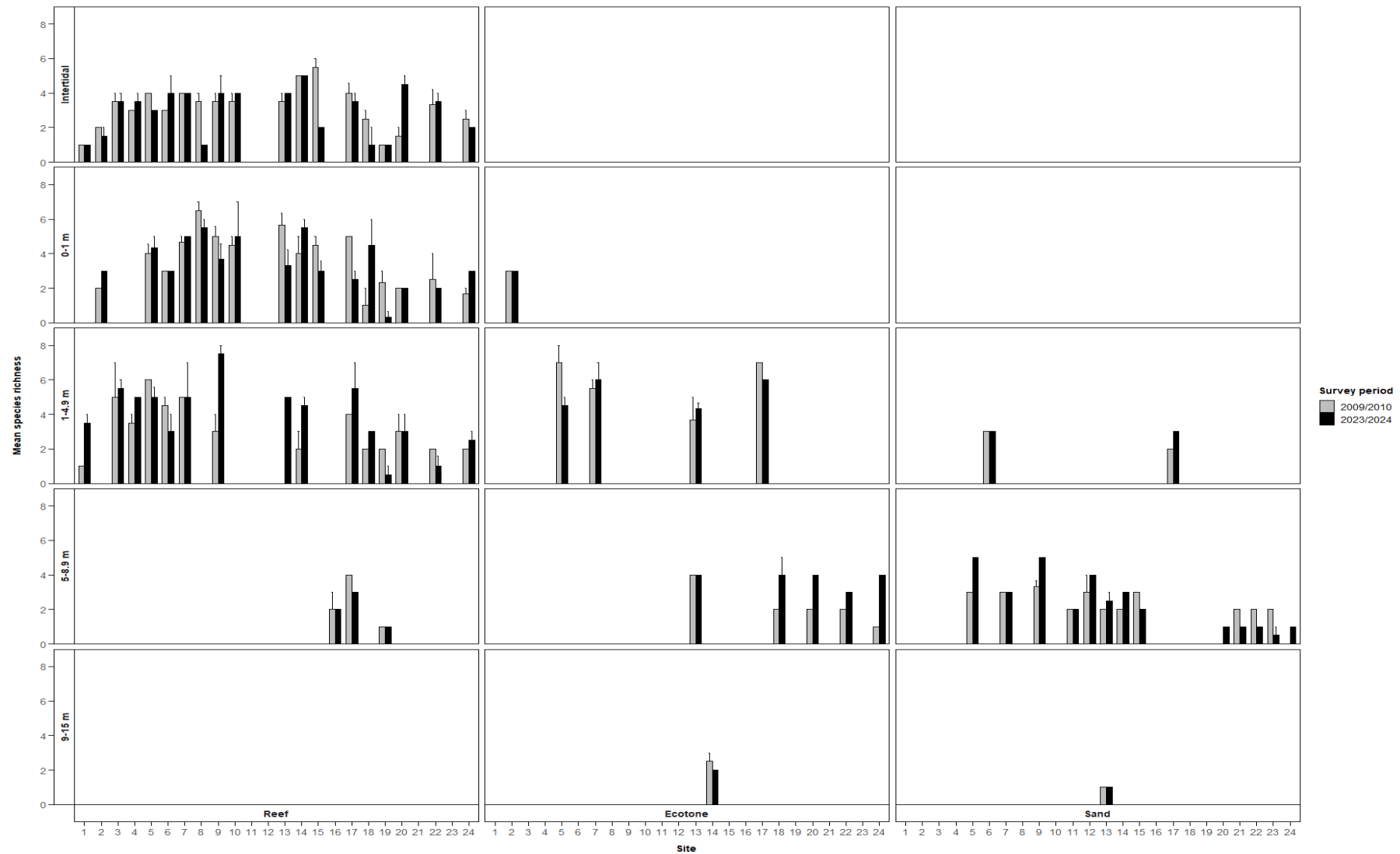


Figure 30: Species richness (mean $\pm$ SE) of introduced species recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

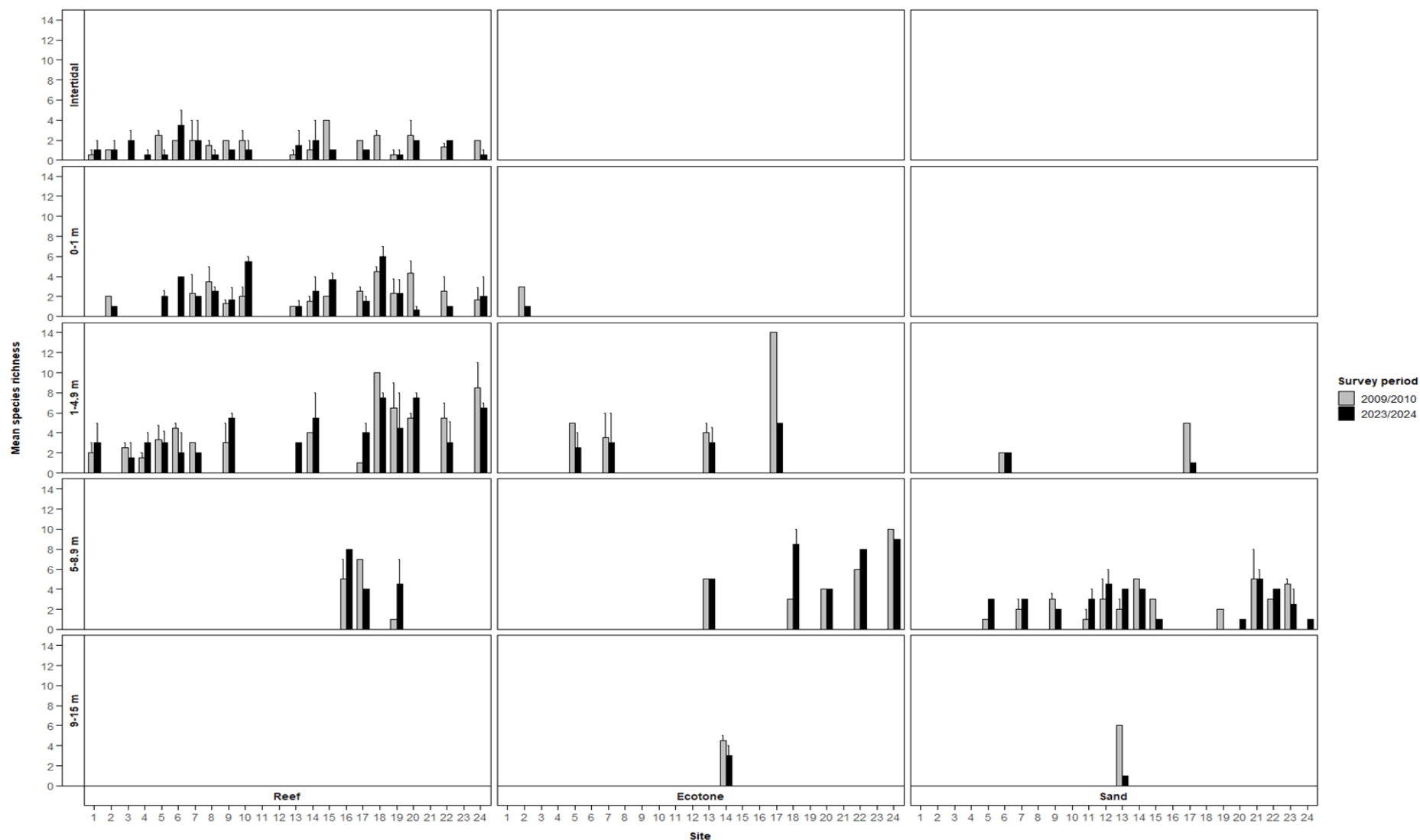


Figure 31: Species richness (mean $\pm$ SE) of the least abundant species (only 1 individual found) recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

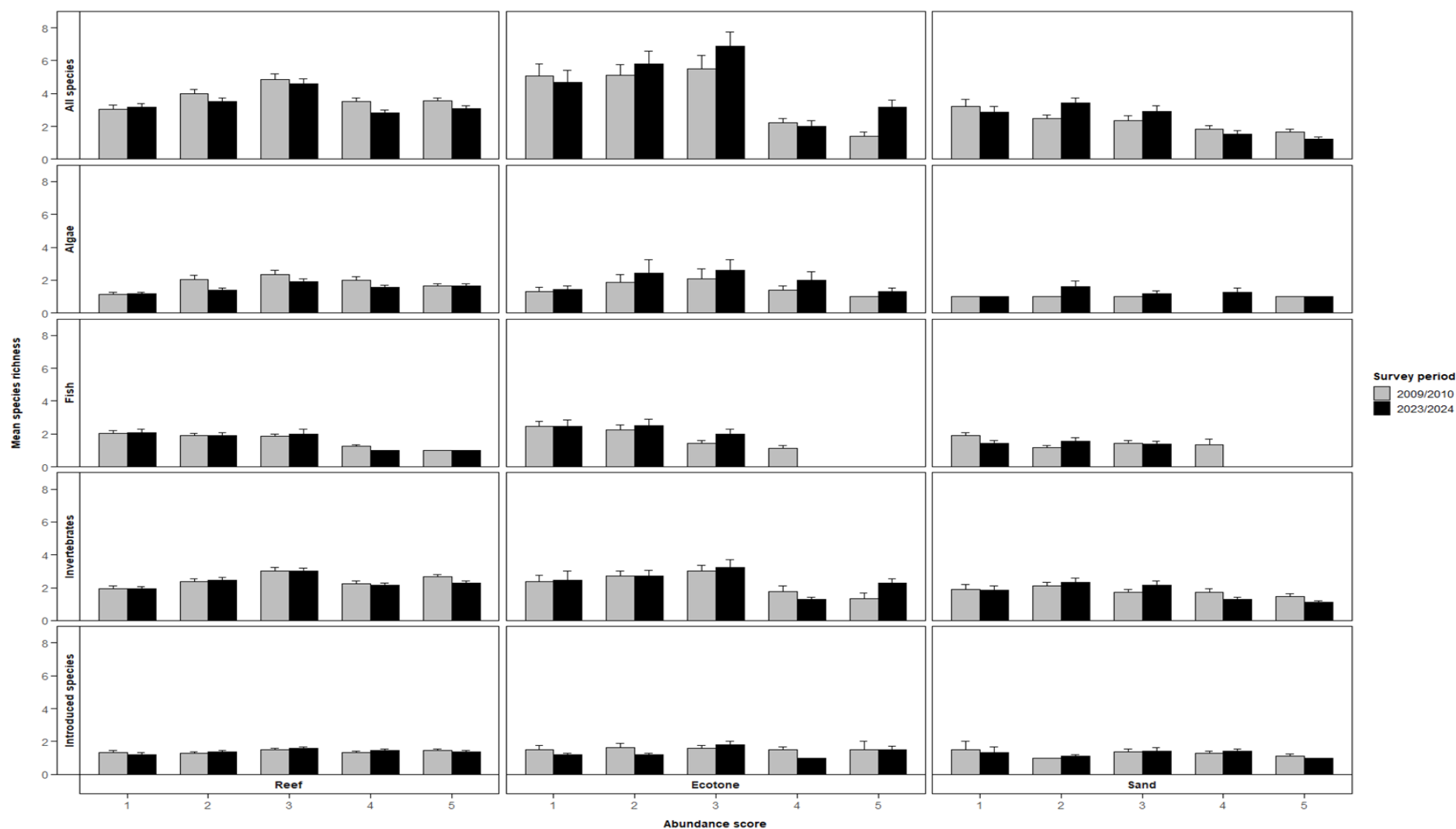


Figure 32: Species richness (mean $\pm$ SE) across abundance scores (1-5) for all taxa, recorded during timed swims in Reef, Ecotone, and Sand habitats of the Derwent in 2010 (grey) and 2024 (black). Abundance scores represent estimated individuals per species per swim: 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = >1000.

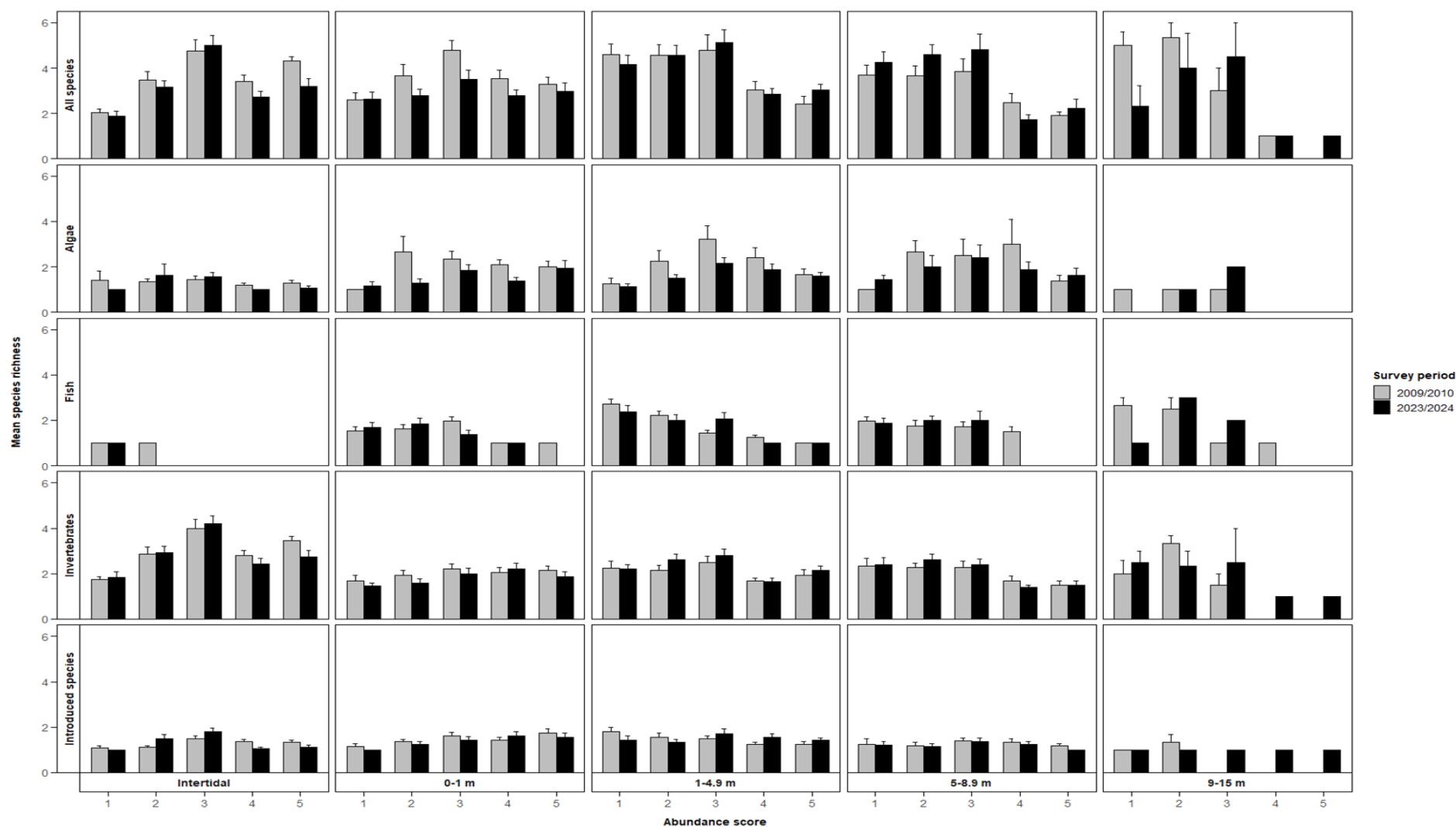


Figure 33: Species richness (mean $\pm$ SE) across abundance scores (1-5) for all taxa, recorded during timed swims in different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m) of the Derwent in 2010 (grey) and 2024 (black). Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+.

Fish

Sixty fish species were counted during the timed surveys, 12 less than 2010. Species richness in reef habitats largely mirrored the spatial patterns observed in 2010 (Figure 34). At 0-1 m depth, richness peaked in the mid-estuary and declined toward both the upper and lower estuary sites. At 1-4.9 m depth, richness generally increased from upper to lower estuary, although some variability was evident in the lower estuary sites. For instance, site 22 had less than half the species richness in 2024 compared to 2010. The majority of fish observations over sand habitats occurred at depths of 5-8.9 m, following a similar spatial distribution to 2010. Most observations were clustered in the mid-estuary, with a species richness comparable to 2010. However, a notable decline was observed in the lower estuary, where often only one species was recorded in 2024, despite previous records of five or more species at those sites.

Five fish species were overlapping as the most abundant in 2010 and 2024 timed surveys (Figure 35). *Forsterygion varium* was the most abundant fish species in reef and ecotone habitats in 2024 and was found at most sites with no clear spatial patterns determined (Figure 35b). Although *F. varium* was recorded at more reef sites overall, its highest rank abundance (rank = 3) was most frequently observed in ecotone habitats, spanning the upper, mid, and lower estuary. This is in comparison to *Trachinops caudimaculatus*, which was previously the most abundant fish in the 2010 timed surveys (Figure 35e). This species declined considerably in 2024, becoming absent across several sites where it was once observed. *Neodax balteatus* exhibited a range shift into the upper estuary reef habitats, with rank abundance of 1-2 at several sites between Bedlam Walls and Naval Cadets, which had no records in 2010 (Figure 35c). Likewise, new and increased observations occurred over sand, with up to 100 individuals (rank = 3) recorded at both sites 9 and 12 (Rosny Point North and South). *Forsterygion gymnotum* similarly shifted further north in 2024, with most observations occurring on reef habitats between Bedlam Walls and Rosny Point in the upper and mid-estuary, and two recordings on ecotone habitats (Figure 35a). Abundance of *Notolabrus tetricus* varied minimally, with the species present across all three habitats (Figure 35d). Its highest rank abundance (=3) was recorded at several lower estuary sites on ecotone habitat. Notable declines occurred at site 16 (Tranmere Reef) over reef habitat, with previous observations of 101-1000 individuals (rank = 4) reducing to only 2-10 individuals (rank = 2).

Parablennius tasmanianus, previously among the most abundant fish in 2010, showed a marked decline in 2024 (Figure 36a). Several sites where it was recorded in 2010 were absent of the species during the most recent survey. In 2024, most observations were clustered between Bedlam Walls and Rosny Point, with a notable absence from other mid and lower estuary sites. In contrast, *Pempheris multiradiata* showed an extended distribution in 2024, occurring across the entire estuary in reef and ecotone habitats (Figure 36b). The species was recorded in higher relative abundance compared to the previous survey, when it was largely confined to the area between Tranmere Reef and Alum Cliffs.

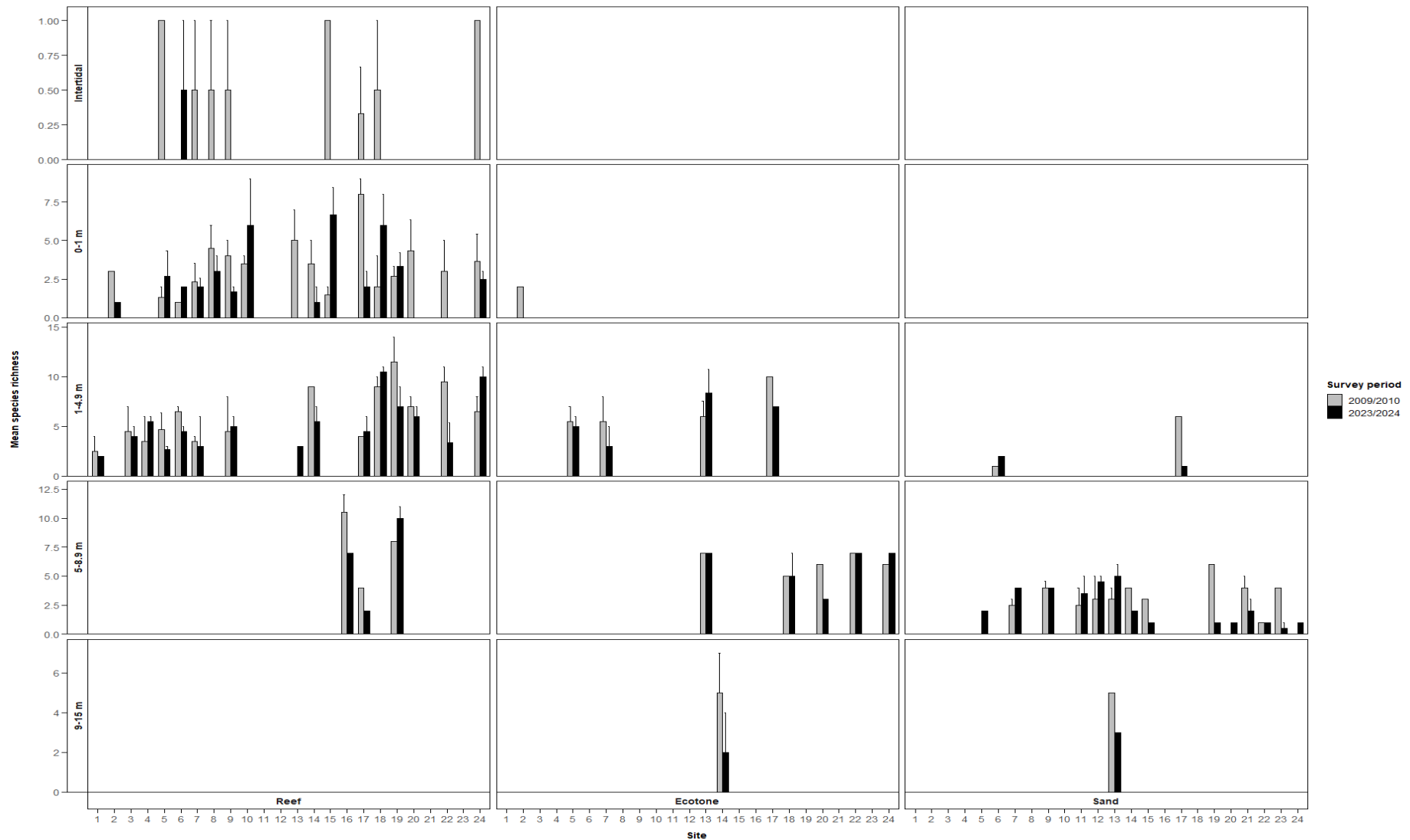


Figure 34: Species richness (mean $\pm$ SE) of fish species recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

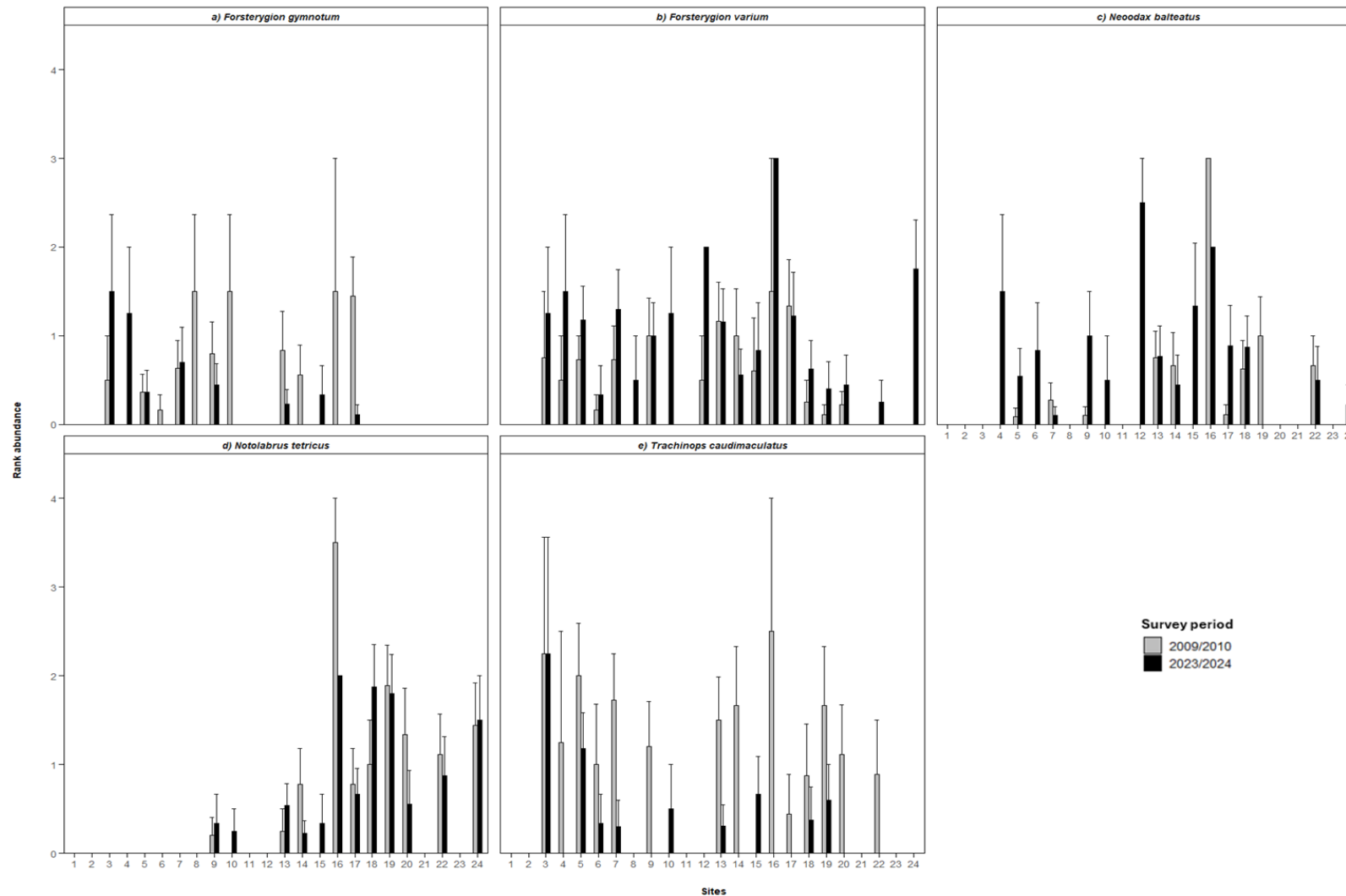


Figure 35: Rank abundance (mean $\pm$ SE) of the most abundant fish species in Derwent in 2010 (grey) and 2024 (black). The five species shown were among the top 6 most abundant in both survey periods, and are presented alphabetically. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 1).

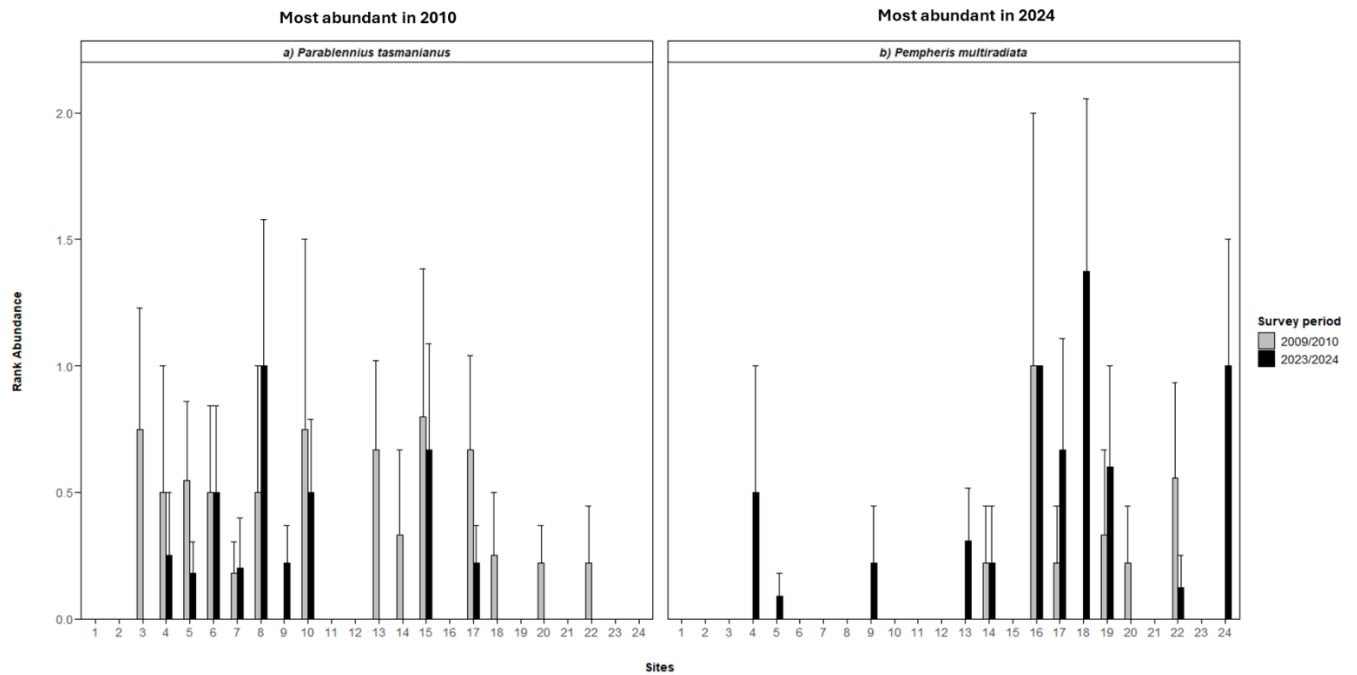


Figure 36: Rank abundance (mean $\pm$ SE) of fish species uniquely among the top 6 most abundant in 2010 (a) and 2024 (b) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill. Species are presented alphabetically. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 2).

Macro-invertebrates

One hundred and forty species of macro-invertebrates were recorded in 2024. As observed in the standard surveys, invertebrate species richness increased from north to south, particularly in intertidal habitats (Figure 37). A similar but more even distribution was evident at 0-1 m and 1-4.9 m depths, although slight declines were recorded at several lower estuary sites – specifically, sites 20, 22 and 24 at 0-1 m, and site 22 at 1-4.9 m. Most observations at 5-8.9 m depth occurred over sand habitats, with minimal variation from the 2010 surveys. Overall, no major changes in invertebrate species richness were detected between survey periods, with only minor site-level fluctuations occurring across the estuary.

We assessed the top eight abundant invertebrate species across both survey periods and how they changed through time. Several species remained among the most abundant species (Figure 38). *Asterias amurensis* increased in rank abundance at most sites, occurring at several new sites across all habitat types (Figure 38a). Previously, the highest abundance was predominantly observed in sand habitats, which remained consistent in 2024. However, rank abundance over both reef and ecotone habitats increased by several orders of magnitude in 2024 as well. In contrast, *Crassostrea gigas* declined by more than ten-fold at nearly all sites and was absent from all ecotone habitats in 2024 (Figure 38b). *Mytilus galloprovincialis* had relatively similar abundance across survey periods, although their spatial distribution extended south, particularly for ecotone habitats, with three lower estuary sites recording high

relative abundance (=5; Figure 38c). Likewise, *Petrolisthes elongatus* maintained a similar spatial distribution and abundance patterns to 2010, although a slight decline in relative abundance is evident across most sites, particularly within the mid to lower estuary (Figure 38f). For example, Collegiate Rowing Club (site 8) declined by two orders of magnitude, as did Battery Point (site 15). *Ostrea angasi* had considerable variability, with no clear spatial patterns (Figure 38d) although it was typically more abundant on deeper reef and ecotone transects, such as the Tranmere reef 5-10 m transect where it was particularly abundant. Over the past decade it appears that several sites in the upper estuary may have increased in relative abundance, although at levels that are not readily differentiated from background spatial variation. In 2024, *Patiriella regularis* appeared to decline on reef habitats throughout the mid-estuary but was newly recorded at two reef sites in the upper estuary (Figure 38e). A few upper and mid-estuary sites also showed increases in abundance of the species on ecotone habitats; particularly sites 2 and 7 (Dowsings Point and 200 m North of the Tasman Bridge), both of which increased multiple orders of magnitude to > 1000 individuals (rank = 5).

Myxicola infundibulum and *Unidentified barnacles* were among the most abundant invertebrates in 2010 but not 2024 (Figure 39a and 39c). Previously, *M. infundibulum* had high relative abundance (=5) over sand habitats, while *Unidentified barnacles* were recorded exclusively on reef habitats, particularly in the upper and mid-estuary. In 2024, both species showed a marked decline in occurrence and abundance, with no observations of *Unidentified barnacles*. This may be due to more successful species identification in 2024. In contrast, both *Heliocidaris erythrogramma* and *Elminius modestus* showed marked increases in rank abundance across the entire estuary, emerging as among the most abundant species in 2024 (Figure 39b and 39d).

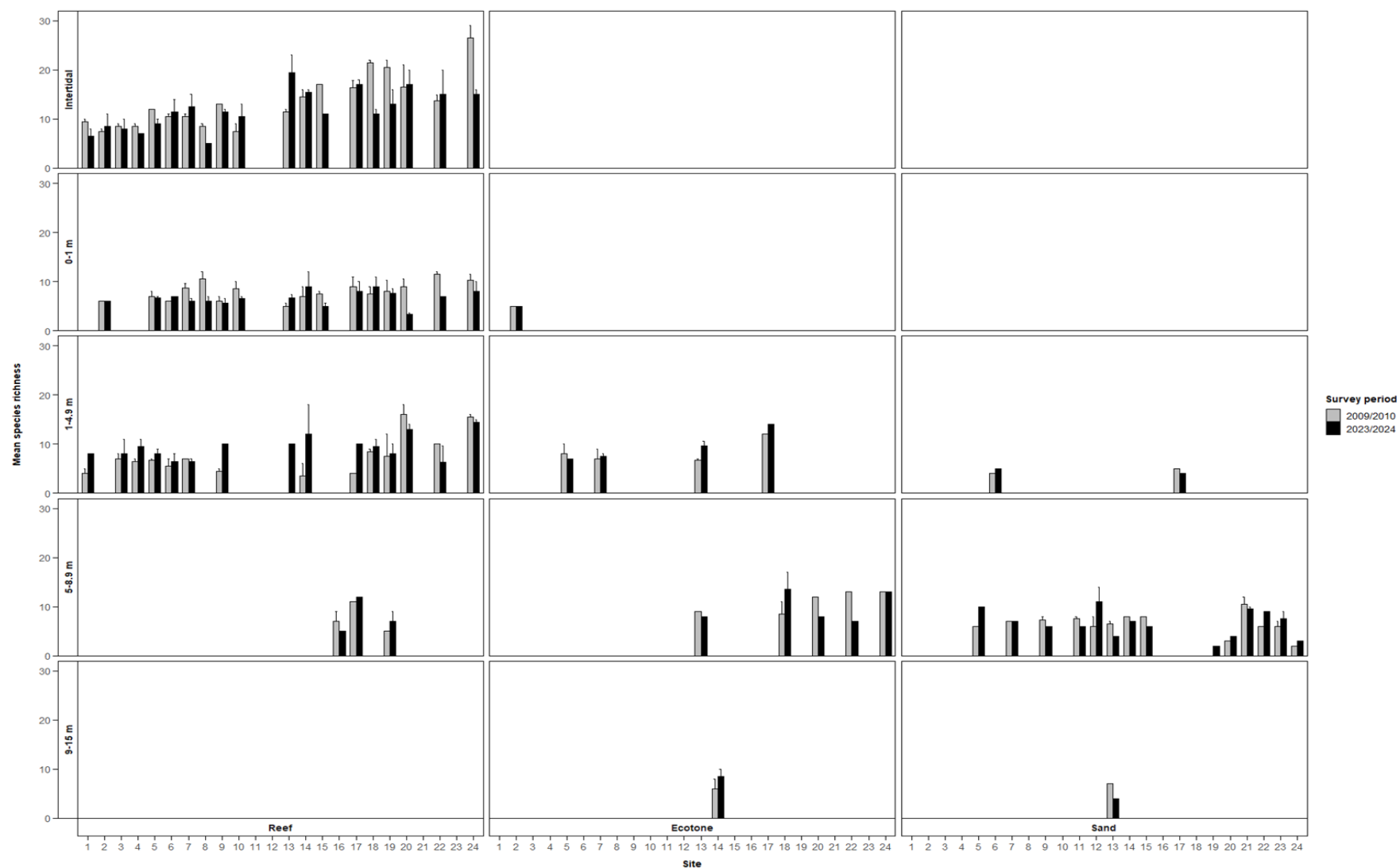


Figure 37: Species richness (mean $\pm$ SE) of invertebrate species recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

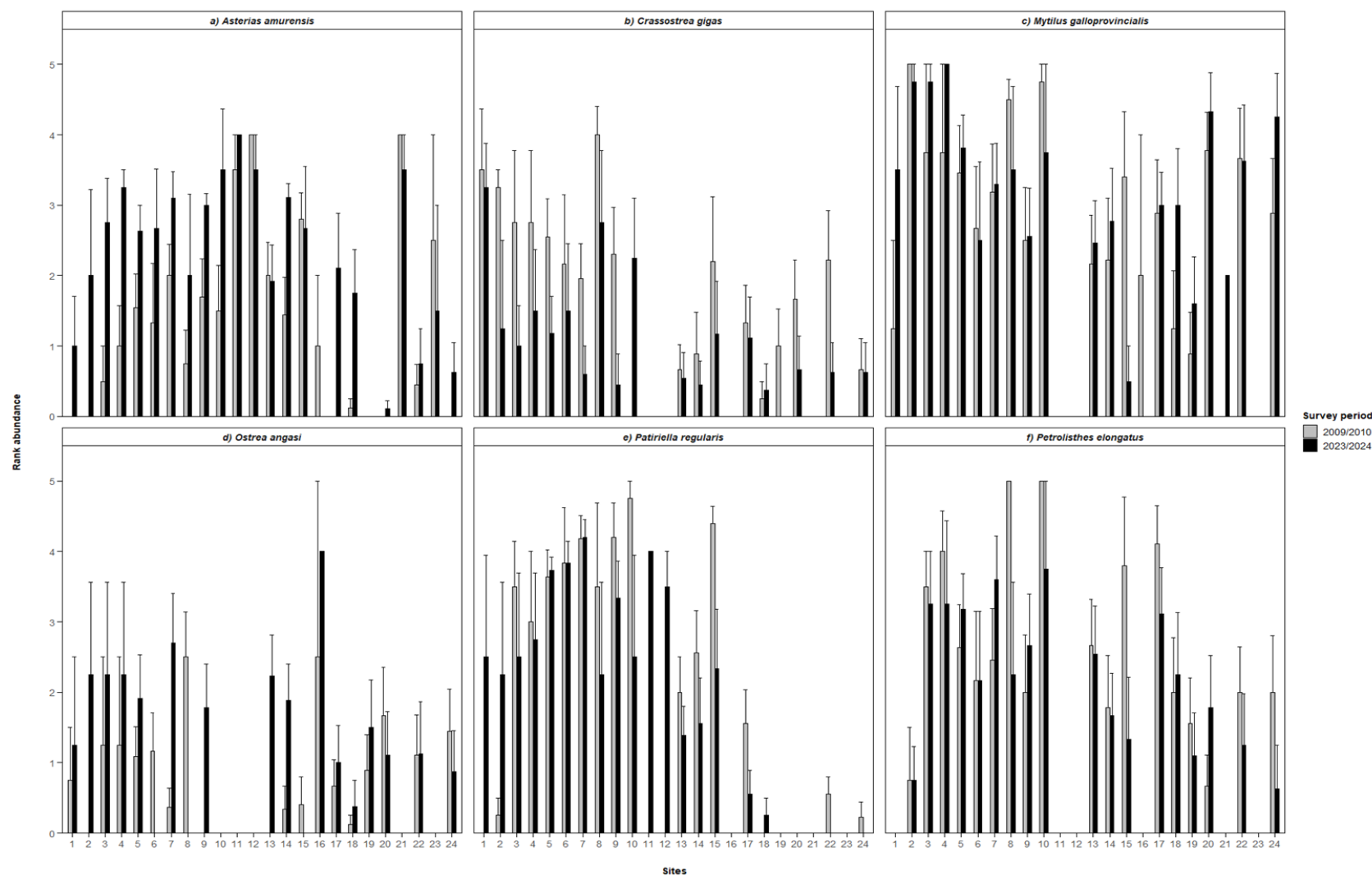


Figure 38: Rank abundance (mean $\pm$ SE) of the most abundant invertebrate species in Derwent in 2010 (grey) and 2024 (black). The six species shown were among the top 8 most abundant in both survey periods, and are presented alphabetically. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 3).

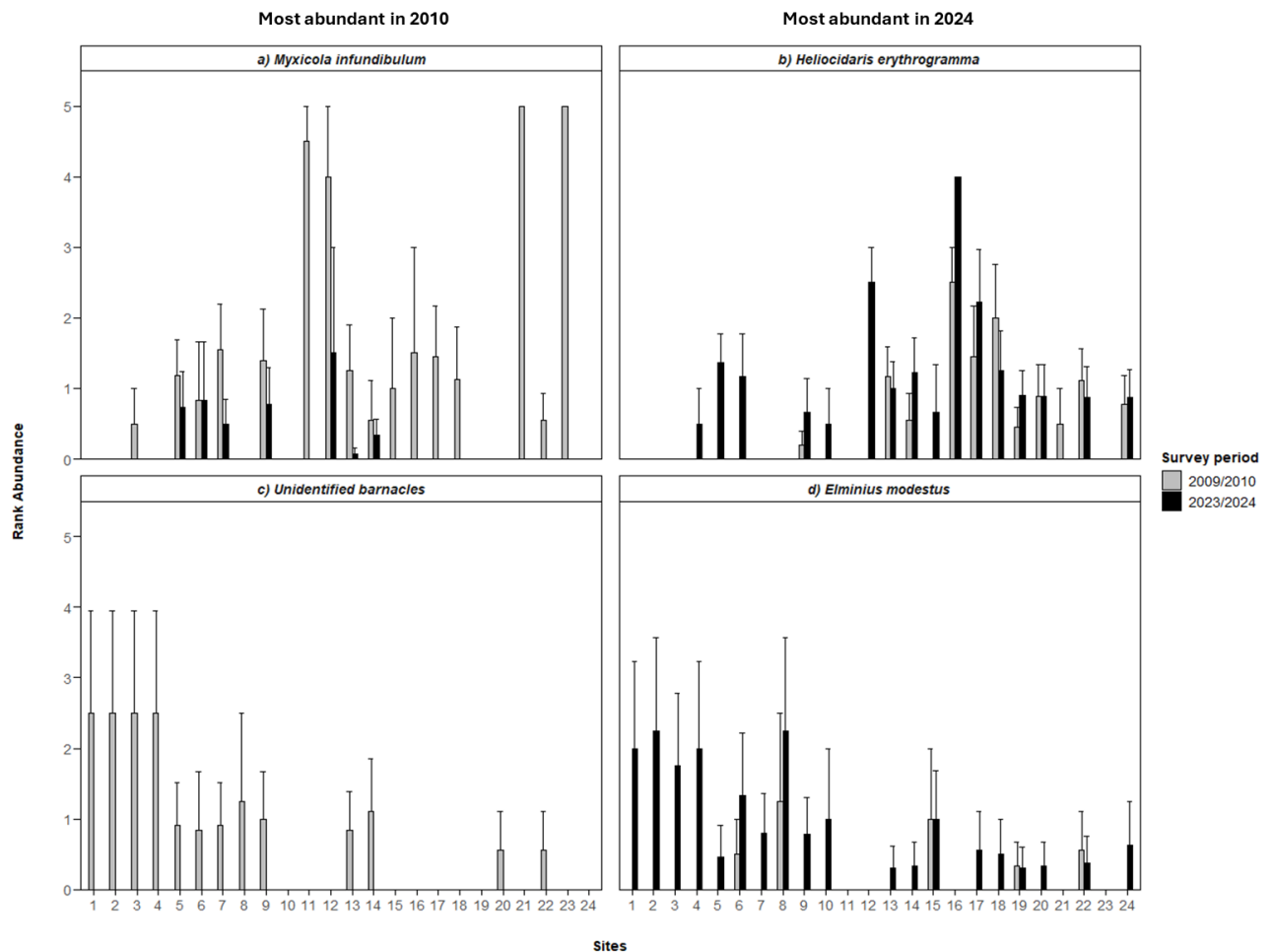


Figure 39: Rank abundance (mean $\pm$ SE) of invertebrate species uniquely among the top 8 most abundant in 2010 (a, c) and 2024 (b, d) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 4).

Algae

Forty-one species of algae and plants were counted during the timed survey. Similar to the standard surveys, algal species richness increased from upper to lower estuary – a spatial pattern observed in 2024 across multiple depths for reef habitats; intertidal, 0-1 m, and 1-4.9 m (**Error! Reference source not found.**). Interestingly, species richness in the upper estuary declined in 2024, however this was primarily within the intertidal habitats, and does not represent the same habitat sampled during the standard surveys. Algal observations at 5-8.9 m depths were spread across all three habitat types, with the highest richness recorded in ecotone habitats located in the lower estuary.

Analysis of algal assemblages across the estuary revealed that, among the eight most abundant species, six were consistently among the most abundant in both 2010 and 2024 (Figure 41). In 2024, *Cystophora moniliformis*, *Ecklonia radiata*, *Lessonia corrugata* and

Undaria pinnatifida were dominant in reef habitats of the mid and lower estuary (Figure 41b, 41c, 41d, and 41f). *C. moniliformis* declined in relative abundance on reef habitats at some sites but expanded its spatial distribution further north. In contrast, *E. radiata* increased in abundance at all sites. Both species also exhibited increased occurrence and relative abundance within ecotone habitats, particularly in the lower estuary, where *E. radiata* exceeded 1,000 individuals (rank = 5) at several sites. *U. pinnatifida*, an introduced brown seaweed, expanded into additional sites in the lower estuary.

Comparatively, in 2024 *Codium* spp. and *Ulva* spp. (Figure 41a and 41e) remained relatively evenly distributed across the estuary, though overall abundance declined compared to 2010. Abundance peaked slightly in the mid-estuary for *Ulva* spp., in both reef and ecotone habitats, but notable site-level variability was evident. For example, sites 11 and 12 in the mid-estuary recorded no observations of *Ulva* spp. in either survey, while nearby sites 16, 20 and 24 all peaked with rank abundances of 3-4 in 2024. *Codium* spp. was recorded in 2024 further upstream, at sites 4-7, with increased rank abundance, but declined at several other sites in the mid-estuary.

The algal community shifted somewhat in 2024, with species once abundant in 2010 no longer among the top abundant species in 2024. *Dictyopteris muelleri* declined significantly in both abundance and distribution, with its range shifting further north (Figure 42a), while unidentified filamentous brown algae were not recorded at all in 2024 (Figure 42c). Comparatively, the introduced species *Grataloupia turuturu* showed a substantial increase in abundance in 2024, expanding its range into the mid-estuary for reef habitats, and into the lower estuary for ecotone habitats (Figure 42b). A similar spatial distribution was observed for the native *Sargassum fallax*, predominantly present in the mid and lower estuary (Figure 42d). Both species were recorded at several new sites in 2024, indicating range expansions and overall increases in abundance. Site-specific variation was evident, with certain sites showing much reduced abundance for *S. fallax* (e.g., sites 16, 18 and 19) and others increasing substantially in 2024 (e.g., sites 14, 17, 20).

We assessed the relative abundance of *Ecklonia radiata* as an indicator of water clarity across different depth zones and observed both increased abundance and an expanded depth range in 2024, particularly within 0-1 m, 1-4.9 m and 5-8.9 m subtidal zones (Figure 43). Notably, sites 13 and 14 (Bellerive Bluff and Bellerive/Howrah Bluff) exhibited substantial increases, rising from sparse coverage (rank 1-2; ≤ 10 individuals), to near-complete coverage (rank 4-5; 101-1000+ individuals). Similarly, site 9 (Rosny Point North) showed a marked increase in the 1-4.9 m depth, where *E. radiata* was previously absent but reached a rank abundance of 4-5 in 2024. Sites 17 and 18 (Tranmere Point and The Grange, Taroona) showed comparable trends. At site 17, abundance in the 1-4.9 m zone increased dramatically, from a single individual (rank 1) to an estimated 101-1000+ individuals (rank 4-5). Both sites also recorded new occurrences at 5-8.9 m depth, with moderate rank abundances (ranks 2-3), indicating a deepening distribution.

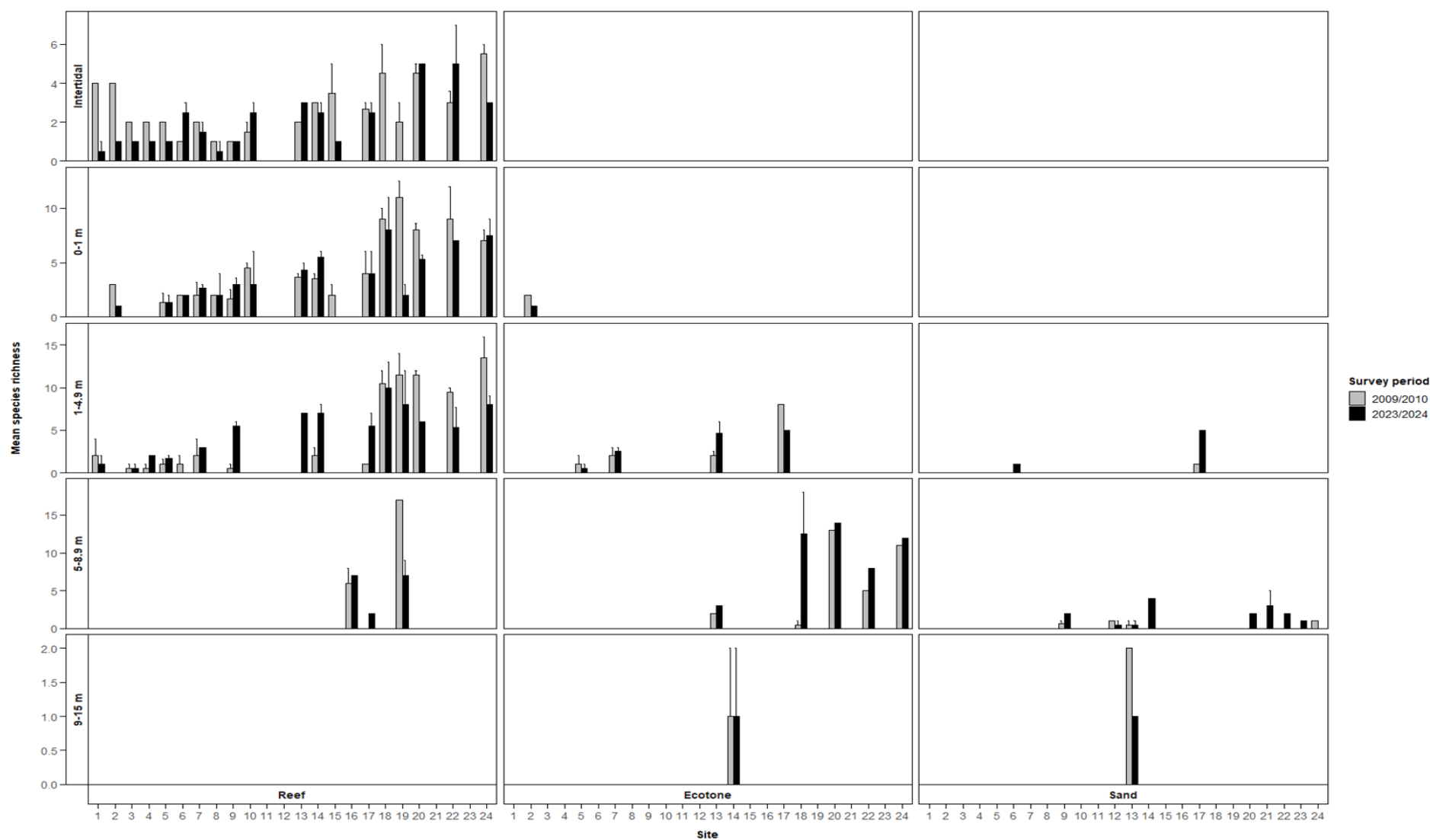


Figure 40: Species richness (mean $\pm$ SE) of algal species recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black), in different habitats (Reef, Ecotone, Sand) and different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m).

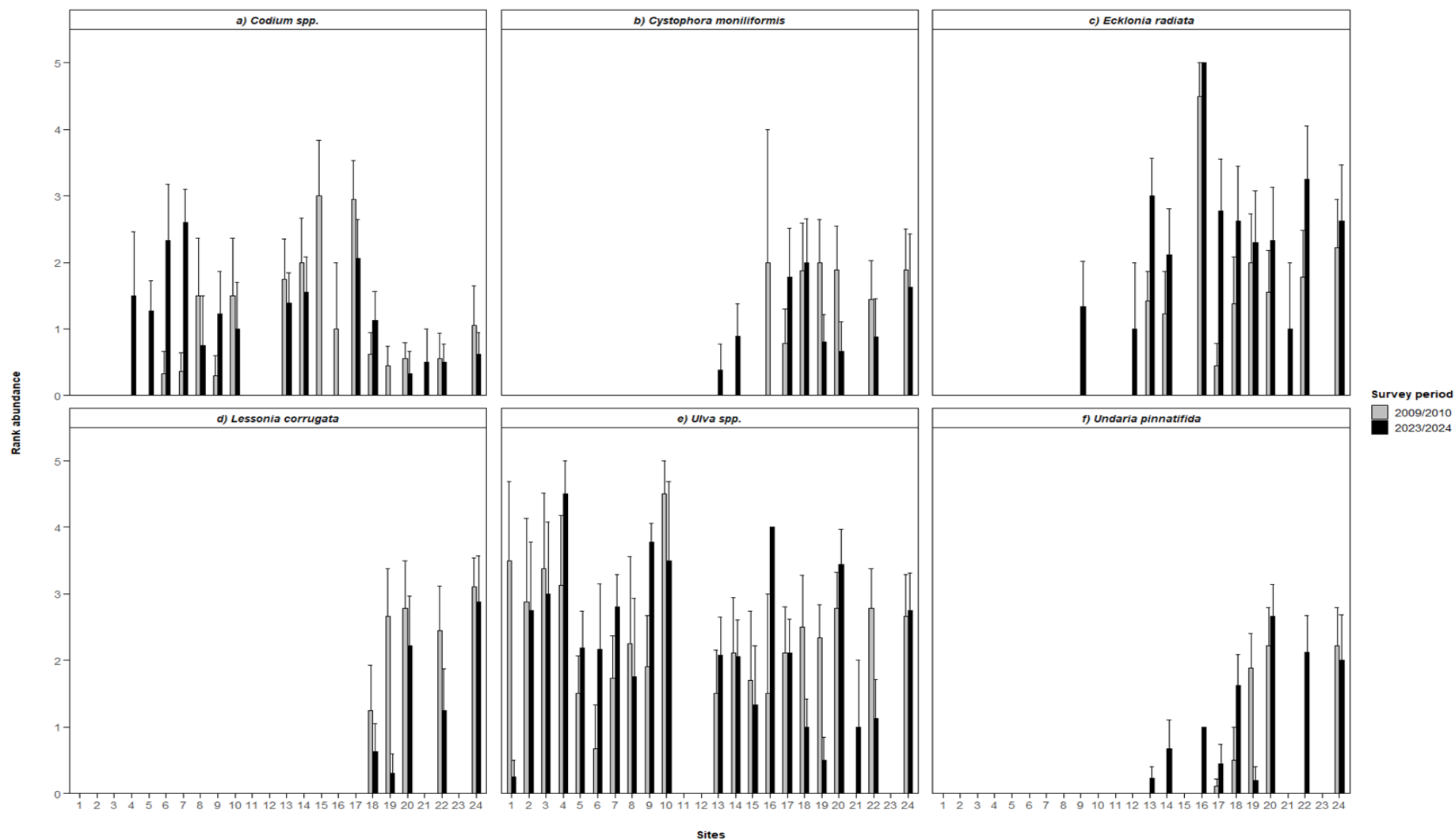


Figure 41: Rank abundance (mean $\pm$ SE) of the most abundant algal species in Derwent in 2010 (grey) and 2024 (black). The six species shown were among the top 8 most abundant in both survey periods, and are presented alphabetically. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 5)

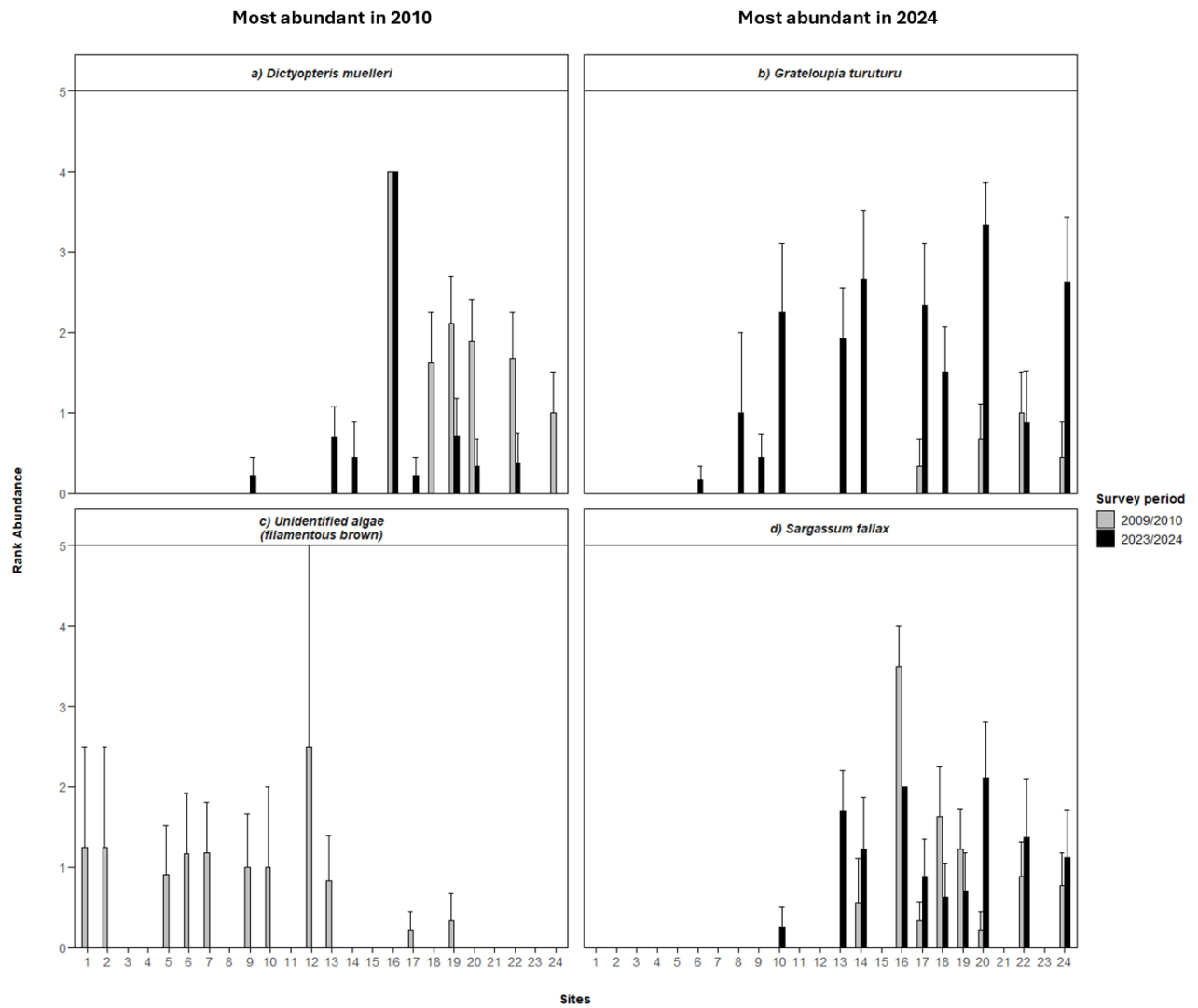


Figure 42: Rank abundance (mean $\pm$ SE) of algal species uniquely among the top 8 most abundant in 2010 (a, c) and 2024 (b, d) in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill. Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Habitat-specific patterns are provided in the Appendix (Appendix 6).

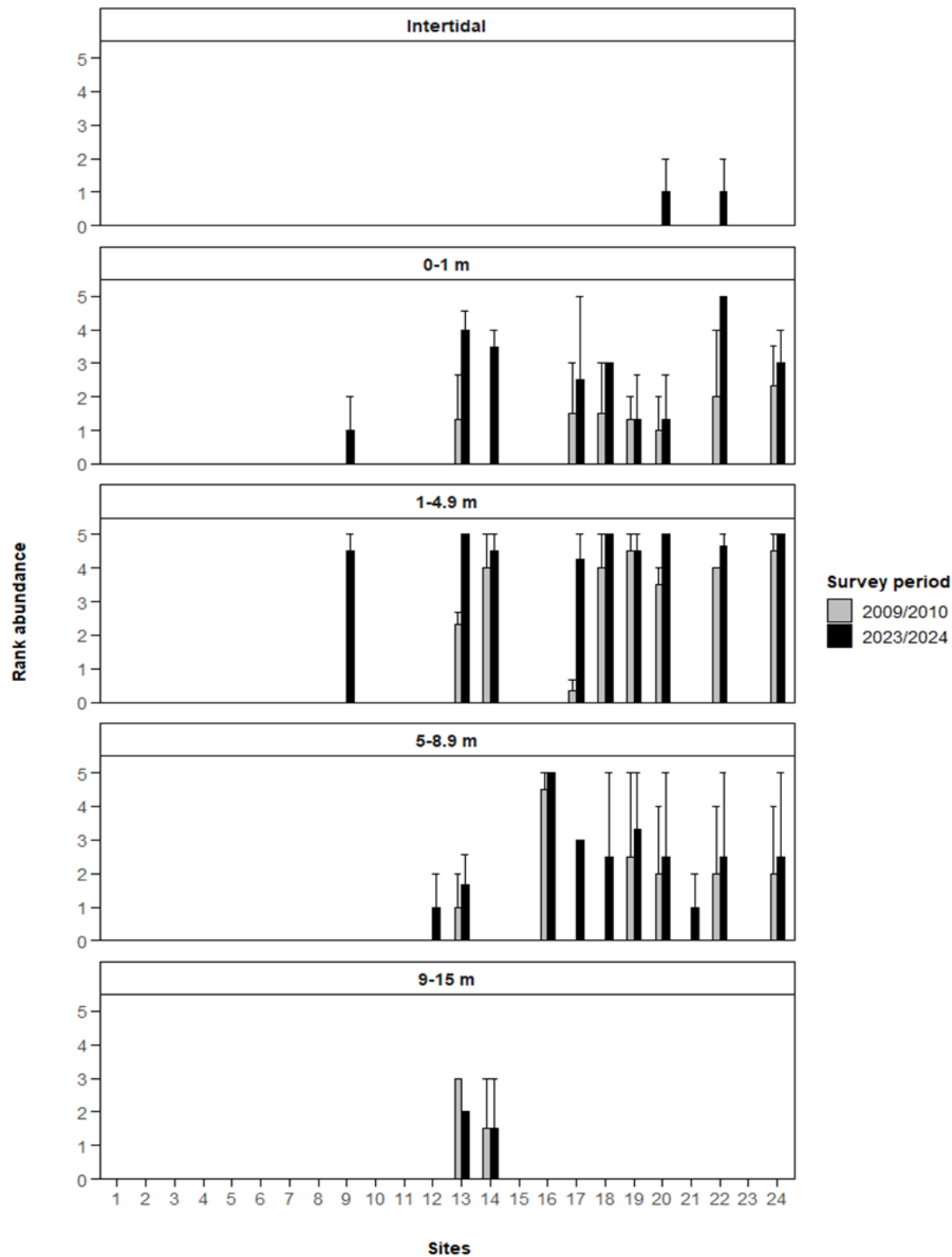


Figure 43: Rank abundance (mean $\pm$ SE) of *Ecklonia radiata* recorded during timed swims in the Derwent in 2010 (grey) and 2024 (black) at different depths (Intertidal, 0-1 m, 1-4.9 m, 5-8.9 m, 9-15 m). Abundance score; 1 = only 1 individual, 2 = 2-10, 3 = 11-100, 4 = 101-1000, 5 = 1000+. Note site 17 (Tranmere reef) is offshore, with no reef shallower than 5 m.

3.3 Targeted *Ostrea angasi* survey

The results are shown in Appendix 20 where most collected oysters were able to be positively identified to species by internal features. The sub-sample of these specimens were then further validated using a taqman qPCR assay developed by Coutts et al. (2022) for positive identification of *O. angasi*, results of this molecular analysis is found in Appendix 22.

Overall, these results show a strong correlation with the species site and depth distributions recorded in the earlier surveys, validating the general patterns observed. This shows *C. gigas* dominating the intertidal zone where *O. angasi* was rarely present. This continued through the 0-1 m depth zone before a transition through the 1-5 m depth zone, with *O. angasi* then becoming the dominant species in the 5-10 m zone, although *C. gigas* were still occasionally present in this zone. At times, some collected specimens were not able to be identified to species based on external features alone. Likewise, some intermediate specimens were not even able to be identified using the key internal morphologies (inner marginal teeth and hinge striations, Appendix 21) as these were missing.

For the genetic testing with the qPCR assay, eight of the twelve specimens initially identified as *O. angasi* tested positive, one tested negative suggesting an incorrect morphological ID, while the remaining three were intermediate as they had only 1 of 3 replicates return a positive result. Of the sixteen specimens of *C. gigas* tested, fourteen were negative, while one returned positive results as *O. angasi* and another came back as an intermediate result. The intermediate results were likely caused by the presence of non-DNA contamination preventing successful PCR and cannot be confidently assigned positive or negative. These positive/intermediate results in the *C. gigas* group had previously been flagged as somewhat intermediate in form and lacking clear internal features. These were initially classified as likely *C. gigas* based on external shell morphology but were prioritised for genetic testing based on the lack of clear differentiating features.

Of note, is that no specific *O. angasi* reefs were found during the study, although the rocky reef at Tranmere reef did have extensive cover between 5-10 m. An extensive oyster reef was found at Cadbury Point in 2-3 m depth, however, all samples collected there were found to be *C. gigas*.

4. DISCUSSION

This study replicates a series of baseline surveys on rocky reefs and adjacent sediments in the Derwent Estuary completed in 2010, with the aims of identifying biodiversity changes through time. This survey revealed several distinct trends in the fish, invertebrate and algal communities in both space and time, building upon the 2010 baseline and establishing a second time point from which to continue decadal monitoring of key ecological components of the system.

4.1 Standard fixed-depth transect-based subtidal reef surveys

Fishes:

Overall, the spatial distribution of fish communities remained broadly consistent through time, with fish species richness increasing from upper to lower estuary, with the greatest richness in the open, marine-influenced sites – as expected for an estuarine system and evidenced in the baseline survey (Barrett et al. 2012). However, overall richness declined for

fish communities through time. Of particular note is the decreased richness from Tranmere Point to Blackmans Bay Blowhole (site 17 to 24). Previously a species-rich area of the estuary, this region has declined considerably since 2010 and now matches much of the mid and upper estuary. Generally, the riverine extreme of an estuary produces a more depauperate environment, with reduced species richness and abundance, due to a lower saline environment (Edgar et al. 1999; Neves et al. 2011). This was evident in the historic baseline surveys of 2010 within the Derwent Estuary. However, the current survey suggests a community restructuring, with compositional changes and a general trend of community simplification; with several sites supporting simpler, less-diverse assemblages. Tranmere Point, Alum Cliffs and White Rock, South Arm (sites 17, 20, and 22) – all lower estuary sites – now more closely resemble the richness of the mid-estuary site, Rosny Point North, despite their marine-influenced locations. Localised changes within the estuary are evident though, with species richness increasing from 2010 for several upper and mid-estuary sites, from Bedlam Walls to Rosny Point. This could be a result of localised resilience to the habitat, or habitat improvement, which correlates with the increased algal cover seen at these sites. However, the likely overall cause of the observed decline in richness at the marine end of the system will require further time-series surveys to be undertaken to determine if this is an ongoing real trend, or an artefact of the generally poor visibility encountered during the 2024 survey following a major storm in late February.

Estuary-wide changes in salinity can also result in community restructuring, with increasing salinity during periods of low river flow or frequency of flooding events driving an increase in the richness and abundance of marine and estuarine species as they extend their range upstream (Baptista et al. 2015; Lauchlan & Nagelkerken 2020). An extended period of increased salinity leading up to the 2024 survey may explain the observed increase in marine and estuarine species observed at locations such as Cadbury Point and Bedlam Walls North during the 2024 surveys.

The riverine extremes of estuaries are typically vulnerable to flood events, which are often exacerbated by the urban environments that commonly surround estuaries (Harrison et al. 2022). Flooding can significantly alter an estuary by increasing water flow, sedimentation and nutrient concentrations, and changing the habitats and species within the estuary (Cooper 2002; Rogers & Woodroffe 2016; Henderson et al. 2024). In severe cases floods can result in the complete loss of ecosystems (Davis et al. 2022). No major flooding events have occurred in the Derwent Estuary from June 2016 to June 2024; in 2016 the river exceeded the major flood level (7.3 m), causing widespread damage and loss of life (Bureau of Meteorology 2016). With almost eight years separating the last event from these biodiversity surveys, this reprieve may have allowed the upper estuary to stabilise, with high salinity levels resulting in the increased richness observed at these sites since 2010. Notably, the mid to upper estuary is regularly influenced by rainfall events and periods of lower salinity in surface waters that particularly influence intertidal to immediate subtidal species. However, at moderate depths (5 m) more significant flood events are required to displace many marine species and provide species restructuring over multiple year scales.

At an individual species level, we observed marked declines in several formerly common species since the 2010 baseline, including *Trachinops caudimaculatus* and *Latridopsis forsteri*. *Trachinops caudimaculatus* is a resident reef schooling planktivorous species, and is particularly vulnerable to changes in the plankton communities, with this driving changes in

abundance on the open coast communities (Barrett et al. 2007; Bates et al. 2014). Both flooding and eutrophication can disrupt the physical and chemical environment, negatively affecting zooplankton, and by extension, the species that depend on them (Rose et al. 2019; Wang et al. 2021). Ongoing variation in eutrophication in the Derwent (Derwent Estuary Program 2024) resulting from storm events and human actions could be influencing zooplankton availability, resulting in the overall decline observed between 2010 and 2024.

For *Latridopsis forsteri*, a recreationally and commercially targeted species, the potential drivers of the observed decline are possibly more complex, but important to try to understand. This is a resident reef schooling species, typically with variable year to year abundance due in part to its episodic recruitment, which happens in large bursts at infrequent intervals (Edgar 2008). Despite the variability, populations readily build up in no-take marine reserves such as those found at Tinderbox and Maria Island where they remain until reaching sexual maturity at 6-7 years of age (Barrett et al. 2007). Elsewhere, they have declined markedly relative to historical levels, most likely due to their high susceptibility to both recreational and commercial fishing via use of mesh nets (Harries & Croome 1989).

During the 1990s and early 2000s, extensive netting restrictions implemented by the Tasmanian government effectively created a de facto reserve for this species within the Derwent Estuary. These restrictions, along with the species' low susceptibility to capture by handlines, likely contributed to the high numbers of trumpeter observed during the 2010 surveys. However, by the 2024 survey, numbers had declined dramatically. To examine whether this decline reflects natural variation in recruitment or human influence, we contrasted these results with unpublished monitoring data from nearby protected sites within the Tinderbox Marine Reserve. There, numbers declined from a total of 52 to 38 individuals sighted on transects between 2010 and 2024, a much smaller decline than observed in the Derwent Estuary sites. This suggests that poor visibility encountered during the 2024 surveys may have inhibited divers ability to observe *L. forsteri*.

However, recreational fishing remains a major pressure in the Derwent, and increased fishing activity cannot be ruled out as a driver of the decline observed in 2024. Minimum size limits for both commercial and recreational fishing remain well below the estimated size at maturity (Murphy & Lyle 1999), likely hindering successful recruitment. Anecdotally, recreational spearfishing in the Derwent has markedly increased over this period, potentially prompted by the high numbers of trumpeter observed in 2010. This combination of increased spearfishing pressure and impaired recruitment may be contributing to the sharp decline in Trumpeter numbers observed in 2024, although similar declines have also been observed in the marine reserve at Maria Island over this period (Barrett, unpublished data). Increased protection of Trumpeter in the Derwent Estuary through spatial restrictions on spearfishing activities could be one way of assisting in the recovery of Trumpeter stocks, improve conservation and biodiversity values, as well as enhancing snorkelling activities and improving public safety in areas of high recreational use.

In addition to abundance changes, minor range shifts also occurred for several species. Two key species of interest are *F. varium* and *F. gymnotum*, both introduced triplefins and primarily considered cryptic species. While the standard fish surveys showed little change in their abundance or distribution, cryptic fish surveys revealed a substantial increase in both metrics in 2024 for these species. These results also highlighted clear range expansions for

both species. Range shifts in marine species are often driven by changing environmental conditions, including warming waters and salinity changes (Hastings et al. 2020; Pinsky et al. 2020), as well as competition and resource allocation (Sorte et al. 2010; Lord & Whitlatch 2015). Range expansions within an estuary, particularly into the upper estuary as we've observed, can occur when increased temperatures and salinity result in more available, favourable habitats throughout the estuary (Baptista et al. 2015). As discussed earlier, a period of relative stability in the estuary in recent years, driven by low flood frequency, may explain much of the range and abundance increase seen at present. Whilst neither *F. varium* nor *F. gymnotum* are considered invasive, these introduced species are a conspicuous component of the reef fish fauna in the mid to upper estuary. Changes in the environment may be favouring these cryptic introduced species, which prefer sheltering in complex rocky reefs (Willis & Anderson 2003), with some sites in 2024 increasing significantly in algal coverage.

In contrast to the fish communities sighted on the broad fish transects, the overall richness and abundance of cryptic fish species (sighted on the narrower detained seabed searches) increased substantially since the 2010 baseline. Our results indicate a clear shift northward for species richness, and coupled with an increase in overall abundance, these findings suggest potential environmental changes that support cryptic species, and their preference for increased habitat complexity. This observation includes the native Tasmanian blenny, *Parablennius tasmanianus*, which increased in abundance within the upper estuary. However, this was matched by a reduction in abundance of gobies in the same region.

Mobile Invertebrates:

These same environmental changes may also be supporting the increase in abundance seen for several cryptic macro-invertebrates, including the two introduced asteroids: the northern Pacific seastar (*Asterias amurensis*) and the New Zealand common cushion star (*Patiriella regularis*). The northern Pacific seastar is a highly invasive echinoderm that has been established in the Derwent Estuary since the 1980's, with significant impacts on aquaculture and commercial shellfish industries (Ross et al. 2004; Hutson et al. 2005). Its establishment also coincides with the decline of the Spotted handfish (*Brachionichthys hirsutus*) in the Derwent, likely due to ecological impacts such as predation on preferred spawning substrates (Bruce et al. 1997; Stuart-Smith et al. 2020). With no recent population assessments in Tasmania, the significant increase observed in the 2024 survey within the estuary should be taken into careful consideration for ongoing management of this pest species. Likewise, *P. regularis* is a dominant species within the upper estuary, and is considered a major competitor to the native, endangered Derwent seastar, *Marginaster littoralis* (Beeton 2009; Barrett et al. 2012), a species now thought to be extinct (O'hara et al. 2018). The population of *P. regularis* appears to be thriving in the current estuarine conditions, with significantly increased abundance from Cadbury Point to Battery Point compared to the baseline data. Again, estuaries are naturally dynamic ecosystems, particularly towards the riverine extreme, where species must be adaptable to withstand acute environmental changes. Much of the variation observed in the middle to upper estuary sites where these species increased the most is likely a reflection of this naturally shifting environment, captured during a period of relative stability. In the absence of any major flooding events in recent years, opportunistic species like *A. amurensis* and *P. regularis* can flourish. Given their invasive nature, continued targeted monitoring of both introduced seastars is recommended. Monitoring at similar

spatial coverage and decadal intervals to this repeat study would give sufficient information to understand the trajectory of introduced problem species and their interaction with native species and habitats. While no effective mechanisms have been identified to remediate such invasives at large scales, the information is effective in highlighting the nature of the invasion, possibilities for local area control (e.g. volunteer removals in threatened species hotspots), and the need for ongoing exploration of solutions to the invasive species problem.

This recent lack of major disturbance may also explain increases in the abundance of several common native invertebrate species, including the short-spined sea urchin (*Heliocidaris erythrogramma*). The invasive long-spined sea urchin (*Centrostephanus rodgersii*) is the dominant species responsible for forming urchin barrens, through extensive overgrazing of kelp ecosystems (Ling et al. 2009; Ling & Keane 2024). However the native short-spined sea urchin can also have considerable ecological implications in the environment, particularly in sheltered waters such as the Derwent Estuary. At high densities, *H. erythrogramma* can overgraze algal communities, leading to reduced habitat complexity and biodiversity (Ling et al. 2010). In extreme cases, urchin barrens can form long-term as an alternative stable state for collapsed kelp ecosystems (Scheibling & Filbee-Dexter 2014; Ling et al. 2015). Our results, however, show that while this species increased in abundance, increasing its range into the upper and mid-estuary, it was not found at levels likely to cause urchin barren formation at any sites. Increased urchin abundance has been linked to fishing pressure and the reduction of urchin predators such as the *L. forsterii*. Other positive signs of recovery in the Derwent include the increase in abundance of *Haliotis rubra*, a commercially harvested native abalone, in the lower estuary.

Macroalgae:

The lower estuary continued to provide diverse algal habitats in 2024, with little significant change since 2010. Overall, the wider system maintained the expected trend of species richness declining with distance from the ocean, with algal diversity highest at the marine end of the estuary (Hardwick-Witman & Mathieson 1983; Edgar et al. 1999). One notable change, was an increase in algal cover in the mid to upper estuary sites, particularly for brown and red algae, despite the overall lack of species richness in these regions. In the upper estuarine site at Cadbury Point this was primarily due to a significant cover of unidentified filamentous red algae, while slightly further downstream there was a notable increase in the cover of the red algae *Rhodymenia* spp. Species within this genus (most notably *R. leptophylla*) became the dominant cover at a number of sites that previously lacked algal cover. As with other marine species, a range extension of macroalgae further into the estuary may be from numerous factors, including a more saline environment due to fewer floods, but also more specifically for algae, potentially increased light penetration due to improving water clarity.

In another, and likely related sign of recovery within the estuary, the cover of golden kelp *Ecklonia radiata* has significantly increased across the mid-estuary. The recovery of canopy-forming species like *E. radiata* is a critical first step in the recovery of macroalgal forests (Galobart et al. 2023), and the increased abundance observed in these surveys suggests more favourable environmental conditions within the estuary, including increased light availability at depth. The marked increase in algal cover in the mid-estuary (at sites such as Rosny Point) also appears to be a response to a marked reduction in the cover of a polychaete (tube worm)

– likely a spionid polychaete species – which formed an extensive matting on rocky reefs at many of the mid to upper estuary sites in 2010. During that survey the matting comprised almost 100% cover at transect depths (2-4/5 m) at most sites between Bedlam Walls to Battery Point, effectively trapping silty sediments and inhibiting the settlement of algal gametophytes and development of large sporophytes (Barrett et al. 2012). It was speculated at the time that this may represent an altered state that would be difficult to recover from. However, encouragingly, sites such as Rosny Point and Bellerive Bluff now show strong signs of recovery, with increased coverage of *E. radiata* and *Sargassum fallax* (to a lesser extent) and red algae including *Rhodomenia leptophylla*, *Peyssonnelia* spp., and *Rhodophyta* spp. The apparent reversal of what was once described an alternative stable state dominated by tubeworm matting suggests a shift towards a more biodiverse, healthy ecosystem, potentially driven by a reduction in sediment loading that is thought to facilitate the maintenance of the polychaete mats.

Although no major floods have occurred in recent years, boat traffic throughout the estuary remains consistent, with commercial and recreational vessels operating daily. Eriksson et al. (2004) found that boating activities – and the resulting boat-generated waves – can influence both the vegetation and macrobenthic infauna, although responses are often taxa-specific. Similarly, Bishop & Chapman (2004) observed significant changes in infauna assemblages within Sydney Harbour’s wash zone during a temporary cessation of ferry traffic. During this period, assemblages began to resemble those in no-wash zones, suggesting that the wash disturbance directly affects benthic community structure. Eriksson et al. (2004) also reported a sharper decline in vegetation cover and species richness with depth in marinas or inlets exposed to ferry routes compared to those not subject to boating disturbances. In the Derwent, frequent vessel activity – including regular MONA ferry transits and other commercial traffic – may be contributing to ecological changes observed in the upper estuary, via a marked increase in shoreline wave action. Although boat-generated disturbance is often associated with negative effects like sediment resuspension (regularly observed in the 2024 surveys), in this context, it may have also facilitated the removal of sediments trapped by the tubeworm matting at various sites along the ferry route – potentially aiding habitat recovery within the estuary and contributing to increased algal cover. More detailed studies are required to better characterise the overall consequences of significant vessel wake heights in this region.

4.2 Timed intertidal and subtidal surveys

This component of the study aimed to assess temporal changes in biodiversity within the Derwent Estuary in a broader context to the fixed depth transect approach discussed above. This allowed a wider range of habitats to be searched, from intertidal shores to the full depth-range of reefs present and over adjacent sediments. While the primary focus was on describing shifts in species richness and composition over time, the timed surveys – consistent with the 2010 methodology – were also designed to maximise the detection of rare or cryptic species, and those not typically encountered on fixed depth transect surveys, such as intertidal species.

As expected, despite extensive searches conducted in both intertidal and subtidal habitats, *Marginaster littoralis*, the native endangered Derwent River seastar, was not recorded during

the 2024 surveys, confirming the tentative extinction of this species as evaluated by O'hara et al. (2018). Similarly, two other endangered seastar species, *Parvulastra vivipara* (the live-bearing seastar) and *Smilasterias tasmaniae* (the Bruny Island seastar), were also absent from all surveyed sites (including the intertidal habitats). This is in line with the surveys conducted in 2010, where no individuals were recorded for any of these rare species (Barrett et al. 2012). Both *P. regularis* and *P. elongatus*, suggested to be a source of competition for *M. littoralis* in 2010, remain dominant in the estuary, with timed surveys confirming similar patterns of abundance and distribution to the standard surveys.

Five independent sightings of the endangered spotted handfish (*Brachionichthys hirsutus*) were recorded in 2024, marking an increase from the single individual observed at Bellerive Bluff in 2010. Most sightings occurred on sand and ecotone habitats at 5-8.9 m, consistent with known habitat preferences, although one individual was unexpectedly recorded on reef habitat at Battery Point in the 0-1 m depth zone. This species was detected at four additional sites compared to 2010, a promising indication of potential recovery and a reflection of the ongoing conservation efforts targeting this highly range-restricted species.

Overall, the patterns observed by the timed-swim approach were generally consistent with those observed on the fixed transect surveys, with a general increase in many species further up the estuary in response to relatively stable conditions in the years prior to the 2024 survey. These included introduced species, such as the fishes *F. varium* and *F. gymnotum* and the seastars *A. amurensis* and *P. regularis*, as well as a range of native species including fishes such as *N. balteatus*, mobile invertebrates such as urchin *H. erythrogramma* and the lobster *J. edwardsii*, and sessile invertebrates such as the native oyster *Ostrea angasi*. Some of these may be responding to the observed increases in algal cover as species, such as the urchins and the rock whiting *N. baleatus* that are herbivorous. Algal cover increased significantly in the mid to upper estuary, particularly due to a major rise in the red algae *Rhodomenia leptophylla* (described from the transect studies) which dominated sites from Bedlam Walls to Rosny Point in 2024. Previously, the extensive polychaete matting present at these sites had precluded algal cover forming.

Timed surveys conducted across multiple depth zones enabled us to assess depth-specific patterns and changes in the distribution of many species through time. While these are detailed in the tables presented as Appendices 14-19, we have simplified reporting here by only focussing on the most significant depth-related change observed (on the set of species examined by the timed-swim method). The most notable change was in the abundance of *Ecklonia radiata*, a canopy-forming kelp species, as this serves as an effective indicator of water clarity due to its dependence on light availability for growth and distribution (Miller et al. 2011; Blain et al. 2020; Janssen et al. 2024). Improved water quality allows greater light penetration at depth, which can support increased abundance and deeper distribution of kelp, like *E. radiata* (Deregibus et al. 2016; Chen et al. 2024). In 2024, the rank abundance of this species increased across multiple sites, with new observations also recorded at several locations and depth zones. At all sites where *E. radiata* was previously recorded in the 0-1 m, 1-4.9 m and 5-8.9 m depth ranges, its abundance increased in 2024. Some sites – particularly Rosny Point North, Bellerive Bluff and Tranmere Point – showed dramatic increases in coverage, potentially reflecting improved water quality in the mid and lower estuary. This expansion in kelp coverage at deeper depths, coupled with the disappearance of widespread

tubeworm matting previously recorded between Bedlam Walls and Battery Point, suggests an overall positive ecological shift and signs of recovery in the estuary.

More general patterns observed from the timed swim surveys include an increasing spread of the introduced red algae *Grateloupia turuturu* upstream from Bellerive/Howrah Bluff to Cornelian Bay, as expected for this recently introduced invasive species. The species first appeared in the vicinity of Bicheno in the mid 2000's (Prof. Gary Saunders, pers.obs.) and has since spread throughout eastern and southeastern Tasmania, as well as in the Tamar estuary. However, despite it being highly abundant in the immediate intertidal, there was no evidence of this annual species displacing native species. Likewise, the large introduced brown seaweed *Undaria pinnatifida* remained at low levels and in the outer estuary sites, although as this is an annual species that is most prevalent from mid-winter to early summer, surveys undertaken in late spring are more suited to understanding its current influence. While the European shore crab *Carcinus maenas* was a recent arrival to the Derwent Estuary at the time of the 2010 study and was anticipated to both spread more widely and increase its abundance in following years, there has been no evidence of this during the recent survey. The New Zealand porcelain crab, *Petrolisthes elongatus* remains the most abundant introduced crab in the system and was extremely abundant under almost any rocks overturned in the mid to upper estuary, as per the 2010 study. However, this is primarily an intertidal species, so the threat is mostly constrained to that habitat.

Some of the more marine species displayed significant changes through time, including a marked increase in the eleven-armed seastar *Coscinasterias muricata*, although such variability in recruitment and subsequent population growth is well documented for this species (Glockner-Fagetti & Phillips 2020). Notable new inclusions included mado sweep, *Atypichthys strigatus*, sighted at a number of locations, as well as several pink snapper, *Chrysophrys auratus*, seen at Battery Point. Both of these are range-extending species and initial indicators of the increasing influence of coastal warming in this region, particularly over summer. While another range-extending species, the yellowtail kingfish, *Seriola lalandi*, was not sighted on surveys, it was present in large numbers at times, resulting in shore and boat fishing activities limiting access at a number of our sites during the survey. At one site, Howrah Bluff, the seabed was covered in hundreds of snagged lures, with associated fishing line attached, representing a significant marine pollution hazard and safety issue.

In response to increasing interest in re-establishing native oyster (*Ostrea angasi*) reefs throughout the historical range of this species (<https://www.nature.org/en-us/about-us/where-we-work/asia-pacific/australia/stories-in-australia/restoring-australias-lost-shellfish-reefs/>), we paid particular attention to validating our field observations of this species in the Derwent and the extent that they overlap the distribution of the invasive introduced Pacific oyster (*Crassostrea gigas*, now recently re-named as *Magallana gigas*), and may be confused with on field surveys. This included sub-sampling at a range of depths and locations throughout the estuary and undertaking lab-based identification by a taxonomic expert using key diagnostic features, then further validation of these results with a molecular qPCR assay for full confidence. Typically, these identifications confirmed field-based observations of the distribution of these species. These showed both species to be distributed throughout most of the sites surveyed, with *C. gigas* dominating the intertidal, gradually transitioning to depths around 5 m or more where *O. angasi* became dominant. *O. angasi* was notably abundant on the fixed transect surveys at 5 m depth at Tranmere reef, as well as at sites such as Crayfish

Point on the timed swim surveys. However, the lab-based identification and subsequent DNA testing confirmed that these species can have significant overlap in shell morphology that can cause mis-identification in field surveys and it is recommended that future surveys do include a component of additional expert validation to validate field observations. Notably also, is that *C. gigas* can also form distinct and extensive oyster reefs, based on observation of these reefs at Cadbury Point.

5. Conclusions

This study provides a comprehensive temporal comparison of biodiversity on rocky reefs and adjacent sediments in the Derwent Estuary, highlighting both persistent threats and encouraging signs of ecological recovery.

Notably, invasive species such as *Asterias amurensis*, *Patiriella regularis* and *Petrolisthes elongatus* remain widespread, particularly in the mid-estuary, underscoring the importance of ongoing management strategies that mitigate the extent of disturbed and nutrient enriched habitats that facilitate such species. While many introduced species are found in the Derwent Estuary, there were no major changes since the previous survey in 2010, aside from the expected increase in the recently introduced red algae *Grateloupia turuturu*, now highly conspicuous in the immediate subtidal zone at many sites.

Despite this, some notable positive trends were observed. The key observation was a marked increase in algal species on rocky reefs in the mid to upper estuary, with the red algae *Rhodomenia leptophylla* dominating reefs seaward of Dowsings Point to Rosny Point, and increasing cover of the golden kelp, *Ecklonia radiata* from Rosny Point seaward. This included increasing depth distribution of *E. radiata* on mid-estuary reefs and a massive decline in a tough polychaete matting that had previously covered these same sites, likely excluding algal settlement and growth. It is likely that the loss of this matting and increased light penetration (allowing algae to grow deeper) is the result of declining silt and organic enrichment in the estuary over this time, as the polychaete matting is a common indicator of highly organically enriched estuaries (Quiroga et al. 1999).

These changes in algal cover are likely to have ongoing influences on fish and mobile invertebrate species that are supported by the algal production, allowing species such as sea urchins to increase at these sites.

A range of species appear to have increased their abundance and upstream distribution, including the introduced fish species *F. gymnotum* and *F. varium*, and seastars *A. amurensis* and *P. regularis*, however, it is anticipated that this is likely due to a prolonged period of low major flood activity prior to the 2024 surveys. During major floods, much of the upper and mid estuary is inundated with freshwater that is toxic to most marine species, sharply limiting their upstream distribution in subsequent months to years.

While many other species changed in abundance and distribution as may be expected in a dynamic estuarine environment, the most significant change was in the abundance of the fish, the bastard trumpeter, *Latridopsis forsteri*. This species, found in resident schools on a number of reefs in 2010 underwent a major (more than tenfold) decline by 2024, one not observed in other surveys at Tinderbox marine reserve where they are protected. It is likely

that much of this decline is in response to a major increase in recreational spearfishing in the Derwent Estuary over this period as fishing practices evolve.

Together, these findings reinforce the importance of long-term ecological monitoring in dynamic systems like the Derwent and provide essential data for continued conservation and management of this ecologically and economically important estuary. They document natural and anthropogenic changes, and responses to human pressures and associated management responses, hopefully underpinning similar future surveys in the future.

6. Acknowledgements

We would like to acknowledge Graham Edgar and Elizabeth Oh, Carolina Zagal, Dane Jones and Amelia Fowles for contribution to the initial baseline survey in 2010, with support funding from an ARC-Linkage grant and a Caring for our Country Grant to the Derwent Estuary Program. For the 2024 survey and reporting, we would like to thank the Derwent Estuary Program for inception and funding (Ursula Taylor and Akira Weller-Wong). WE would also like to thank Hunter Forbes, German Soeler, Jillian Fitzpatrick, Olivia Stowell, Lotus Blossom-Dumpleton, Lara Denis-Roy, Dan Haley and Marianne Pelletier for their assistance with fieldwork and data collection. Finally, Adam Davey and Alex Coutts for expert taxonomic identification of oysters followed by DNA-based validation.

7. References

- Alves AT, Petsch DK, and Barros F (2020) Drivers of benthic metacommunity structure along tropical estuaries. *Scientific Reports* **10**:1739
- Baptista J, Martinho F, Nyitrai D, Pardal MA, and Dolbeth M (2015) Long-term functional changes in an estuarine fish assemblage. *Marine pollution bulletin* **97**:125-134
- Barrett N, Edgar G, Zagal CJ, Oh E, and Jones D (2012) Surveys of intertidal and subtidal biota of the Derwent Estuary-2010.
- Barrett NS, Edgar GJ, Buxton CD, and Haddon M (2007) Changes in fish assemblages following 10 years of protection in Tasmanian marine protected areas. *Journal of Experimental Marine Biology and Ecology* **345**:141-157
- Bates AE, Barrett NS, Stuart-Smith RD, Holbrook NJ, Thompson PA, and Edgar GJ (2014) Resilience and signatures of tropicalization in protected reef fish communities. *Nature Climate Change* **4**:62-67
- Beeton R (2009) Advice to the Minister for the Environment, Heritage and the Arts from the Threatened Species Scientific Committee (the Committee) on Amendment to the list of Threatened Species under the Environment Protection and Biodiversity Conservation Act 1999 (EPBC Act). . In: Ministry for the Environment HaTA, (ed), Canberra
- Bishop MJ, and Chapman MG (2004) Managerial decisions as experiments: An opportunity to determine the ecological impact of boat-generated waves on macrobenthic infauna. *Estuarine, Coastal and Shelf Science* **61**:613-622
- Blain CO, Rees TaV, Christine Hansen S, and Shears NT (2020) Morphology and photosynthetic response of the kelp *Ecklonia radiata* across a turbidity gradient. *Limnology and Oceanography* **65**:529-544
- Bruce BD, Green MA, and Last P (1997) Aspects of the biology of the endangered spotted handfish, *Brachionichthys hirsutus* (Lophiiformes: Brachionichthyidae) off southern Australia. *ishConferenCe*:369

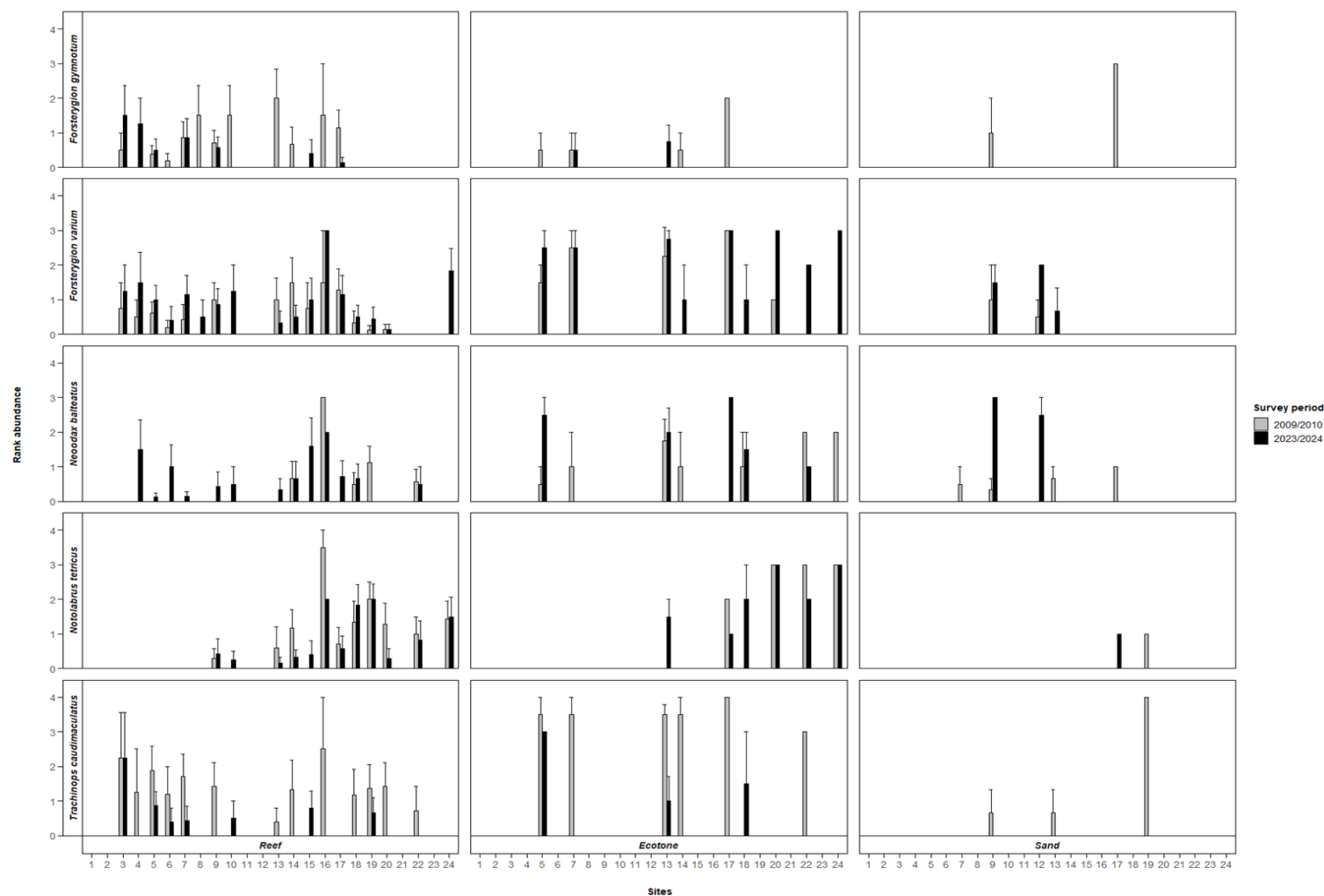
- Bureau of Meteorology. (2016) Tasmanian record major flooding event - June 2016.
- Chen J, Wang K, Li X, Zhao X, Cheng X, Liu Z, Zhang J, and Zhang S (2024) Remote sensing inversion of water quality parameters (TSM, Chl-a, and CDOM) in subtidal seaweed beds and surrounding waters. *Ecological Indicators* **167**:112716
- Cooper JaG (2002) The role of extreme floods in estuary-coastal behaviour: contrasts between river- and tide-dominated microtidal estuaries. *Sedimentary Geology* **150**:123-137
- Coutts, A., O'Brien, A., Weeks, A., Swearer, S., van Rooyen, A., Branigan, S., Morris, R. (2022) An environmental DNA approach to informing restoration of the functionally extinct oyster, *Ostrea angasi*. *Aquatic Conservation: Marine and Freshwater Ecosystems*. **32**: 1732–1744
- Davis TR, Larkin MF, Forbes A, Veenhof RJ, Scott A, and Coleman MA (2022) Extreme flooding and reduced salinity causes mass mortality of nearshore kelp forests. *Estuarine, Coastal and Shelf Science* **275**:107960
- De Souza Machado AA, Spencer K, Kloas W, Toffolon M, and Zarfl C (2016) Metal fate and effects in estuaries: A review and conceptual model for better understanding of toxicity. *Science of The Total Environment* **541**:268-281
- Dep. (2020) State of the Derwent estuary — 2020 update. An update and review of environmental data and activities. Hobart, Australia
- Dep. (2021) Metal contamination in fish and shellfish of the Derwent estuary - A summary of results to December 2020. Hobart
- Deregibus D, Quartino ML, Campana GL, Momo FR, Wiencke C, and Zacher K (2016) Photosynthetic light requirements and vertical distribution of macroalgae in newly ice-free areas in Potter Cove, South Shetland Islands, Antarctica. *Polar Biology* **39**:153-166
- Derwent Estuary Program (2024) URL www.reportcard.derwentestuary.org.au/2023
- Edgar GJ. (2008) Australian marine life: the plants and animals of temperate waters. New Holland, Sydney
- Edgar GJ, Barrett NS, and Last PR (1999) The distribution of macroinvertebrates and fishes in Tasmanian estuaries. *Journal of Biogeography* **26**:1169-1189
- Eriksson BK, Sandström A, Isæus M, Schreiber H, and Karås P (2004) Effects of boating activities on aquatic vegetation in the Stockholm archipelago, Baltic Sea. *Estuarine, Coastal and Shelf Science* **61**:339-349
- Galobart C, Ballesteros E, Golo R, and Cebrian E (2023) Addressing marine restoration success: evidence of species and functional diversity recovery in a ten-year restored macroalgal forest. *Frontiers in Marine Science* **Volume 10 - 2023**
- Glockner-Fagetti A, and Phillips NE (2020) Species assemblage and recruitment patterns of echinoderms on shallow rocky reefs in central New Zealand. *New Zealand Journal of Marine and Freshwater Research* **54**:286-304
- Hardwick-Witman MN, and Mathieson AC (1983) Intertidal macroalgae and macroinvertebrates: Seasonal and spatial abundance patterns along an estuarine gradient. *Estuarine, Coastal and Shelf Science* **16**:113-129
- Harries D, and Croome R (1989) A review of past and present inshore gill netting in Tasmania with particular reference to the Bastard trumpeter *Latridopsis Forsteri* Castelnau.
- Harrison LM, Coulthard TJ, Robins PE, and Lewis MJ (2022) Sensitivity of estuaries to compound flooding. *Estuaries and Coasts* **45**:1250-1269
- Hastings RA, Rutterford LA, Freer JJ, Collins RA, Simpson SD, and Genner MJ (2020) Climate Change Drives Poleward Increases and Equatorward Declines in Marine Species. *Current Biology* **30**:1572-1577.e1572
- Henderson CJ, Olds AD, Goodridge Gaines LA, Mosman JD, Perry HJ, Borland HP, and Gilby BL (2024) Flood effects on estuarine fish are mediated by seascape composition and context. *Marine Biology* **171**:138

- Hutson K, Ross D, Day R, and Ahern JJ (2005) Australian scallops do not recognise the introduced predatory seastar *Asterias amurensis*. *Marine Ecology Progress Series* **298**:305-309
- Janssen AR, Bishop MJ, Mayer-Pinto M, and Dafforn KA (2024) Morpho-physiological traits and tissue burdens of *Ecklonia radiata* linked to environmental variation in an urban estuary. *Marine Environmental Research* **199**:106572
- Lauchlan SS, and Nagelkerken I (2020) Species range shifts along multistressor mosaics in estuarine environments under future climate. *Fish and Fisheries* **21**:32-46
- Ling SD, Ibbott S, and Sanderson JC (2010) Recovery of canopy-forming macroalgae following removal of the enigmatic grazing sea urchin *Heliocidaris erythrogramma*. *Journal of Experimental Marine Biology and Ecology* **395**:135-146
- Ling SD, Johnson CR, Ridgway K, Hobday AJ, and Haddon M (2009) Climate-driven range extension of a sea urchin: inferring future trends by analysis of recent population dynamics. *Global Change Biology* **15**:719-731
- Ling SD, and Keane JP (2024) Climate-driven invasion and incipient warnings of kelp ecosystem collapse. *Nature Communications* **15**:400
- Ling SD, Scheibling RE, Rassweiler A, Johnson CR, Shears N, Connell SD, Salomon AK, Norderhaug KM, Pérez-Matus A, Hernández JC, Clemente S, Blamey LK, Hereu B, Ballesteros E, Sala E, Garrabou J, Cebrian E, Zabala M, Fujita D, and Johnson LE (2015) Global regime shift dynamics of catastrophic sea urchin overgrazing. *Philosophical Transactions of the Royal Society B: Biological Sciences* **370**:20130269
- Lord J, and Whitlatch R (2015) Predicting competitive shifts and responses to climate change based on latitudinal distributions of species assemblages. *Ecology* **96**:1264-1274
- Macleod C, and Coughanowr C (2019) Heavy metal pollution in the Derwent estuary: History, science and management. *Regional Studies in Marine Science* **32**:100866
- Miller SM, Hurd CL, and Wing SR (2011) Variations in growth, erosion, productivity, and morphology of *Ecklonia radiata* (Alariaceae; Laminariales) along a fjord in southern New Zealand. *Journal of Phycology* **47**:505-516
- Murphy RJ, and Lyle JM. (1999) Impact of gillnet fishing on inshore temperate reef fishes, with particular reference to banded morwong. *Tasmanian Aquaculture and Fisheries Institute*
- Neves LM, Teixeira TP, and Araújo FG (2011) Structure and dynamics of distinct fish assemblages in three reaches (upper, middle and lower) of an open tropical estuary in Brazil. *Marine Ecology* **32**:115-131
- Nyrstar Hobart. (2017) Triennial public environment report 2015–2017. Hobart, Tasmania
- O'hara TD, Mah CL, Hipsley CA, Bribiesca-Contreras G, and Barrett NS (2018) The Derwent River seastar: re-evaluation of a critically endangered marine invertebrate. *Zoological Journal of the Linnean Society* **186**:483-490
- Pinsky ML, Selden RL, and Kitchel ZJ (2020) Climate-Driven Shifts in Marine Species Ranges: Scaling from Organisms to Communities. *Annual Review of Marine Science* **12**:153-179
- Potter IC, Chuwen BM, Hoeksema SD, and Elliott M (2010) The concept of an estuary: A definition that incorporates systems which can become closed to the ocean and hypersaline. *Estuarine, Coastal and Shelf Science* **87**:497-500
- Quiroga E, Soto R, Rozba~Zyloz N, Poliquetos L, and Su Y (1999) The spionid polychaetes (Polychaeta: Spionidae) and its importance in the community structure: a study case in the Bay of Iquique, northern Chile (20 degree 11'S; 70 degree 10'W). **63**:1-16
- Rasmussen JA, Ingleton T, Bennett WW, Pearson RM, Ca M, Foulsham E, Hanslow D, Scanes PR, and Connolly RM (2024) The effects of estuarine outflows on coastal marine ecosystems in New South Wales, Australia. *Marine pollution bulletin* **208**:116915

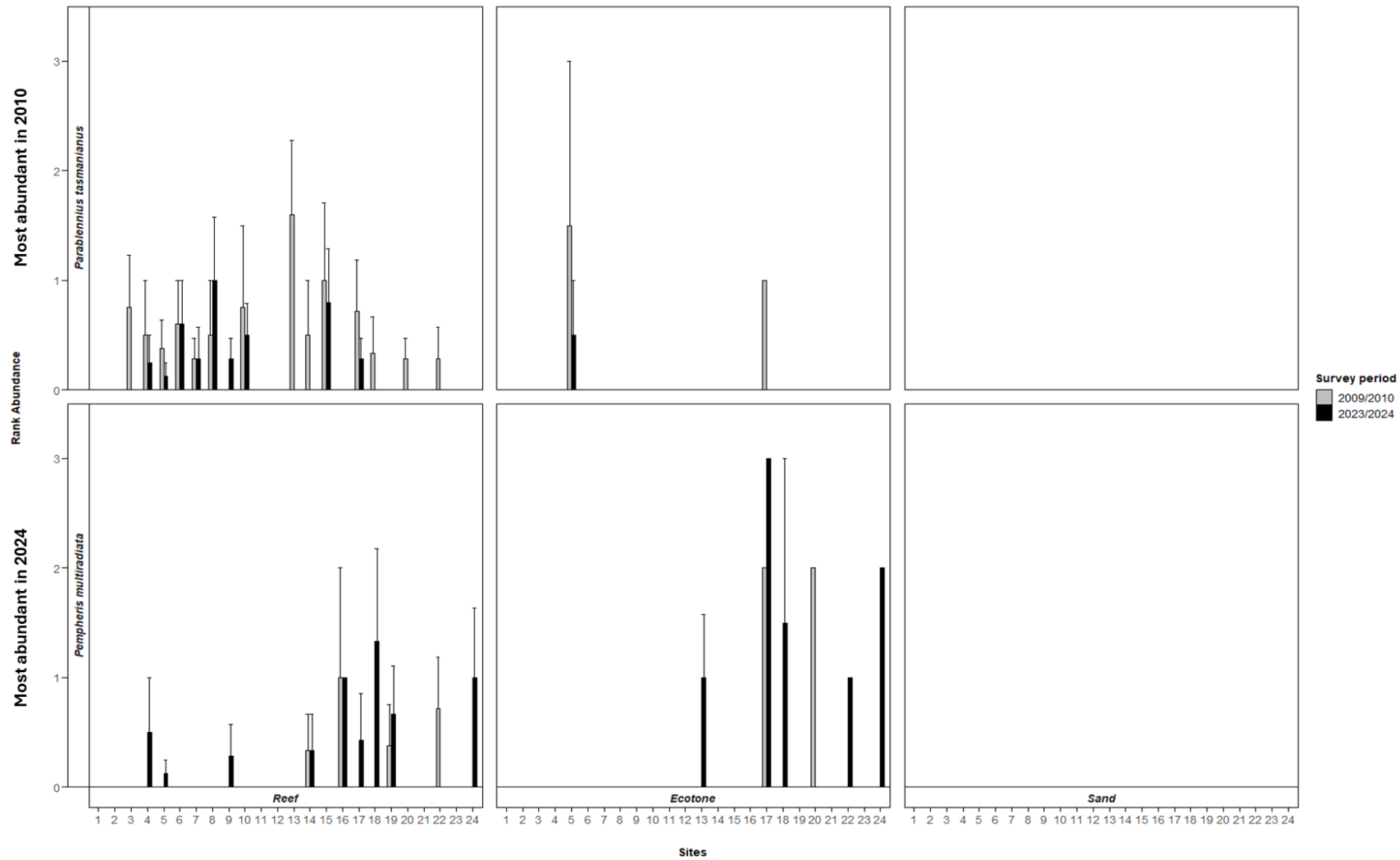
- Rogers K, and Woodroffe CD (2016) Geomorphology as an indicator of the biophysical vulnerability of estuaries to coastal and flood hazards in a changing climate. *Journal of Coastal Conservation* **20**:127-144
- Rose TH, Tweedley JR, Warwick RM, and Potter IC (2019) Zooplankton dynamics in a highly eutrophic microtidal estuary. *Marine pollution bulletin* **142**:433-451
- Ross D, Johnson C, and Hewitt C (2006) Abundance of the introduced seastar, *Asterias amurensis*, and spatial variability in soft sediment assemblages in SE Tasmania: Clear correlations but complex interpretation. *Estuarine, Coastal and Shelf Science* **67**:695-707
- Ross DJ, Johnson CR, Hewitt CL, and Ruiz GM (2004) Interaction and impacts of two introduced species on a soft-sediment marine assemblage in SE Tasmania. *Marine Biology* **144**:747-756
- Scheibling RE, and Filbee-Dexter K (2014) Sea urchin barrens as alternative stable states of collapsed kelp ecosystems. *Marine Ecology Progress Series* **495**:1-25
- Sorte CJB, Williams SL, and Carlton JT (2010) Marine range shifts and species introductions: comparative spread rates and community impacts. *Global Ecology and Biogeography* **19**:303-316
- Stuart-Smith J, Edgar GJ, Last P, Linardich C, Lynch T, Barrett N, Bessell T, Wong L, and Stuart-Smith RD (2020) Conservation challenges for the most threatened family of marine bony fishes (handfishes: *Brachionichthyidae*). *Biological Conservation* **252**:108831
- Wang H, García Molinos J, Heino J, Zhang H, Zhang P, and Xu J (2021) Eutrophication causes invertebrate biodiversity loss and decreases cross-taxon congruence across anthropogenically-disturbed lakes. *Environment International* **153**:106494
- Willis T, and Anderson M (2003) Structure of cryptic reef fish assemblages: Relationships with habitat characteristics and predator density. *Marine Ecology Progress Series* **257**:209-221

8. Appendix

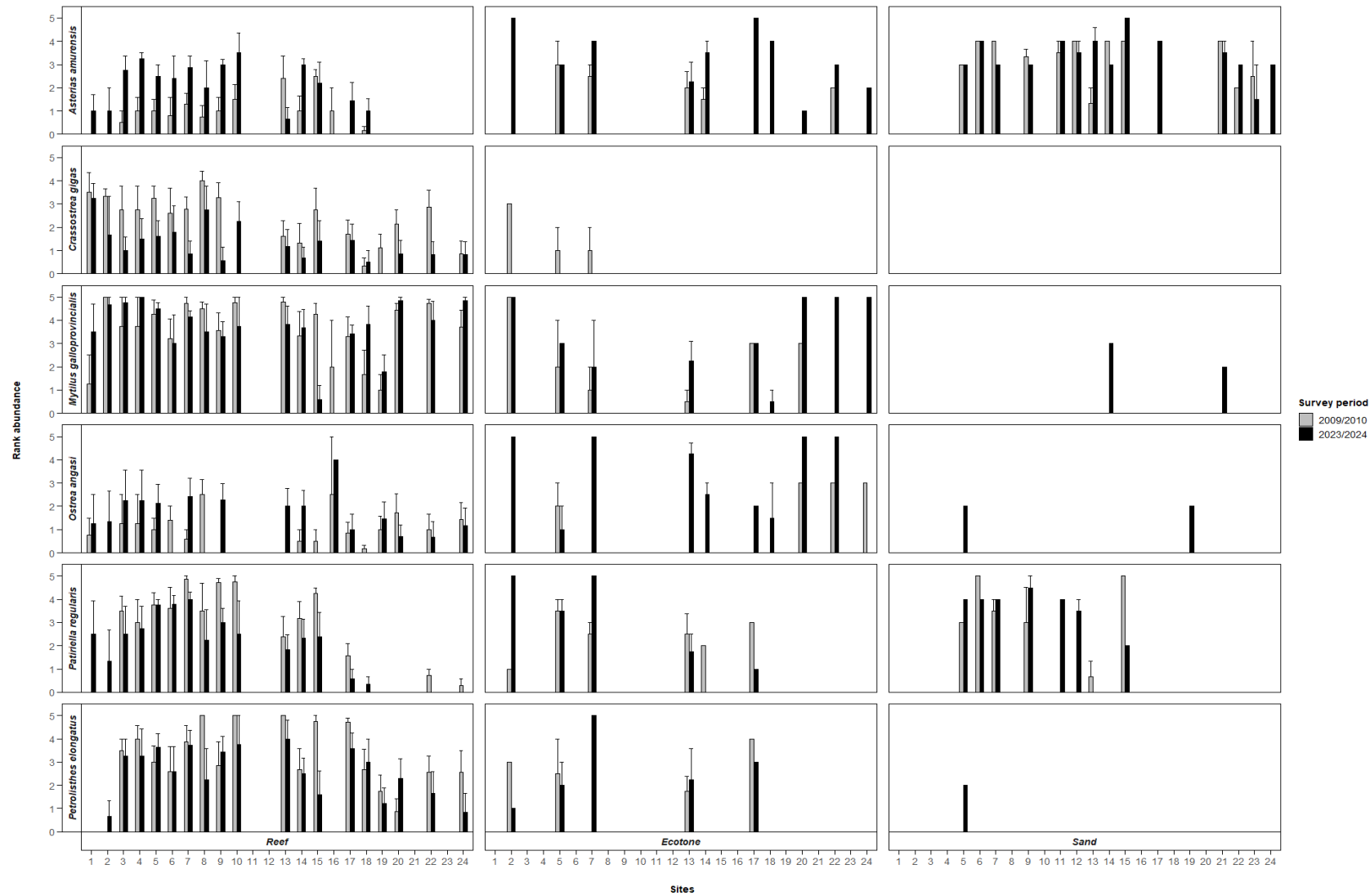
Appendix 1: Habitat-specific rank abundance (mean $\pm$ SE) of the most abundant fish species in Derwent in 2010 (grey) and 2024 (black). The five species shown were among the top 6 most abundant in both survey periods, and are presented alphabetically.



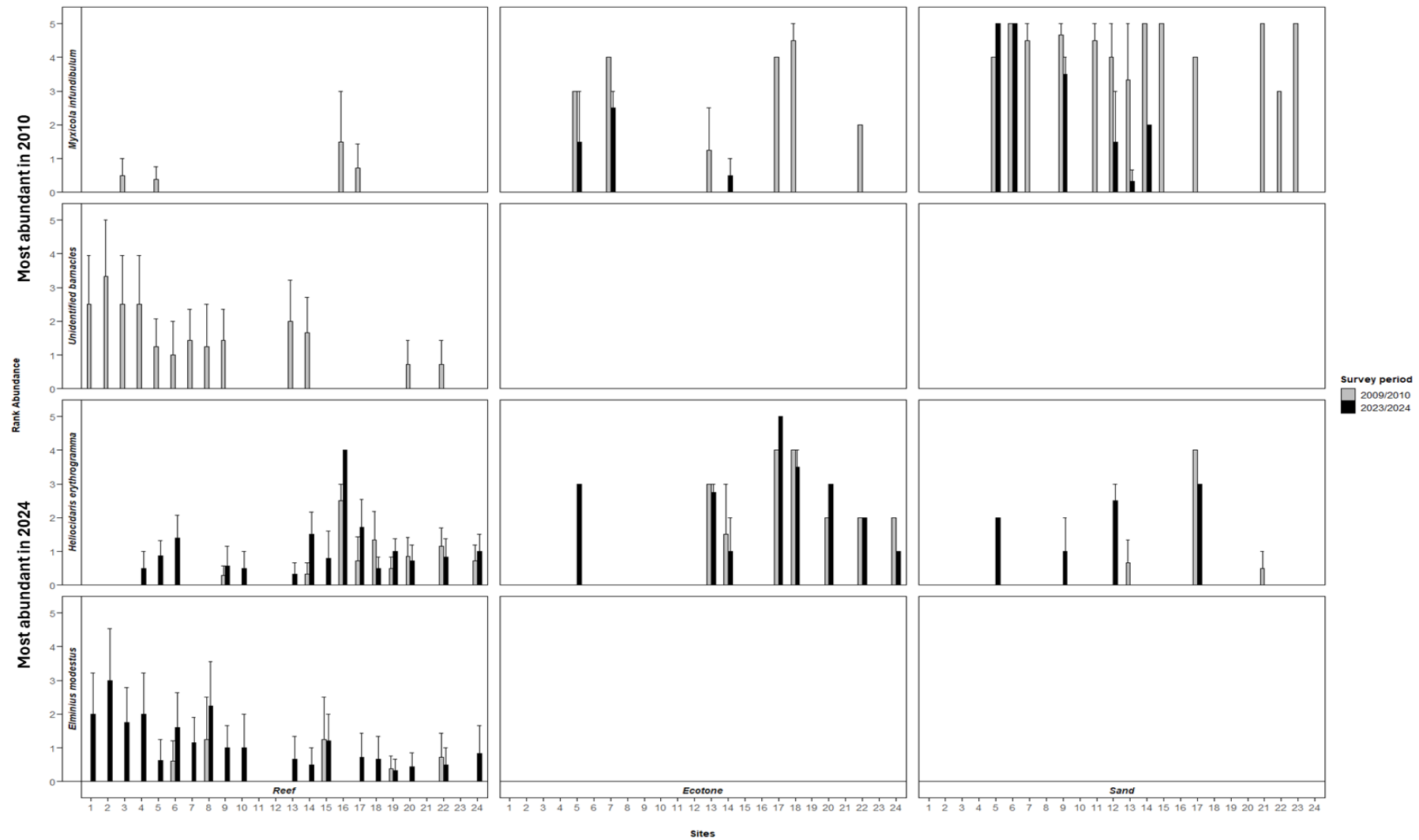
Appendix 2: Habitat-specific rank abundance (mean $\pm$ SE) of fish species uniquely among the top 6 most abundant in 2010 and 2024 in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill.



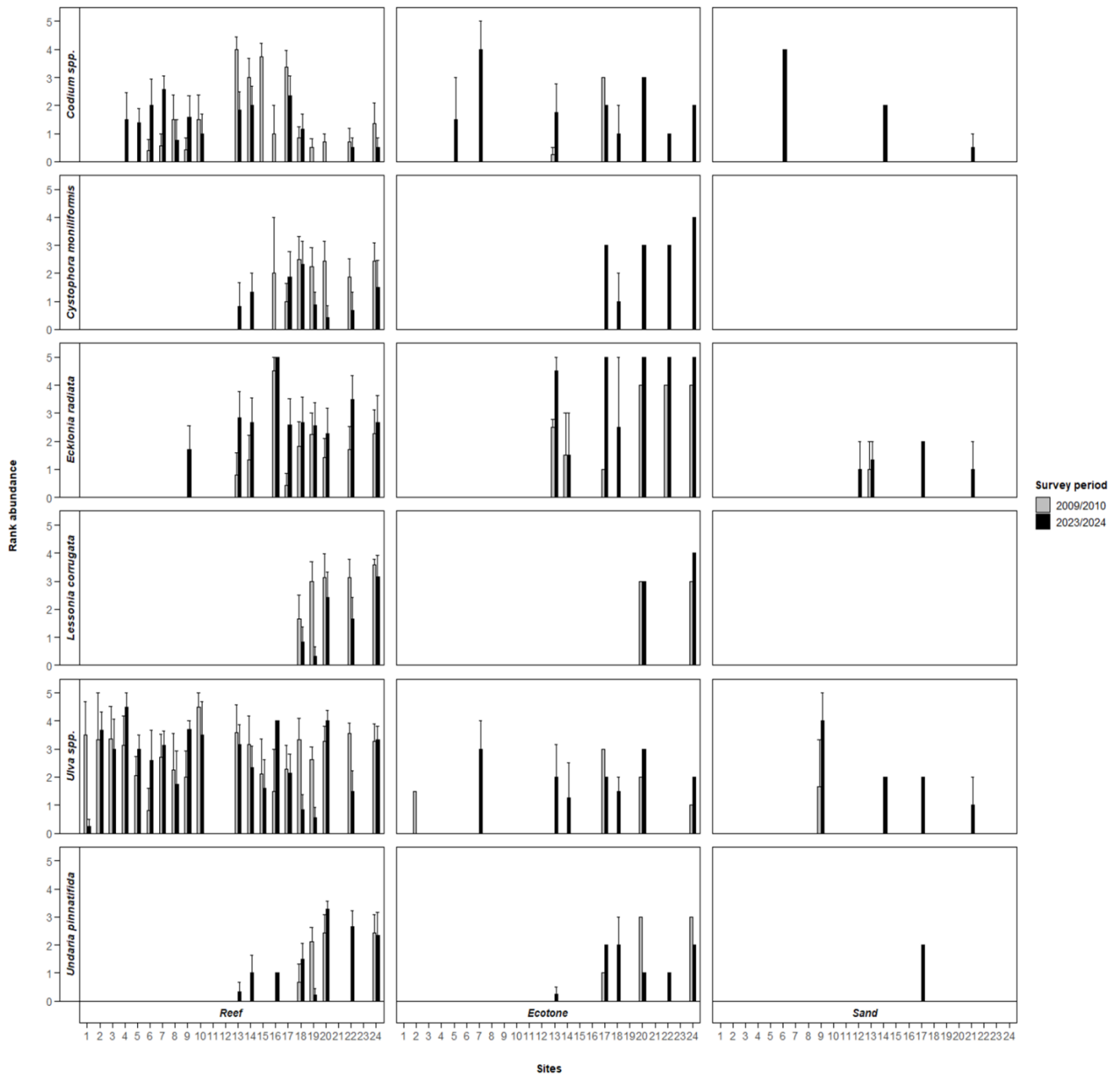
Appendix 3: Habitat-specific rank abundance (mean $\pm$ SE) of the most abundant invertebrate species in Derwent in 2010 (grey) and 2024 (black). The six species shown were among the top 8 most abundant in both survey periods, and are presented alphabetically.



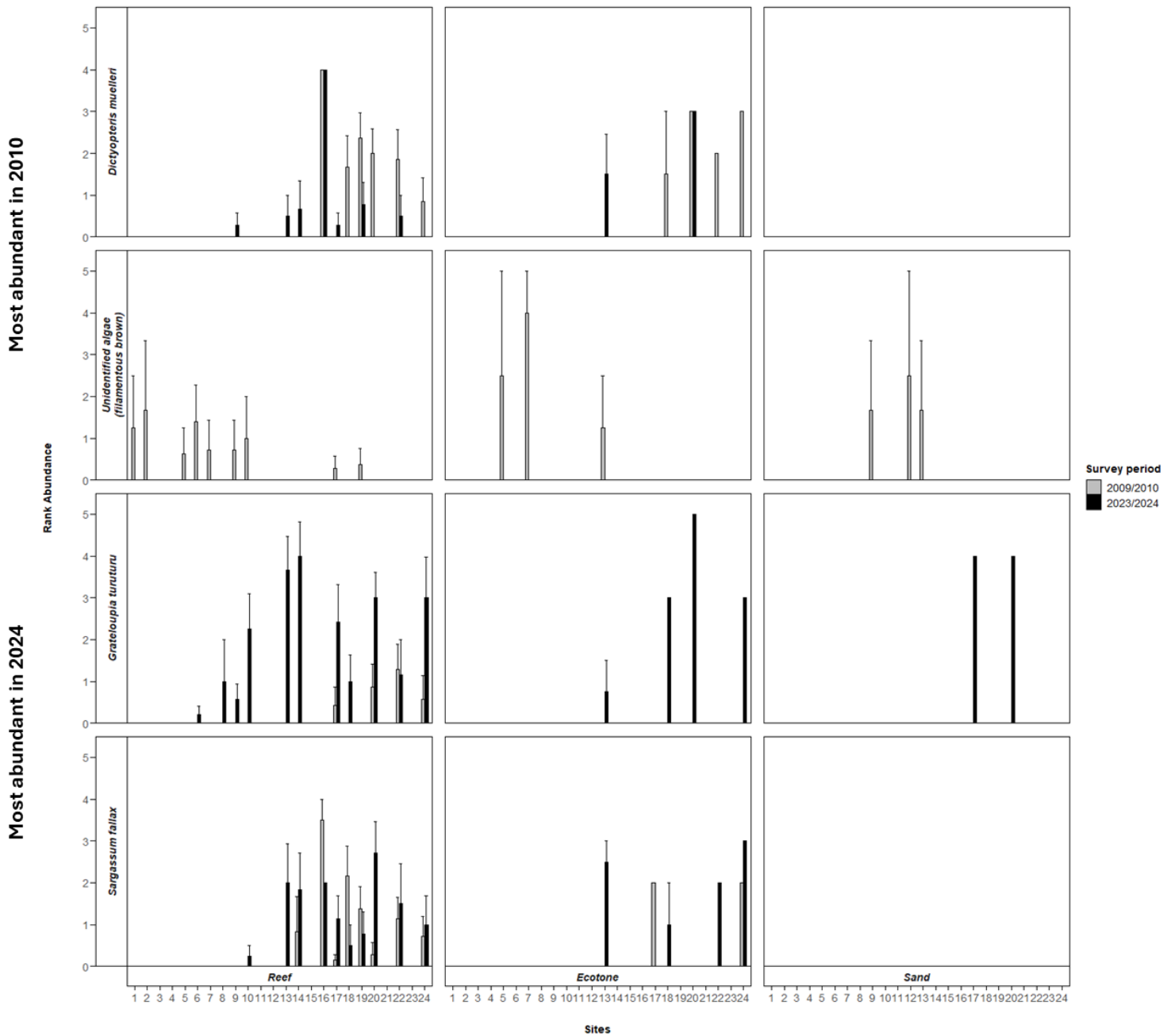
Appendix 4: Habitat-specific rank abundance (mean $\pm$ SE) of invertebrate species uniquely among the top 8 most abundant in 2010 and 2024 in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill.



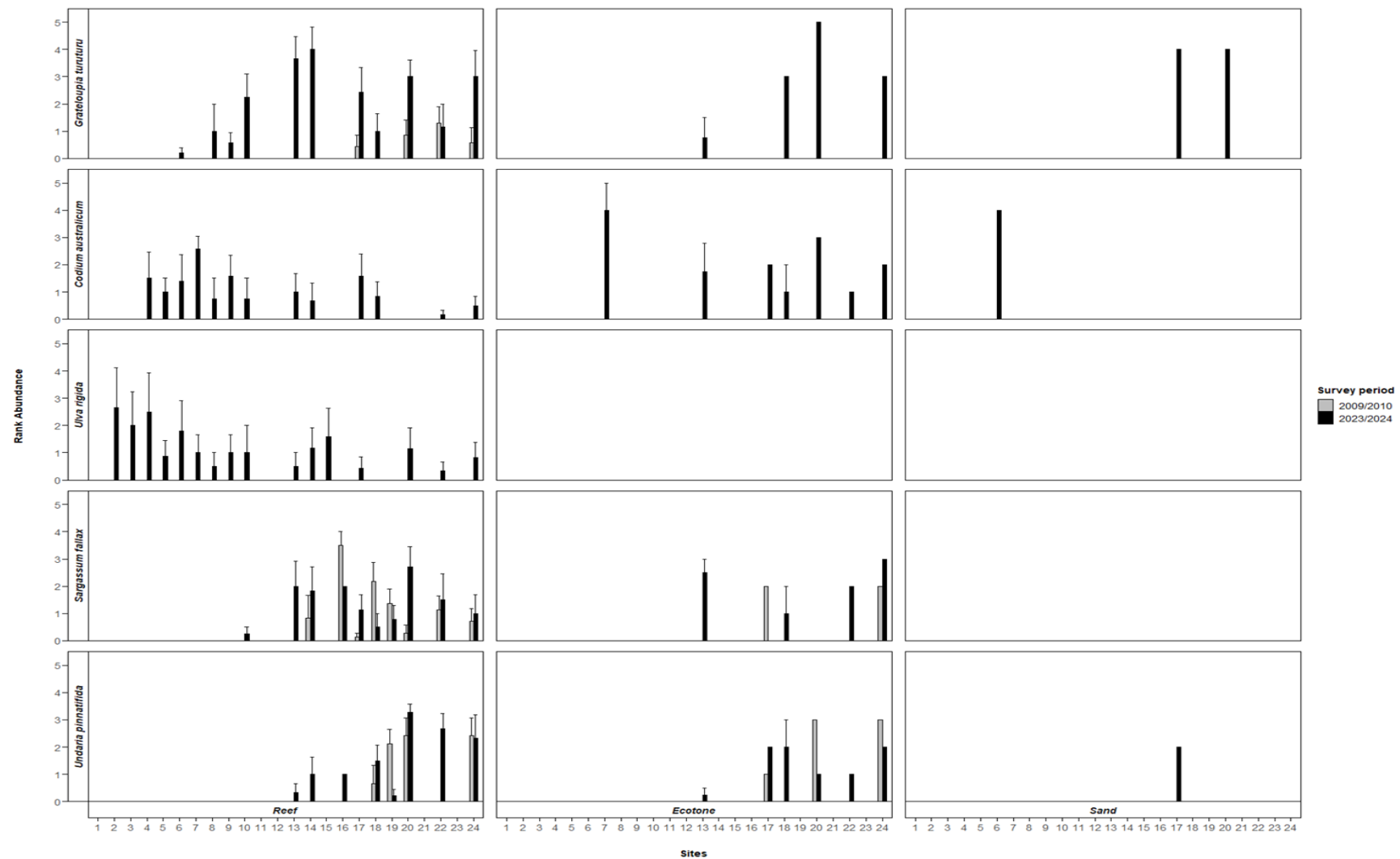
Appendix 5: Habitat-specific rank abundance (mean $\pm$ SE) of the most abundant algal species in Derwent in 2010 (grey) and 2024 (black). The three species shown were among the top 8 most abundant in both survey periods, and are presented alphabetically.



Appendix 6: Habitat-specific rank abundance (mean $\pm$ SE) of algal species uniquely among the top 8 most abundant in 2010 and 2024 in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill.



Appendix 7: Habitat-specific rank abundance (mean $\pm$ SE) of algal species uniquely among the top 8 most abundant in 2024 in the Derwent. While each species peaked in abundance during one period, changes across both survey periods are indicated by grey (2010) and black (2024) fill.



Appendix 8: Total fish abundances recorded during standard surveys in February-April 2010, * = introduced species.

METHOD 1, 2009/2010	Sites	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
Species	Number of transects Common name/Depth	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	
		1	4	4	3	2	2	4	4	5	2	3	4	4	5	4	5	5	5	5	
<i>Acanthaluteres spilomelanurus</i>	Bridled leatherjacket	5										2				18					25
<i>Acanthaluteres vittiger</i>	Toothbrush leatherjacket								9	26				2	3	138		2	1	9	190
<i>Aplodactylus arcitidens</i>	Marblefish											1			1		3	1		3	9
<i>Aracana aurita</i>	Shaws cowfish												1			1					12
<i>Arenigobius bifrenatus</i>	Bridled goby	5													2	1	3	1	3	1	5
<i>Arripis</i> spp.	Australian salmon												15								15
<i>Blenniud</i> spp.	Blenny											1									1
<i>Cephaloscyllium laticeps</i>	Draughtboard shark													2	2			5		1	10
<i>Chirodactylus spectabilis</i>	Banded morwong																				6
<i>Cristiceps australis</i>	Crested weedfish											6									6
<i>Dinolestes lewini</i>	Long-fin pike													502		61	171		7	1	742
<i>Diodon nichthemerus</i>	Globe fish											3	3		1	1	1	1			10
<i>Dotalabrus aurantiacus</i>	Castelnaus wrasse												2	12	1	1		2	8	12	38
<i>Eubalichthys gunnii</i>	Gunn's leatherjacket																		1		1
<i>Favonigobius lateralis</i>	Long-finned goby	40	29	13																	82
<i>Forsterygion gymnotum</i>	Estuarine threefin*		35		2		3	2	1												43
<i>Forsterygion varium</i>	Many-rayed threefin*		18	12	55		10	12	52	18	21	172					2				372
<i>Girella zebra</i>	Zebra fish											1									1
<i>Gobiid</i> spp.	Unidentified goby			1			5														6
<i>Hippocampus abdominalis</i>	Bigbelly seahorse	5	1								1										7
<i>Hippocampus</i> spp.	Seahorse					1															1
<i>Latridopsis forsteri</i>	Bastard trumpeter									1			386		13	130	232	5	3	114	884
<i>Meuschenia australis</i>	Brown-striped leatherjacket												2		6	1		2		2	13
<i>Meuschenia freycineti</i>	Six-spine leatherjacket												2		1						3
<i>Myliobatis australis</i>	Eagle ray												1								1
<i>Neoodax balteatus</i>	Little rock whiting			1			2		2	10	3	17	4	11		24		1	3	2	82
<i>Neosebastes scorpaenoides</i>	Common gurnard perch				1												1				2
<i>Nesogobius hinsbyi</i>	Hinsbys goby						2														2
<i>Notolabrus fucicola</i>	Purple wrasse											2	11	1	7	13	2	4	8	19	67
<i>Notolabrus tetricus</i>	Blue-throat wrasse			1					3	6		23	43	126	109	40	77	107	76	57	668
<i>Olisthops cyanomelas</i>	Herring cale													1	15						16
<i>Parablennius tasmanianus</i>	Tasmanian blenny		6	8	2						3										19
<i>Pempheris multiradiata</i>	Common bulkseye											9					76		35	1	121
<i>Pentaceropsis recurvirostris</i>	Long-snouted boarfish													1						3	4
<i>Phyllopteryx taeniolatus</i>	Weedy seadragon															1	4				5
<i>Pictilabrus laticlavus</i>	Senator wrasse												5	39	18		7	10	28	8	115
<i>Platycephalid</i> spp.	Flathead					1															1
<i>Pseudocaranx dentex</i>	White trevally																20				20
<i>Pseudocaranx georgianus</i>	Silver trevally												85								85
<i>Pseudogoniistius nigripes</i>	Magpie perch																		1		1
<i>Pseudolabrus rubicundus</i>	Rosy wrasse													1							1
<i>Pseudophycis bachus</i>	Red cod														1		1			2	4
<i>Pseudophycis barbata</i>	Bearded cod										1		1	1	1	1					4
<i>Siphonognathus beddomei</i>	Pencil weed whiting																	1			1
<i>Spiniraja whitleyi</i>	Melbourne skate																			1	1
<i>Syngnathid</i> spp.	Unidentified pipefish											1									1
<i>Trachinops caudimaculatus</i>	Hulafish		2638	3345	1490	106	145	228	107	98	233	1720	73	602	700	35	62	25	10	798	12415
<i>Trachurus declivis</i>	Jack mackerel									20							20				40
<i>Urolophus cruciatus</i>	Banded stingaree					1		2	2		1										6
<i>Vincentia conspersa</i>	Southern cardinalfish														1						1
Species richness (total species per site)		4	6	7	5	4	6	4	7	7	8	12	15	13	17	13	17	14	13	18	
Total species richness		50																			

Appendix 9: Total fish abundances recorded during standard surveys in 2023-2024, * = introduced species.

METHOD 1, 2023/2024	Sites	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
	Number of transects	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	
Species	Common name / Depth	1	4	4	3	2	2	2	4	4	2	3	4	3	5	4	4	5	5	5	
<i>Acanthaluteres spilomelanurus</i>	Bridled leatherjacket								8	40						10					58
<i>Acanthaluteres vittiger</i>	Toothbrush leatherjacket			1		3				1								1		2	8
<i>Apodactylus arcidens</i>	Marblefish																3	1	5	1	10
<i>Aracana aurita</i>	Shaws cowfish																1		1		2
<i>Atypichthys strigatus</i>	Bridled goby											20		3			12	26			61
<i>Cephaloscyllium laticeps</i>	Draughtboard shark													1	1				1	1	4
<i>Chirodactylus spectabilis</i>	Banded morwong														1	1	6				8
<i>Cristiceps australis</i>	Crested weedfish													1							1
<i>Dinolestes lewini</i>	Long-fin pike															2			39	70	111
<i>Diodon nichthemerus</i>	Globe fish				2				2				1				2	1		1	9
<i>Dotalabrus aurantiacus</i>	Castelnaus wrasse												3	2	2	1		2	1	1	12
<i>Enoplosus armatus</i>	Old wife																			2	2
<i>Eocallionymus papilio</i>	Painted stinkfish											1									1
<i>Forsterygion gymnotum</i>	Estuarine threefin*		3	9	7		14	6													39
<i>Forsterygion varium</i>	Many-rayed threefin*		6	39	11	2	30	4	11	7	42	9		6	6		31	2	1		207
<i>Haletta semifasciata</i>	Blue weed whiting													1							1
<i>Heteroclinus johnstoni</i>	Johnston's weedfish										1										1
<i>Heteroclinus tristis</i>	Sharp-nose weedfish																		1		1
<i>Latridopsis forsteri</i>	Bastard trumpeter												3				22	2	17	19	63
<i>Lotella rhacina</i>	Rock cod																		2		2
<i>Nemadactylus macropterus</i>	Jackass morwong																		1		1
<i>Neodax balteatus</i>	Little rock whiting	1		55		14	1	5	30	15	30	15	3	4	1	9	3	2		4	192
<i>Nesogobius hinsbyi</i>	Hinsbys goby	10																			10
<i>Nesogobius maccullochi</i>	Girdled goby	1			3																4
<i>Notolabrus fucicola</i>	Purple wrasse										2				1	3	5	3	27	5	46
<i>Notolabrus tetricus</i>	Blue-throat wrasse			2				45	1	1	3	14	19	16	26	27	28	79	120	19	400
<i>Olisthops cyanomelas</i>	Herring cale																	2		2	4
<i>Parablennius tasmanianus</i>	Tasmanian blenny				1	2	2				2										7
<i>Pempheris multiradiata</i>	Common bullseye			2				20	6	7		6			100			15	67	4	227
<i>Pentaceropsis recurvirostris</i>	Long-snouted boarfish																			1	1
<i>Pictilabrus laticlavus</i>	Senator wrasse								3	1		1	7	17	8		1	17	22	4	81
<i>Pseudogoniistius nigripes</i>	Magpie perch																		1	1	2
<i>Pseudolabrus rubicundus</i>	Rosy wrasse																	1			1
<i>Pseudophycis bachus</i>	Red cod			1																	1
<i>Scorpius aequipinnis</i>	Sea sweep															1					1
<i>Scorpius lineolata</i>	Silver sweep												2				5	1	38	4	50
<i>Siphonognathus beddomei</i>	Pencil weed whiting					2			2				1			9	1	13		4	32
<i>Trachinops caudimaculatus</i>	Hulafish		830	709	28	6	6	1			9			54	3	6	10	155	1	684	2502
<i>Trinorfolkia clarkei</i>	Clark's threefin						1				3										4
<i>Urolophus cruciatus</i>	Banded stingaree									2											2
Species richness (total species per site)		3	3	8	5	7	6	6	8	8	8	7	8	10	10	10	14	17	17	19	
Total species richness		40																			

Appendix 10: Invertebrate and cryptic fish abundances recorded during standard surveys in February-April 2010, * = introduced species.

METHOD 2, Inverts/Cryptic Fish	2009/2010	Site	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total	
Groupings	Species	Number of transects Common name / Depth	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	5	
Cryptic fish	<i>Blennioid spp.</i>	Blenny	1	4	4	3	2	2	4	4	5	2	3	4	4	5	4	5	5	5	5	5	2
	<i>Bovichtus angustifrons</i>	Dragonet											2				1						1
	<i>Cephaloscyllium laticeps</i>	Draughtboard shark								1												2	3
	<i>Forsterygion gymnotum</i>	Estuarine threefin*						9	1														10
	<i>Forsterygion varium</i>	Many-rayed threefin*				46		27	8	34	14	12	71	1				8					221
	<i>Gobioid spp.</i>	Unidentified goby	9	11	5				4	22	5												56
	<i>Heteroclinus johnstoni</i>	Johnstons weedfish		1										1	1								3
	<i>Hippocampus abdominalis</i>	Bigbelly seahorse	1	1			1			1	1												5
	<i>Nesogobius hinsbyi</i>	Hinsbys goby					20						1										21
	<i>Nesogobius maccullochi</i>	Girdled goby				1	3					2											6
	<i>Parablennius tasmanianus</i>	Tasmanian blenny					2	3	1		1												7
	<i>Pempheris multiradiata</i>	Common bullseye								2	1							32			17		52
	<i>Phyllopteryx taeniolatus</i>	Weedy seadragon																1					1
	<i>Scorpaena papillosa</i>	Southern rock cod							1	3	2		1	2	2	6	2	16	4		2		41
	<i>Trinorfolkia clarkei</i>	Clark's threefin						2	5	6				2	4	2	1	1	4				27
	<i>Urolophus cruciatus</i>	Banded stingaree																			1		1
Crustaceans	<i>Carcinus maenas</i>	Green crab*				3						3	8										14
	<i>Decapoda spp.</i>	Unidentified hermit crab									3			1		1	3	1	2				11
	<i>Halicarcinus ovatus</i>	Three-pronged flat spider crab											1										1
	<i>Jasus edwardsii</i>	Southern rock lobster													20	1		3	6	8	28		66
	<i>Malacostraca spp.</i>	Unidentified shrimp	5																				5
	<i>Metacarcinus novaezelandiae</i>	Piecrust crab*		3	2	2				1													8
	<i>Nectocarcinus tuberculosus</i>	Velvet crab															1						1
	<i>Pagurid spp.</i>	Hermit crab																	1				1
	<i>Paragrapsus laevis</i>	Shore crab				1																	1
	<i>Petrolisthes elongatus</i>	New Zealand porcelain crab*			13		1	17			1												32
	<i>Plagusia chabrus</i>	Red bait crab														2	1	10	1	10	4		28
	<i>Strigopagurus strigimanus</i>	Stridulating hermit crab																		1			1
Cnidaria	<i>Phlyctenactis tuberculosa</i>	Swimming anemone																		31			31
Echinoderms	<i>Amblypneustes ovum</i>	Short-spined urchin												48			43						91
	<i>Asterias amurensis</i>	Northern Pacific seastar*	16	10	9	2	5	9	5	2	26												84
	<i>Australostichopus mollis</i>	Sea cucumber									1					1	1	1	2				6
	<i>Cenolia trichoptera</i>	Feather star													25	8		24	27	90	450		624
	<i>Comanthus tasmaniae</i>	Tasmanian feather star													2				7	5			14
	<i>Coscinasterias muricata</i>	Eleven-armed seastar							1			3	1				1	5	2				13
	<i>Goniocidaris impressa</i>	Pencil urchin																	2				2
	<i>Goniocidaris spp.</i>	Unidentified urchin																	4				4
	<i>Goniocidaris tubaria</i>	Thorny sea urchin								1													1
	<i>Heliocidaris erythrogramma</i>	Purple urchin							105	20	42	163	100	11	49	17	23	191	153	95			969
	<i>Meridiastra calcar</i>	Eight-armed seastar								1			1		253		220	14					489
	<i>Nectria ocellata</i>	Ocellate seastar																		1			1
	<i>Patiriella regularis</i>	New Zealand common cushion star*	232	80	19	173	244	507	124	97	372	3						2					1853
	<i>Petricia vernicina</i>	Velvet seastar															1		1	1	5		8
	<i>Tosia australis</i>	Southern biscuit star												8			1		1	1			11
	<i>Tosia magnifica</i>	Magnificent biscuit star							1	1		3						1	2				8
	<i>Uniophora granifera</i>	Granular seastar																2	1				3
Molluscs	<i>Agnewia tritoniformis</i>	Murex snail														3	2	1					6
	<i>Argobuccinum pustulosum</i>	Pustular triton															2	1					3
	<i>Austrocochlea spp.</i>	Top snail														3		3					6
	<i>Cabestana spengleri</i>	Spengler's trumpet								1		2				22	14	28					67
	<i>Cabestana spp.</i>	Predatory whelks														14	2						16
	<i>Cabestana tabulata</i>	Shouldered triton																3					3
	<i>Cominella lineolata</i>	Lined whelk														7		2					9
	<i>Dicathais orbita</i>	White rock shell															2		7	2	1		12
	<i>Haliotis rubra</i>	Blacklip abalone												1		3		1	7	17	5		34
	<i>Lunella undulata</i>	Common warrener													1	40	7	7	17				72
	<i>Maoricolpus roseus</i>	New Zealand screwshell									60												60
	<i>Pleuroploca australasia</i>	Tulip shell									1						1						2
	<i>Scutus antipodes</i>	Elephant snail															2			1			3
Species richness (total species per site)			3	6	5	7	4	8	8	14	16	8	12	11	8	17	18	24	22	13	11		
Total species richness			59																				

Appendix 11: Invertebrate and cryptic fish abundances recorded during standard surveys in 2023-2024, * = introduced species.

METHOD 2, Inverts/Cryptic Fish	2023/2024	Site	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
Groupings	Species	Number of transects Common name / Depth	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	4
			1	4	4	3	2	2	2	4	4	2	3	4	3	5	4	4	4	5	5	5
Cryptic fish	<i>Bovichtus angustifrons</i>	Dragonet														1		1		1		3
	<i>Cephaloscyllium laticeps</i>	Draughtboard shark																		1	2	3
	<i>Conger verreauxi</i>	Southern conger																			1	1
	<i>Cristiceps australis</i>	Crested weedfish								1											1	1
	<i>Diodon nictemerus</i>	Globefish						1		1									1			3
	<i>Eocallionymus papilio</i>	Painted stinkfish				1	1															2
	<i>Forsterygion gymnotum</i>	Estuarine threefin*		9	15	20		39	9	1		6	12	7								118
	<i>Forsterygion varium</i>	Many-rayed threefin*	23	130	34	2	72	26	70	74	20	40			6	38		233	18	3	8	797
	<i>Heteroclinus johnstoni</i>	Johnstons weedfish												2								2
	<i>Heteroclinus tristis</i>	Longnose weedfish					1					1										2
	<i>Hippocampus abdominalis</i>	Bigbelly seahorse										1										1
	<i>Neosebastes scorpaenoides</i>	Common gurnard perch																	1			1
	<i>Nesogobius hinsbyi</i>	Hinsbys goby	7	1	1	8	5	1		1												24
	<i>Nesogobius maccullochi</i>	Girdled goby	4																			4
	<i>Parablennius tasmanianus</i>	Tasmanian blenny	2	1	5		2	1														11
	<i>Pempheris multiradiata</i>	Common bullseye		1	1				20	14	35				55			1		39	74	240
	<i>Pseudaphritis urvillii</i>	Freshwater flathead			1																	1
	<i>Pseudophycis bachus</i>	Rod cod									1											1
	<i>Scorpaena papillosa</i>	Southern rock cod									1		1			2				1	7	12
	<i>Trachichthys australis</i>	Southern roughy									13											13
	<i>Trinorfolkia clarkei</i>	Clark's threefin	3	1			5	1		1	6	3	3	1	1	1		1		1	5	32
	<i>Vincentia conspersa</i>	Southern cardinalfish	1	12	3					1											2	19
Crustaceans	<i>Carcinus maenas</i>	Green crab*				1																1
	<i>Halicarcinus rostratus</i>	Beaked flat spider crab	2			1																3
	<i>Jasus edwardsii</i>	Southern rock lobster								16	24				11	2	5	14		54	42	168
	<i>Metacarcinus novaezelandiae</i>	Piecrust crab*				6																6
	<i>Nectocarcinus integrifrons</i>	Rough rock crab				1																1
	<i>Pagurixus handrecki</i>	Handrecks hermit crab									3	8										11
	<i>Petrolisthes elongatus</i>	New Zealand porcelain crab*			5	2																7
	<i>Plagusia chabrus</i>	Red bait crab																1	7	1	2	11
	<i>Strigopagurus strigimanus</i>	Stridulating hermit crab																1				1
	<i>Phlyctenactis tuberculosa</i>	Swimming anemone																		1		1
	<i>Asterias amurensis</i>	Northern Pacific seastar*	97	117	130	267	200	429	5	95	116	464	19					3				1942
Cnidarians	<i>Astrostele scabra</i>	Seven-armed starfish																2		1		3
	<i>Australostichopus mollis</i>	Sea cucumber										1	1	5				1				8
	<i>Cenolia trichoptera</i>	Feather star								3	1				30	17	1	35	35		735	857
	<i>Comanthus tasmaniae</i>	Tasmanian feather star													1				1	2	6	10
	<i>Coscinasterias muricata</i>	Eleven-armed seastar								1	1			19			32		39	5		97
	<i>Goniocidaris tubaria</i>	Thorny sea urchin																	3			3
	<i>Helicoidaris erythrogramma</i>	Purple urchin			2	2	101	49	81	128	275	386	93	21	46	8	10	51	2	6		1261
	<i>Holopneustes inflatus</i>	Seagrass sea urchin						2														2
	<i>Meridiastra calcar</i>	Eight-armed seastar														208		220	53			481
	<i>Neothyonidium spp.</i>	Unidentified sea cucumber																	89			89
Echinoderms	<i>Patiriella regularis</i>	New Zealand common cushion star*	667	859	408	495	629	1084	797		429	471	30					1				5870
	<i>Tosia magnifica</i>	Magnificent biscuit star								2		1	1	1				1				6
	<i>Uniophora granifera</i>	Granular seastar																	1			1
	<i>Argobuccinum pustulosum</i>	Pustular triton											1					1				2
	<i>Cabestana spengleri</i>	Spengler's trumpet														8	1	2	1			12
	<i>Charonia lampas</i>	Pink lady																		1		1
	<i>Dicathais orbita</i>	White rock shell														3						3
	<i>Fusinus novaehollandiae</i>	New holland spindle										4										4
	<i>Haliotis rubra</i>	Blacklip abalone												1	15	18	1	11	16	5		67
	<i>Lunella undulata</i>	Common warrener												124	151							275
	<i>Maoricolpus roseus</i>	New Zealand screwshell									1											1
	<i>Penion mandarinus</i>	Mandarin penion									1		3						1			5
	<i>Pleuroploca australasia</i>	Tulip shell									1		1									2
	<i>Polycera capensis</i>	Crowned nudibranch		2	8			1		1		7			2	1		2	2			26
	<i>Ranella australasia</i>	Australian rock whelk												2		26		11		1		40
	<i>Scutus antipodes</i>	Elephant snail									3					3						6
Species richness (total species per site)			2	11	13	11	11	8	10	12	18	11	15	12	9	15	6	21	15	15	12	
Total species richness			59																			

Appendix 12: Average algal percentage covers per site, recorded during standard surveys in February-April 2010, * = introduced species.

METHOD 3, 2009/2010		1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
Groupings	Site Number of transects Species / Depth	2	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4
		1	4	4	3	2	2	2	4	4	2	3	4	4	4	5	4	5	5	5	5
Brown algae	<i>Acrocarpia paniculata</i>	0	0	0	0	0	0	0	0	0	0	0	25.4	22.7	2.8	4.1	2.3	0.5	6.9	7.5	72.2
	<i>Carpoglossum confluens</i>	0	0	0	0	0	0	0	0	0	0	0	0	11.9	7.3	48.5	13.4	21	7.7	20.2	130
	<i>Caulocystis cephalornithos</i>	0	0	0	0	0	0	0	0	0	0	1.5	2	0	0	0	0	0	0	0	3.5
	<i>Caulocystis uvifera</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0	0	0.3
	<i>Cladostephus spongiosus</i>	0	0	0	0	0	0	0	0	0	0	0.5	2.8	4.5	0.9	0.3	2.7	0	0	0	11.7
	<i>Cystophora moniliformis</i>	0	0	0	0	0	0	0	0	0	0	0	0.6	2.2	0	0.6	3.5	0	0	0	6.9
	<i>Dictyopteris muelleri</i>	0	0	0	0	0	0	0	0	0	0	0	2	3.6	11.3	0	6.4	0	0	0	23.3
	<i>Dictyota dichotoma</i>	0	0	0	0	0	0	0	0	1.2	0	0.4	0	0	0	4.3	0	0	0	0	5.9
	<i>Ecklonia radiata</i>	0	0	0	0	0	0	0	9.2	5.4	0	0	28.2	18.8	2	32	24.6	40.6	23.5	73.6	257.9
	<i>Halopteris paniculata</i>	0	0	0	0	0	0	0	1.2	0	0	0	0	0	0.1	0	0	0	0	0	1.3
	<i>Lessonia corrugata</i>	0	0	0	0	0	0	0	0	0	0	0	0	17.3	43.7	0	21	0.9	14.1	0.4	97.4
	<i>Macrocystis pyrifera</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	2	10.9	0.8	1.4	15.1
	<i>Perithalia caudata</i>	0	0	0	0	0	0	0	0	0	0	0	0	1.2	0	0	0	0	0	0	1.2
	<i>Phaeophyceae spp.</i>	0	0	0	0	0	0	0	1	0	0	0	0	0	2.5	0	0	0	0	0	3.5
	<i>Sargassum fallax</i>	0	0	0	0	0	0	0	0	10.2	0	10	3.7	6.4	0.9	1.6	5.3	1.9	1.2	0	41.2
	<i>Sargassum spp.</i>	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0	0	0	0	0	0	0.2
	<i>Sargassum verruculosum</i>	0	0	0	0	0	0	0	0	0	0	0	1.6	3.2	0	0	0.5	0.2	0	0	5.5
	<i>Sporochnus comosus</i>	0	0	0	0	0	0	0	3.2	0	7.9	0	0	1	0	0	0	0	0	0	12.1
	<i>Undaria pinnatifida</i>	0	0	0	0	0	0	0	0	0	0	0	3.6	1.9	1.8	0	0	0.2	0	0	7.5
	<i>Xiphophora gladiata</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0.3
	<i>Zonaria spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0.2	0	0	1
Red algae	<i>Aeodes nitidissima</i>	0	0	0	0	0	10.6	1.2	2.8	0.2	0	0	0	0	0	0	0	0	0	0	14.8
	<i>Areschougia spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	1.8	0	0	0	0	0	0	1.8
	<i>Ballia callitricha</i>	0	0	0	0	0	0	0	0	1	0	0	0	2	0.9	0	1.8	0.8	0	2.6	9.1
	<i>Callophyllis rangiferina</i>	0	0	0	0	0	0	0	0	0	0	0	0	7.7	3.1	2.4	5.3	1.9	6.7	0.8	27.9
	<i>Ceramium spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1	0	1.6	0	2.6
	<i>Champia viridis</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.7	0.4	1.1
	<i>Corallinales spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0.6	11.3	12.7	10.1	6.7	26.9	6.3	13.8	88.4
	<i>Delisea pulchra</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.4	0	1.4
	<i>Echinothamnion hystrix</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0.7	0	0.8	0	0	0	1.5
	<i>Euptilota articulata</i>	0	0	0	0	0	0	0	0	0	0	0	0	5.3	0	0	0.1	3.1	0.4	0	8.9
	<i>Gelidium pusillum</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	3.5	1	0	0	4.5
	<i>Gigartina muelleriana</i>	0	0	0	0	0	0	0	0	0	0	0	0.6	0	0	0	0	0	0	0	0.6
	<i>Gracilaria ramulosa</i>	14.6	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	14.6
	<i>Grateloupia turuturu</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0.2
	<i>Halymenia spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0	0	0.3
	<i>Hemineura frondosa</i>	0	0	0	0	0	0	0	0	0	0	0	0.4	2.3	1.4	1	6.8	1.2	9.1	0	22.2
	<i>Hypnea spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	3.6	0	0	0	0	0	0	3.6
	<i>Laurencia spp.</i>	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0	0.7	0	1.1
	<i>Lenormandia marginata</i>	0	0	0	0	0	0	0	0	0	0	0	0	2.9	0.6	3.4	0	3.4	10.3	9.6	30.2
	<i>Mastophoropsis canaliculata</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	2.6	0	0	0	0	0	2.6
	<i>Peyssonnelia novaehollandiae</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0.5	0.9
	<i>Peyssonnelia spp.</i>	0	0	0	0	0	0	0	13	13	0	63.8	44.4	21.1	16.3	49.6	13.9	23.7	8.2	9.1	276.1
	<i>Phacelocarpus alatus</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.9	1.6	3.5
	<i>Phacelocarpus peperocarpus</i>	0	0	0	0	0	0	0	0	0	0	0	0	4.6	0.2	0	0	5.1	15.6	0.9	26.4
	<i>Plocamium angustum</i>	0	0	0	0	0	0	0	0	0	0	0	1.8	25.8	4.8	0	3.6	2.5	5.2	2.6	46.3
	<i>Plocamium cartilagineum</i>	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0	0	0	0.2
	<i>Plocamium dilatatum</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.6	2.6	6.5	10.7
	<i>Pollexfenia lobata</i>	0	0	0	0	0	0	0	0	0	0	0	5.4	9.9	14.8	7.7	7.3	1.5	0.7	0	47.3
	<i>Polyopes constrictus</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1	1	0	0	2
	<i>Pterocladia capillacea</i>	0	0	0	0	0	0	0	0	0	0	0	0.6	0.5	3.9	0	0	0	0	0	5
	<i>Ptilonia australasica</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0.9	1.2
	<i>Rhodophyta spp.</i>	24.6	1.6	17.6	2.4	0	0.6	8.6	3.2	0	6.8	0.4	0.4	0	2.7	0.6	14.1	16.3	3.6	2.4	105.9
	<i>Rhodymenia spp.</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	3.5	0	0	0	0.4	6	9.9
	<i>Sonderophycus coriaceus</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.9	0.9
	<i>Stenogramma interruptum</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0	2.4	2.9
	<i>Thamnoclonium dichotomum</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0.5
	<i>Tsengia feredayae</i>	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0.3	0.3	0	0	0	1

Appendix 12 cont: Average algal percentage covers per site, recorded during standard surveys in February-April 2010, * = introduced species.

METHOD 3, 2009/2010	Site	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total	
	Number of transects	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4		
Groupings	Species / Depth	1	4	4	3	2	2	4	4	5	2	3	4	4	5	4	5	5	5	5		
Green algae	<i>Bryopsis</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0.2	
	<i>Caulerpa geminata</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.5	1.5	
	<i>Caulerpa longifolia</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.6	
	<i>Caulerpa simpliciuscula</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	5.6	0.8	0	0.3	0	1	0.2	7.9
	<i>Caulerpa trifaria</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	1.2	0	0	0.7	0	2.6	4.3	8.8
	<i>Chlorophyta</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0.4
	<i>Codium</i> spp.	0	0	0	0	0	0	0.6	1.6	0.2	7.4	1.3	0	0	0	0	0	0	0	0	0	11.1
	<i>Ulva</i> spp.	0.6	0	0	0	0	0	0	0	1.6	0	0	3.2	0	1.8	0.5	0	0.4	0	0	0	8.1
Algae	<i>Algae</i> spp.	37.8	0	0	0	0	0	0	0	4.4	0	6.9	33	9.6	12.7	15.4	6.8	9.5	0	1	137.1	
Sessile Invertebrates	<i>Animalia</i> spp.	0	85.8	84.2	83.6	92.6	66	71.8	58.2	0	72.6	0	0	0	0	0	1.6	0	1	3	620.4	
	<i>Anthozoa</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0.9	0	0.5	0.4	0	0	1.8	
	<i>Ascidacea</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0.9	0	0	0	0	0	0.9	
	<i>Bryozoa</i> spp.	0	0	0	0	0	0	0	0.8	0	0	0	0	0.4	0	0	0	0	1.1	5	7.3	
	<i>Corynactis australis</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0.4	
	<i>Crassostrea gigas</i>	2.4	1.4	0.8	0	0	1.4	0	0	0	0	0	0	0	0	0	0	0	0	0	0	6
	<i>Culicia tenella</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0.5	1	
	<i>Erythropodium</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	4	0	4	
	<i>Galeolaria</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0.2	
	<i>Herdmania grandis</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.6	0.6	
	<i>Hydroida</i> spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0.2	
	<i>Mytilus galloprovincialis</i>	0.6	0	0	0	0	0	0	0	0	0	0	0	0	0	2.2	4	0	0	0	6.8	
	<i>Ostrea angasi</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0.4	12.6	1.5	0	0	0	14.5	
	<i>Oulactis mucosa</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0.2	
	<i>Phlyctenanthus australis</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0	0	0	0.8	
	<i>Porifera</i> spp.	0	0	0	0	0	0	0	0	0.2	0	0.2	0.6	24.6	17.4	6.2	26.2	14	53	51.1	193.5	
	<i>Pyura australis</i>	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.1	0	0.1	
	<i>Pyura gibbosa</i>	0	0	0	0	0	0	0	0	0	0	0	0.2	0	3.6	0	3.7	0	0	0	7.5	
Other	Bare rock	0	0	0	0	0	0	0	0	0	0	1.5	0	0	0	0	0	0	1.8	0	3.3	
	Cobble	34.6	88.2	10	6.4	51	39.8	21.8	0	0	2	10.8	0	0	0	0	0	11.5	0	0	276.1	
	Gravel	31	11.2	0	0	0	34	26	0	0	0	9	0	18.7	0	0	10.8	0	8.2	0	148.9	
	Sand	0	0	0	0	0	0	0	25	0	0	7.2	19.4	0	10	9.7	6.9	4.4	0	2.1	84.7	
	Shell	0	0	0	0	0	0	0	0	0	23.8	0	0	0	0	1.8	0	0	0	0	25.6	
	Silt on reef	32	0.6	11.6	0	8	0	0	0	0	0	0	0	0	0	0	0	0	0	0	52.2	
	Turf/sand/sediment matrix	0	0	0	0	0	0	0	0	69.8	0	0	0	0	16.7	0	0	0	1	0	87.5	
	<i>Heterozostera tasmanica</i>	21.6	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	21.6	
Species richness (total species per site)		10	6	5	3	3	6	6	10	12	5	15	22	32	38	22	41	30	39	34		
Total species richness		93																				

Appendix 13: Average algal percentage covers per site, recorded during standard surveys in 2023-2024, * = introduced species.

METHOD 3, 2023/2024	Site Number of transects Species / Depth	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
		2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	
		1	4	4	3	2	2	2	4	4	2	3	4	3	5	4	4	5	5	5	
Brown algae	Acrocarpia paniculata	0	0	0	0	0	0	0	0	0	0	0	8.4	9	0	1	0.2	1.1	10.9	3.4	34
	Carpoglossum confluens	0	0	0	0	0	0	0	0	0	0	0	2.8	3.2	6.6	18.3	5.9	29.8	31.7	18.6	116.9
	Caulocystis uvifera	0	0	0	0	0	0	0	0	0	0	4.3	0	0	0	0	0	0	0	0	4.3
	Cladostephus spongiosus	0	0	0	0	0	0	0	0	0	0	0	1.6	1.8	0	0	0	0	0	0	3.4
	Colpomenia sinuosa	0	0	0	0	0	0	0	0	0	0	0.1	0	0	0	0	0	0	0	0	0.1
	Cystophora moniliformis	0	0	0	0	0	0	0	0	0	0	2.1	2	5.7	0.5	0.4	2.2	0	0	0	12.9
	Dictyopteris muelleri	0	0	0	0	0	0	0	0	2.8	0	7	0	1.4	0	0.2	0.2	0.7	0	0.1	12.4
	Dictyota nigricans	0	0	0	0	0	0	0	0	0	0	0	0	0	2.1	0	1.4	0	0	0	3.5
	Ecklonia radiata	0	0	0	0	0	0	27.2	72.2	43.2	0	23.3	56.4	48.7	39.3	36.7	30.7	37.8	10	71.4	496.9
	Halopteris paniculata	0	0	0	0	0	0	0	0	0	0	7	0	9.2	0	0	0	0.2	0	0	16.4
	Lessonia corrugata	0	0	0	0	0	0	0	0	0	0	0	0	0	14.7	0.2	3.2	0	2	0	20.1
	Macrocystis pyrifera	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	4.1	0.5	6.2	0	10.8
	Perithalia caudata	0	0	0	0	0	0	0	0	0	0	0	0	6.1	0	0	0	1.9	1	0	9
	Phaeophyceae spp.	0	0	0	0	0	0	0.8	0	0	0.8	0.8	0	0	0	0	0	0	0	0.6	3
	Sargassum fallax	0	0	0	0	0	0	0	5.8	2	1.2	0.5	0.8	2.1	5	1.8	3.1	0	0.3	0.5	23.1
	Sargassum vestitum	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.4	0.8	0.9	3.1
	Undaria pinnatifida	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0.5	0	0.3	0	0	1
	Xiphophora gladiata	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0.8
	Zonaria spp.	0	0	0	0	0	0	0	0.6	0	0	0.7	0	0.4	2.1	0	0.1	1.1	0	0.4	5.4
Red algae	Aeodes nitidissima	0	0	0	0	0	1.2	0.2	0	0	0	0	0	0	0	0	0	0	0	0	1.4
	Areschougia congesta	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0	0	0.2	0.6
	Ballia callitricha	0	0	0	0	0	0	0	0	2	0	0	0	1	0.4	1.8	0.6	0	0	0.6	6.4
	Ballia spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	4	0.7	1.9	6.6
	Callophyllis rangiferina	0	0	0	0	0	0	0	0	0	0	0	0	0.1	2.1	12.8	0	6.7	10.8	0	32.5
	Champia spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0	0.8
	Champia viridis	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0.3
	Corallinales spp.	0	0	0	0	0	0	0	0.2	0.4	0	0.5	2.2	31.2	3.7	13	3	40.9	30.5	11.7	137.3
	Craspedocarpus ramentaceus	0	0	0	3.4	0	0.4	3.2	0.2	0	0	0	0	0	0	0	0	0	0	0	7.2
	Dasythamniella plumigera	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.1	0.1
	Delisea pulchra	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0.3
	Echinothamnion hystrix	0	0	0	0	0	0	0	0	0	0	0.3	0	0	0	0	0	2.2	0	0	2.5
	Euptilota articulata	0	0	0	0	0	0	0	0	0	0	0	0	0.7	0	0	0	0	1.2	0.2	2.1
	Gigartina muelleriana	0	0	0	0	0	0	0	0	0	0	0	0	0	1.4	3.8	0	0	0	0	5.2
	Gracilaria cliftonii	0	0	0	0	0	0	0	0	0	0	0	0.8	0	0	0	0	0	0	0	0.8
	Grateloupia turuturu	0	0	0	0	0	0	0	0.4	0.4	0	1.4	0	0	0.5	0.2	0	0.3	0	0	3.2
	Haliptilon roseum	0	0	0	0	0	0	0	0	0	0	0	1.4	0	0	0	0	0	1.3	0	2.7
	Hemineura frondosa	0	0	0	0	0	0	0	0	0	0	0	2.6	0	0.5	0.6	0.1	0.8	4	0	8.6
	Kallymenia spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1	0	0	1
	Laurencia spp.	0	0	0	0	0	0	0	0.8	1.2	0	11	0	0.7	0	2.6	0	0.2	1.6	0.3	18.4
	Lenormandia marginata	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0.9	0.6	5.7	5.7	7	20.1
	Mychodea aciculare	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0.2
	Peyssonnelia spp.	0	0	0	0	0	0	3.6	52.2	21	0.6	23	21.8	21.3	5.2	17.9	9	22.8	22.8	14.3	235.5
	Phacelocarpus apodus	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	2	0.5	0	2.5
	Phacelocarpus peperocarpus	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	2.8	9.5	0.9	13.6
	Phacelocarpus sessilis	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.1	0.1
	Plocamium angustum	0	0	0	0	0	0	0	0	0	0	0	0	2.3	0.2	0	0.1	8.9	7.4	1.1	20
	Plocamium costatum	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.6	9.7	5.6	15.9
	Plocamium dilatatum	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.5	0	1.5
	Plocamium patagiatum	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0.5
	Pollexfenia lobata	0	0	0	0	0	0	0	0	0	0	0.6	0	2.8	2.9	6.6	0.9	0.6	3.8	0	18.2
	Polyopes constrictus	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.2	0.2	1.4
	Ptilonia australasica	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.7	0.7
	Rhodophyta spp.	30	0	0	10.2	2.2	19.4	6	0	12.4	25	2.2	0.6	0	0	0	0.9	0	3	0.5	112.4
	Rhodomenia spp.	0	35	85.8	14	0	0.8	50.2	17	27	2.6	0.9	1.4	0.8	3.5	1	5.5	1	3.2	6.3	256
	Sonderophycus coriaceus	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	1.3	1.8
	Thamnoclonium dichotomum	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.1	1.1
	Tsengia feredayae	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.2	0	0	1.2

Appendix 133 cont: Average algal percentage covers per site, recorded during standard surveys in 2023-2024, * = introduced species.

METHOD 3, 2023/2024	Site	1	3	4	5	6	7	9	13	14	15	17	18	19	20	22	24	25	26	27	Total
	Number of transects	2	2	2	2	2	2	2	2	2	2	4	2	4	4	4	4	4	4	4	
	Species / Depth	1	4	4	3	2	2	2	4	4	2	3	4	3	5	4	4	5	5	5	
Green algae	Apjohnia laetevirens	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0	0	0	0	0.5
	Caulerpa geminata	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.7	2.3	3
	Caulerpa longifolia	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0	0	0	0	0.2
	Caulerpa simpliciuscula	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.6	0	1.6
	Caulerpa trifaria	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0	0.8
	Chaetomorpha coliformis	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0	0	0.4	0.6	0.1	1.4
	Chlorophyta spp.	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0	0	0	0	0.6	0	1
	Cladophora spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.8	0.5	0.1	2.4
	Codium spp.	0	0	0	3.4	31.2	5	3.2	0	1.2	64.4	0.8	0	0.6	0	0	0.2	0.1	0	0	110.1
	Ulva spp.	0	0	0	0	0	0	3	0	2	0	0.4	1.2	0.2	0	1.9	8.9	1.3	0	0	18.9
Algae	Algae spp.	30	0	0	50.2	82.2	0	7.4	0	46.6	0	20.2	25.6	13.5	16	8	46	2.5	12.9	10.2	371.3
Sessile Invertebrates	Anthothoe albocincta	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0.5
	Anthozoa spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.3	0	0.2	0	0.5
	Ascidacea spp.	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0.3	1	1	2.7
	Bryozoa spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	1.6	10	12.1
	Corynactis spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0.2
	Crassostrea gigas	13.8	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	13.8
	Culicia spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0.2
	Galeolaria caespitosa	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0	0	0	0	0	0	0.4
	Herdmania grandis	0	0	0	0	0	0	0	0	0	0	0	0	0	0.4	0	0	0	0	0	0.4
	Hydroida spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.8	0	0.5	0.5	1.8
	Magallana gigas	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.9	0	0	0	1.9
	Mytilid spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.6	0	0	1.6
	Mytilus galloprovincialis	0.8	0	1.2	25.6	0	8.2	0	0	0	0	0.2	7.4	0	50.4	30.4	18.7	2.2	0	0	145.1
	Ostrea angasi	0	6.6	0	8.6	0	14.2	0	3.6	0.4	0	0.2	0	0	0	1.3	0	0	0	0	34.9
	Ostrea spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.2	0	0	0.2
	Porifera spp.	0	2.4	7.4	0	3.4	3	1	5.8	5	0	0.6	6	6.6	10.2	11.8	5.3	7.8	16.4	38.7	131.4
	Pyura australis	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1.9	0	0.6	0.2	2.7
	Pyura gibbosa	0	0	0	0	0	0	0	0	0	0	0	0	0.5	1.5	0	0	0	0	0	2
	Sessilia spp.	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.1	0	0.1
Other	Bare rock	0	0	0	0	1.6	0	21	1.8	0	9.6	17	1	1	1.1	0	1.2	0.8	0.7	0.4	57.2
	Cobble	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0.5	0.5
	Gravel	0	0	0	0	2.6	0	0	3.4	0	0	0	0	2.8	1.1	0	2.7	1.2	0	1.8	15.6
	Sand	0	0	0	0	0	0	4	1	0	7.8	30.6	27.4	10.2	4.2	7.9	3.6	7.8	0	0.4	104.9
	Shell	0	10.4	2.4	0	0	0	0	0	0	14	0	0	0	2	0	0	5.1	0	0	33.9
	Silt on reef	32	7.4	22	4.8	0	0	2	0	0	0	0	0	0	0	0	0	0	0	0	68.2
	Turf/sand/sediment matrix	38	54	0	0	0	75.2	6	10.2	0	0	0	0	0	0	0	0	7.9	6.6	7.7	205.6
	Heterozostera nigricaulis	30	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	30
Species richness (total species per site)		7	6	5	8	6	9	15	17	15	9	26	19	32	27	26	31	44	45	47	
Total species richness		95																			

Appendix 14: Rank abundances of algae and plant species recorded during timed surveys (sites 1–9) in February–March 2010. Depth. Depth categories: -1 = intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand. * = introduced species.

2009/2010	Site	1	1	2	2	3	3	4	4	5	5	6	6	7	7	8	8	9	9
	Depth category	-1	1	-1	0	-1	1	-1	1	-1	0	1	0	1	0	1	0	-1	0
	Habitat	R	R	R	E	R	R	R	R	R	R	R	R	R	R	R	R	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2
Brown algae	<i>Acrocarpia paniculata</i>																		
	<i>Bellotia eriophorum</i>																		
	<i>Carpoglossum confluens</i>																		
	<i>Caulocystis cephalornithos</i>																		
	<i>Caulocystis uvifera</i>																		
	<i>Cladosiphon filum</i>																		
	<i>Cladostephus spongiosus</i>																		
	<i>Colpomenia sinuosa</i>																		
	<i>Cystophora monilifera</i>																		
	<i>Cystophora moniliformis</i>																		
	<i>Cystophora retorta</i>																		
	<i>Cystophora retroflexa</i>																		
	<i>Dictyopteris muelleri</i>																		
	<i>Dictyota dichotoma</i>																		
	<i>Dictyota spp.</i>																		
	<i>Distromium flabellatum</i>																		
	<i>Ecklonia radiata</i>																		
	<i>Halopteris paniculata</i>																		
	<i>Hormostira banksii</i>																		
	<i>Lessonia corrugata</i>																		
	<i>Macrocystis pyrifera</i>																		
	<i>Phyllospora comosa</i>																		
	<i>Sargassum fallax</i>																		
	<i>Sargassum spp.</i>																		
	<i>Sargassum verruculosum</i>																		
	<i>Sargassum vestitum</i>																		
	<i>Sporochnus comosus</i>																		
	<i>Undaria pinnatifida</i>																		
	Unidentified algae (brown turf)																		
	Unidentified algae (filamentous brown)																		
	<i>Zonaria turneriana/angustata</i>																		
Green algae	<i>Caulerpa brownii</i>																		
	<i>Caulerpa geminata</i>																		
	<i>Caulerpa longifolia</i>																		
	<i>Caulerpa simpliciuscula</i>																		
	<i>Caulerpa trifaria</i>																		
	<i>Chaetomorpha coliformis</i>																		
	<i>Cladophora spp.</i>																		
	<i>Codium duthieae</i>																		
	<i>Codium spongiosum</i>																		
	<i>Codium spp.</i>																		
	<i>Ulva spp.</i>																		
	Unidentified algae (filamentous green)																		
Plants	<i>Sarcocornia quinqueflora</i>																		
Red algae	<i>Aeodes nitidissima</i>																		
	<i>Dasya spp.</i>																		
	<i>Gigartina recurva</i>																		
	<i>Gracilaria ramulosa</i>																		
	<i>Gracilaria spp.</i>																		
	<i>Grateloupia turuturu</i>																		
	Unidentified algae (filamentous red)																		
	Unidentified algae (red)																		
Seagrasses	<i>Heterozostera tasmanica</i>																		
Unclassified algae	Unidentified algae																		
	Unidentified algae (turf)																		
Species richness (total species per depth & habitat)		4	4	4	0	4	4	3	2	2	2	1	0	2	2	1	0	1	0
Total species richness (total species per site)		7		6		2		2		2		4		2		4		6	

Appendix 14 cont.: Rank abundances of algae and plant species recorded during timed surveys (sites 19-24) in February-March 2010. Depth. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand. * = introduced species.

[illegible]

Appendix 15: Rank abundances of algae and plant species recorded during timed surveys (sites 1-9) in 2024. Depth. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand. * = introduced species.

[illegible]

Appendix 15 cont.: Rank abundances of algae and plant species recorded during timed surveys (sites 10-18) in 2024. Depth. Depth categories: - 1 = intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand. * = introduced species.

2024	Site	10	11	12	13	14	15	16	17	18																			
	Depth category	-1	0	2	2	-1	0	1	2	3	-1	0	1	2	3	-1	0	1	2	3	-1	0	1	2	3				
	Habitat	R	R	S	S	R	R	R	E	S	R	R	R	S	E	R	R	R	S	E	R	R	R	R	E				
	Species / Replicate search	1	2	1	2	1	2	1	2	3	1	1	2	3	1	2	1	2	3	1	1	1	2	1	2				
Brown algae	<i>Acrocarpia paniculata</i>																												
	<i>Carpoglossum confluens</i>																												
	<i>Caulocystis uvifera</i>																												
	<i>Cladostephus spongiosus</i>																												
	<i>Colpomenia sinuosa</i>																												
	<i>Cystophora cf. monilifera</i>																												
	<i>Cystophora grevillei</i>																												
	<i>Cystophora moniliformis</i>																												
	<i>Cystophora subfarcinata</i>																												
	<i>Dictyopteris muelleri</i>																												
	<i>Ecklonia radiata</i>																												
	<i>Glossophora nigricans</i>																												
	<i>Halopteris paniculata</i>																												
	<i>Halopteris</i> spp.																												
	<i>Hormosira banksii</i>																												
	<i>Lessonia corrugata</i>																												
	<i>Macrocystis pyrifera</i>																												
	<i>Myriodesma integrifolium</i>																												
	<i>Perithalia caudata</i>																												
	<i>Sargassum fallax</i>																												
	<i>Sargassum</i> spp.																												
	<i>Sargassum verruculosum</i>																												
	<i>Sporochnus comosus</i>																												
	<i>Undaria pinnatifida</i>																												
	<i>Zonaria angustata</i>																												
	<i>Zonaria turneriana/angustata</i>																												
Green algae	<i>Apjohnia laetevirens</i>																												
	<i>Caulerpa longifolia</i>																												
	<i>Caulerpa simpliciuscula</i>																												
	<i>Caulerpa</i> spp.																												
	<i>Caulerpa trifaria</i>																												
	<i>Codium spongiosum</i>																												
Red algae	<i>Codium</i> spp.																												
	<i>Ulva</i> spp.																												
	<i>Aeodes nitidissima</i>																												
	<i>Craspedocarpus ramentaceus</i>																												
	<i>Grateloupia turuturu</i>																												
	<i>Thamnoclonium dichotomum</i>																												
Seagrasses	Unidentified algae (red)																												
	<i>Heterozostera nigricaulis</i>																												
	<i>Heterozostera tasmanica</i>																												
Species richness (total species per depth & habitat)		3	2	6	0	0	0	1	0	3	3	3	5	7	6	2	6	1	0	3	1	2	3	6	2	0	1	0	7
Total species richness (total species per site)		6				0		1		10				12						1		7		10			24	7	

Appendix 15 cont.: Rank abundances of algae and plant species recorded during timed surveys (sites 19-24) in 2024. Depth. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand. * = introduced species.

2024	Site	19	0	1	2	20	0	1	2	21	22	0	1	2	23	24	0	1	2
	Depth category	-1				-1				2	-1				2	-1			
	Habitat	R	R	R	R	R	R	R	S	S	R	R	R	S	S	R	R	R	S
	Species / Replicate search	1	2	1	2	3	1	2	1	2	1	2	1	2	3	1	2	1	2
Brown algae	<i>Acrocarpia paniculata</i>		3				5		3				4		4		3		4
	<i>Carpoglossum confluens</i>						4		5	4					5			3	3
	<i>Caulocystis uvifera</i>																		
	<i>Cladostephus spongiosus</i>																		
	<i>Colpomenia sinuosa</i>																		
	<i>Cystophora cf. monilifera</i>																		
	<i>Cystophora grevillei</i>																		
	<i>Cystophora moniliformis</i>						3		2	3							5	4	4
	<i>Cystophora subfarcinata</i>																2	3	
	<i>Dictyopteris muelleri</i>						4		3										
	<i>Ecklonia radiata</i>		4				5	4	5	5	2	2	5	4	5	5	4	2	5
	<i>Glossophora nigricans</i>																		
	<i>Halopteris paniculata</i>						4												
	<i>Halopteris spp.</i>											2							
	<i>Hormosira banksii</i>																		
	<i>Lessonia corrugata</i>		3										4		3	3	3	5	3
	<i>Macrocystis pyrifera</i>																	3	3
	<i>Myriodesma integrifolium</i>										2								
	<i>Perithalia caudata</i>						5		4										
	<i>Sargassum fallax</i>		4				3				3		5	4	3		4	2	3
	<i>Sargassum spp.</i>											2							
	<i>Sargassum verruculosum</i>																		
	<i>Sporochnus comosus</i>																		
	<i>Undaria pinnatifida</i>						2				3	3	5	3	3		3	3	3
	<i>Zonaria angustata</i>																		
	<i>Zonaria turneriana/angustata</i>																		
Green algae	<i>Apjohnia laetevirens</i>												3						
	<i>Caulerpa longifolia</i>						3	1	4										
	<i>Caulerpa simpliciuscula</i>								5										
	<i>Caulerpa spp.</i>																		
	<i>Caulerpa trifaria</i>						3		3										
	<i>Codium spongiosum</i>																		
	<i>Codium spp.</i>																		
Red algae	<i>Ulva spp.</i>						3	2			3	5	5	3	3	4	5		
	<i>Aeodes nitidissima</i>																		
	<i>Craspedocarpus ramentaceus</i>																		
	<i>Grateloupia turuturu</i>										2	3	3	4	4		5	4	5
	<i>Thamnoclonium dichotomum</i>																		
Seagrasses	Unidentified algae (red)																		
	<i>Heterozostera nigricaulis</i>																		
	<i>Heterozostera tasmanica</i>																		
Species richness (total species per depth & habitat)		0	0	4	1	1	12	4	9	5	0	5	5	6	5	5	6	6	2
Total species richness (total species per site)							16												

Appendix 16: Rank abundances of invertebrates recorded during timed surveys (sites 1–9) in February–March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	1	1	2	2	3	3	4	4	5	5	6	6	7	7	8	8	9	9	1	1	2	2	3	3
	Depth category	-1	0	1	2	3	4	5	6	7	8	9	1	2	3	4	5	6	7	8	9	1	2	3	4
	Habitat	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	3	4
Barnacles	<i>Catomerus polymerus</i>																								
	<i>Chamaesipho tasmanica</i>																								
Crustaceans	<i>Elminius modestus</i>																								
	<i>Tetractilella purpurascens</i>																								
Crustaceans	Unidentified barnacles	5	5			5	5			5	5			5	5			5	5						
	<i>Bellidilia undecimspinosus</i>																								
Crustaceans	<i>Brachyotus spinosus</i>	2	2																						
	<i>Carcinus maenas</i>					2	3	2		2	2	2	3	2	1	1		4	2	1	1	2	1	3	4
Crustaceans	<i>Cyclograpsus granulatus</i>	2	2			3	4			3	3							3	3						
	<i>Ebalia intermedia</i>																								
Crustaceans	<i>Halicarcinus ovatus</i>	1								1	2							1		2					
	<i>Halicarcinus rostratus</i>																								
Crustaceans	<i>Helograpsus haswellianus</i>	4	4			1				2															
	<i>Jasus edwardsii</i>																								
Crustaceans	<i>Lophopagurus nanus</i>																								
	<i>Metacarcinus novaezelandiae</i>																								
Crustaceans	<i>Munida gregaria</i>																								
	<i>Naxia aurita</i>																								
Crustaceans	<i>Notomithrax ursus</i>																								
	<i>Pagurid spp.</i>																								
Crustaceans	<i>Paguristes frontalis</i>																								
	<i>Pagurixus handrecki</i>																								
Crustaceans	<i>Palaemon dolospina</i>																								
	<i>Palaemon intermedius</i>																								
Crustaceans	<i>Palaemon serenus</i>																								
	<i>Paragrapsus gaimardii</i>	4	4			4	4			4	4			4	4			4	3						
Crustaceans	<i>Paragrapsus laevis</i>																								
	<i>Paragrapsus quadridentatus</i>																								
Crustaceans	<i>Petrolisthes elongatus</i>					3	3	3	3	5	3	3	5	4	4			5	5	5	5	5	5	5	5
	<i>Plagusia chabrus</i>																								
Crustaceans	<i>Strigopagurus strigimanus</i>																								
	Unidentified crab																								
Crustaceans	Unidentified crab (decorator)																								
	Unidentified hermit crab																								
Crustaceans	Unidentified shrimp																								
	<i>Amblypneustes ovum</i>																								
Echinoderms	<i>Amphiura (Amphiura) constricta</i>																								
	<i>Amphiura (Amphiura) elandiformis</i>																								
Echinoderms	<i>Asterias amurensis</i>																								
	<i>Australostichopus mollis</i>																								
Echinoderms	<i>Clarkoma canaliculata</i>																								
	<i>Comanthus tasmaniae</i>																								
Echinoderms	<i>Comanthus trichoptera</i>																								
	<i>Coscinasterias muricata</i>																								
Echinoderms	<i>Goniocidarid spp.</i>																								
	<i>Goniocidarid tubaria</i>																								
Echinoderms	<i>Helicoidarid erythrogramma</i>																								
	<i>Meridiastra calcar</i>																								
Echinoderms	<i>Patriella regularis</i>					1	2	3	5	4	4	4	5	4	5	4	5	5	5	5	5	5	5	5	5
	<i>Petrolisthes elongatus</i>																								
Echinoderms	<i>Tosia australis</i>																								
	<i>Tosia magnifica</i>																								
Echinoderms	<i>Uniophora granifera</i>																								

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 1–9) in February–March 2010. Depth categories: -1= intertidal; 0= 0–0.9m; 1= 1–4.9m; 2= 5–8.9 m; 3= 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site																				
	Depth category																				
	Habitat																				
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	
Molluscs	<i>Afrolittorina praetermissa</i>																				
	<i>Agnewia tritoniformis</i>																				
	<i>Aplysia parvula</i>																				
	<i>Aplysia spp.</i>																				
	<i>Aplysia sydneyensis</i>																				
	<i>Argobuccinum pustulosum</i>																				
	<i>Austrocochlea constricta</i>																				
	<i>Austrocochlea spp.</i>																				
	<i>Austrolittorina unifasciata</i>																				
	<i>Bembicium auratum</i>																				
	<i>Bembicium melanostomum</i>																				
	<i>Bembicium nanum</i>																				
	<i>Brachidontes rostratus</i>																				
	<i>Cabestana spengleri</i>																				
	<i>Cabestana tabulata</i>																				
	<i>Cellana solida</i>																				
	<i>Cellana tramoserica</i>																				
	<i>Chiton glaucus</i>																				
	<i>Chlorodiloma odontis</i>																				
	<i>Clanculus undatus</i>																				
	<i>Crepidina rugosa</i>																				
	<i>Cominella lineolata</i>																				
	<i>Crassostrea gigas</i>	2	2	5	5	3	3	4	3	3	3	5	3	3	5	3	3	4	4	5	
	<i>Dicathais orbita</i>																				
	<i>Diloma concamerata</i>																				
	<i>Haliotis rubra</i>																				
	<i>Ischnochiton australis</i>																				
	<i>Ischnochiton elongatus</i>																				
	<i>Ischnochiton variegatus</i>																				
	<i>Lepsiella vinosa</i>																				
	<i>Maoricolpus roseus</i>																				
	<i>Mytilus galloprovincialis</i>			5	2	5	5	5	5	5	5	5	2	5	5	5	5	5	5	5	
	<i>Notoacmea flammea</i>																				
	<i>Notoacmea petterdi</i>																				
	<i>Octopus tetricus</i>																				
	<i>Ostrea angasi</i>																				
	<i>Patella (Scutellastra) chapmani</i>	3																			
	<i>Patella (Scutellastra) peronii</i>																				
	<i>Patelloida latistrigata</i>																				
	<i>Patelloida victoriana</i>																				
	<i>Pecten fumatus</i>																				
	<i>Phallomedusa solida</i>	3																			
	<i>Plaxiphora albida</i>																				
	<i>Pleurobranchaea maculata</i>																				
	<i>Pleuroploca australasia</i>																				
	<i>Ranella australasia</i>																				
	<i>Sassia verrucosa</i>																				
	<i>Siphonaria diemenensis</i>																				
	<i>Siphonaria funiculata</i>																				
	<i>Siphonaria spp.</i>																				
	<i>Siphonaria tasmanica</i>																				
	<i>Sypharochiton pelliserpentis</i>																				
	<i>Tatea huonensis</i>	5	5																		
	<i>Turbo undulatus</i>																				
	Unidentified chiton																				
	<i>Xenostrobus pulex</i>	5	5			3	3			3	3			3							
	<i>Xenostrobus securis</i>				5																

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 1–9) in February–March 2010. Depth categories: -1 = intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	10	11	12	13	14	15	16	17	18
	Depth category	-1	0	2	2	-1	0	1	2	3
	Habitat	R	R	S	S	R	R	R	S	E
	Species / Replicate search	1	2	1	2	1	2	1	2	1
Barnacles	<i>Catomerus polymerus</i>									
	<i>Chamaesipho tasmanica</i>	5	3							
Crustaceans	<i>Elminius modestus</i>									
	<i>Tetraclitella purpurascens</i>									
Crustaceans	Unidentified barnacles					5	5			
	<i>Bellidilia undecimspinosus</i>							3		
Crustaceans	<i>Brachynotus spinosus</i>									
	<i>Carcinus maenas</i>	2			1					
Crustaceans	<i>Cyclograpsus granulosus</i>					4	4			
	<i>Ebalia intermedia</i>								2	1
Crustaceans	<i>Halicarcinus ovatus</i>		2						1	
	<i>Halicarcinus rostratus</i>			3						2
Crustaceans	<i>Helograpsus haswellianus</i>									
	<i>Jasus edwardsii</i>									
Crustaceans	<i>Lophopagurus nanus</i>							2	1	
	<i>Metacarcinus novaezelandiae</i>									1
Crustaceans	<i>Munida gregaria</i>			2						
	<i>Naxia aurita</i>			2						
Crustaceans	<i>Notomithrax ursus</i>									1
	<i>Pagurid spp.</i>				1					
Crustaceans	<i>Paguristes frontalis</i>									
	<i>Pagurixus handrecki</i>									2
Crustaceans	<i>Palaemon dolospina</i>									
	<i>Palaemon intermedius</i>			3						
Crustaceans	<i>Palaemon seren</i>		3	3						
	<i>Paragrapsus gaimardii</i>	2	2	2		1		3	2	
Crustaceans	<i>Paragrapsus laevis</i>									
	<i>Paragrapsus quadridentatus</i>	1		2						
Crustaceans	<i>Petrolisthes elongatus</i>	5	5	5	5	5	5	5	5	5
	<i>Plagusia chabrus</i>									
Crustaceans	<i>Strigopagurus strigimanus</i>			2						
	Unidentified crab					1				
Crustaceans	Unidentified crab (decorator)			2	2					
	Unidentified hermit crab									
Crustaceans	Unidentified shrimp					2				
	<i>Amblypneustes ovum</i>									
Echinoderms	<i>Amphiura (Amphiura) constricta</i>									
	<i>Amphiura (Amphiura) elandiformis</i>									
Echinoderms	<i>Asterias amurensis</i>	1	3	2	3	4	4	4		
	<i>Australostichopus mollis</i>					4	4	4		
Echinoderms	<i>Clarkoma canaliculata</i>									
	<i>Comanthus tasmaniae</i>									
Echinoderms	<i>Comanthus trichoptera</i>									
	<i>Coscinasterias muricata</i>					1				
Echinoderms	<i>Goniocidaris spp.</i>									
	<i>Goniocidaris tubaria</i>									
Echinoderms	<i>Heliocidaris erythrogramma</i>					3	3	3		3
	<i>Meridiastra calcar</i>									
Echinoderms	<i>Patiriella regularis</i>	5	4	5	5	3	3	5	5	3
	<i>Petrolisthes elongatus</i>									
Echinoderms	<i>Tosia australis</i>									
	<i>Tosia magnifica</i>									
Echinoderms	<i>Uniophora granifera</i>									

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	10	11	12	13	14	15	16	17	18
Depth category		-1	0	2	-1	0	0	2	-1	0
Habitat		R	R	S	R	R	R	R	R	R
Species / Replicate search		1 2 1 2	1 2	1 2	1 2 1 2 3	1 2 3 1 2 3	1 2 1 2 1 2 3	1 2 1 2	1 2 3 1 2	1 2 1 2
Molluscs										
<i>Afrolittorina praetermissa</i>										
<i>Agnestia tritoniformis</i>										
<i>Aplysia parvula</i>										
<i>Aplysia</i> spp.				1						2
<i>Aplysia sydneyensis</i>			1				1			
<i>Argobuccinum pustulosum</i>										
<i>Austrocochlea constricta</i>		2					2		1	2 3
<i>Austrocochlea</i> spp.										1
<i>Austrolittorina unifasciata</i>					4 4		4 4		5 3 3	5
<i>Bembicium auratum</i>	1									
<i>Bembicium melanostomum</i>							2			
<i>Bembicium nanum</i>										
<i>Brachidontes rostratus</i>										5
<i>Cabestana spengleri</i>									1 1	1 2 1
<i>Cabestana tabulata</i>										
<i>Cellana solida</i>									3 2 2	4
<i>Cellana tramoserica</i>										
<i>Chiton glaucus</i>	1				2	2	3 3	1 2	2	3 3 2 1
<i>Chlorodiloma odontis</i>										2 1 2
<i>Clanculus undatus</i>										2
<i>Clypidina rugosa</i>										2
<i>Cominella lineolata</i>										3 4
<i>Crassostrea gigas</i>					3 3	2	4 4	4 4 3	3 3 3 3	2
<i>Dicathais orbita</i>									2 1	4 3 1
<i>Diloma concamerata</i>										2
<i>Haliotis rubra</i>										
<i>Ischnochiton australis</i>					1			1	2 3	1
<i>Ischnochiton elongatus</i>										
<i>Ischnochiton variegatus</i>									1	
<i>Lepsiella vinosa</i>							3			
<i>Maoricolpus roseus</i>						4 2				5 4
<i>Mytilus galloprovincialis</i>	5 5	5 4			5 5 5 5 4	2	5 5 5 5	5 5 4 3	4 5 5 4 5 4	5 5
<i>Notoacmea flammea</i>										2
<i>Notoacmea petterdi</i>									3	
<i>Octopus tetricus</i>										
<i>Ostrea angasi</i>							3	2	1 3 2	1
<i>Patella (Scutellastra) chapmani</i>									2	
<i>Patella (Scutellastra) peronii</i>										
<i>Patelloida latistrigata</i>					2			3		3
<i>Patelloida victoriana</i>										
<i>Pecten fumatus</i>						1				
<i>Phallomedusa solida</i>										
<i>Plaxiphora albida</i>							1			
<i>Pleurobranchaea maculata</i>										1
<i>Pleuroploca australasia</i>										
<i>Ranella australasia</i>										
<i>Sassia verrucosa</i>										
<i>Siphonaria diemenensis</i>					4 4		5 5		5 4 5	5
<i>Siphonaria funiculata</i>					3 2			2 4	5	4
<i>Siphonaria</i> spp.										
<i>Siphonaria tasmanica</i>										
<i>Sypharochiton pelliserpentis</i>			1		3 3	3	3 4	3	4 3 3 4 3	3 3
<i>Tatea huonensis</i>										
<i>Turbo undulatus</i>										3 3
Unidentified chiton								1		2
<i>Xenostrobus pulex</i>										5
<i>Xenostrobus securis</i>										

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	10	11	12	13	14	15	16	17	18																																												
	Depth category	-1	0	2	2	-1	0	1	2	3	-1	0	1	2	-1	0	1	2	-1	0	1	2																																
	Habitat	R	R	S	S	R	R	R	S	E	R	R	R	S	R	R	R	E	R	R	R	E																																
	Species / Replicate search	1	2	1	2	1	2	1	2	3	1	2	3	1	2	3	1	2	1	2	3	1	2																															
Other	<i>Dorvillea australiensis</i>																																																					
	<i>Galeolaria caespitosa</i>																																																					
	<i>Ikeda</i> spp.																																																					
	<i>Myxicola infundibulum</i>			5	4	5	3							5	5	5																																						
	<i>Notoplana australis</i>																																																					
	<i>Paraplanocera</i> spp.																																																					
	<i>Polynoid polychaete</i>																																																					
	<i>Sabellastarte australiensis</i>																																																					
	<i>Sabellastarte</i> spp.			3																																																		
	<i>Sabellid</i> spp.																																																					
	Unidentified flatworm																																																					
	Unidentified invertebrates (tube matting)				3																																																	
	Unidentified polychaete									2																																												
	Unidentified serpulids			4		4																																																
Sessile invertebrates	<i>Actinia tenebrosa</i>																																																					
	<i>Anthothoe albocincta</i>																																																					
	<i>Ascidia sydneyensis</i>			2		1	2																																															
	<i>Ascidella aspersa</i>			5	4		4																																															
	<i>Aulactinia veratra</i>																																																					
	<i>Botrylloides</i> spp.									1																																												
	<i>Ciona intestinalis</i>																																																					
	<i>Cnemidocarpa radicata</i>																																																					
	<i>Corynactis australis</i>																																																					
	<i>Culicia</i> spp.																																																					
	<i>Cyanea rosella</i>																																																					
	<i>Epiactis</i> spp.																																																					
	<i>Eudistoma constrictum</i>																																																					
	<i>Herdmania grandis</i>																																																					
	<i>Herdmania momus</i>																																																					
	<i>Oulactis mucosa</i>																																																					
	<i>Phlyctenanthus australis</i>																																																					
	<i>Pyura australis</i>																																																					
	<i>Pyura gibbosa</i>																																																					
	<i>Pyura stolonifera</i>																																																					
	<i>Sycozoa</i> spp.																																																					
	Unidentified anemone																																																					
	Unidentified ascidians																																																					
	Unidentified sponges																																																					
Species richness (total species per depth & habitat)		9	6	7	10	7	8	4	8	11	12	5	6	4	7	6	7	6	7	9	7	13	16	5	9	6	1	8	8	4	17	17	7	8	8	5	9	14	19	16	7	11	4	5	12	11	21	22	9	6	9	8	6	11
Total species richness (total species per site)		16				12				9				37								38								32				11				42				55												

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 19-24) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 19-24) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 16 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 19-24) in February-March 2010. Depth categories: - 1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	19										20										21					22										23					24																								
	Depth category	-1					0					1					2					-1					0					1					2					2					-1					0					1					2				
	Habitat	R		R			R			R		S	R	R		R			S	E	S	R	R			R		S	E	S	R	R			R		R			S	E																									
	Species / Replicate search	1	2	1	2	3	1	2	1	1	1	2	1	2	3	1	2	1	1	1	2	1	2	3	1	2	1	1	1	2	1	2	1	2	3	1	2	1	1																											
Other	<i>Dorvillea australiensis</i>																																																																	
	<i>Galeolaria caespitosa</i>		3																																																															
	<i>Ikeda spp.</i>																			2	1						1																																							
	<i>Myxicola infundibulum</i>																			5	5						3	2	5	5																																				
	<i>Notoplana australis</i>																																																																	
	<i>Paraplanocera spp.</i>																																																																	
	<i>Polynoid polychaete</i>																																																																	
	<i>Sabellastarte australiensis</i>																																																																	
	<i>Sabellastarte spp.</i>																				2																																													
	<i>Sabellid spp.</i>																																																																	
	Unidentified flatworm																												2																																					
	Unidentified invertebrates (tube matting)																																																																	
	Unidentified polychaete																																																																	
	Unidentified serpulids																												3																																					
Sessile invertebrates	<i>Actinia tenebrosa</i>	2	2																																																															
	<i>Anathoe albocincta</i>				4								3																																																					
	<i>Ascidia sydneiensis</i>																																																																	
	<i>Asciidiella aspersa</i>																																																																	
	<i>Aulactinia veratra</i>													2	3																																																			
	<i>Botrylloides spp.</i>																																																																	
	<i>Ciona intestinalis</i>																																																																	
	<i>Cnemidocarpa radicata</i>																																																																	
	<i>Corynactis australis</i>																																																																	
	<i>Culicia spp.</i>																																																																	
	<i>Cyanea rosella</i>																																																																	
	<i>Epiactis spp.</i>																																																																	
	<i>Eudistoma constrictum</i>																																																																	
	<i>Herdmania grandis</i>																																																																	
	<i>Herdmania momus</i>																																																																	
	<i>Oulactis mucosa</i>	3	3																																																															
	<i>Phlyctenanthus australis</i>																																																																	
	<i>Pyura australis</i>																																																																	
	<i>Pyura gibbosa</i>				4	5	4																																																											
	<i>Pyura stolonifera</i>	3	3												4	5	3	5																																																
	<i>Sycozoa spp.</i>																																																																	
	Unidentified anemone																																																																	
	Unidentified ascidians																																																																	
	Unidentified sponges	0																																																																
	Species richness (total species per depth & habitat)		20	22	4	12	8	3	12	5	0	12	21	8	7	12	14	18	3	12	9	12	16	12	13	11	12	10	10	6	13	5	7	24	29	12	11	8	16	15	2	13																								
	Total species richness (total species per site)		43										44										17					52										9					52																							

Appendix 17: Rank abundances of invertebrates recorded during timed surveys (sites 1-9) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site Depth category Habitat Species / Replicate search	1		2				3				4				5						6						7						8				9													
		-1		1		-1		0		-1		1		-1		1		-1		0		1		2		-1		0		1		2		-1		0		-1		0		1		2							
		R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S	R	S								
		1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2				
Barnacles	<i>Catomerus polymerus</i>																																																		
	<i>Chamaesipho tasmanica</i>																																																		
Crustaceans	<i>Cthamalus antennatus</i>																																																		
	<i>Elminius modestus</i>	3	5			4	5			3	4			3	5			5								3	5																								
	<i>Tetraclitella purpurascens</i>																																																		
	<i>Bellidilia undecimspinoso</i>																																																		
	<i>Carcinus maenas</i>			2		2		1			2															3																									
	<i>Cyclograpsus granulosus</i>			2		3		4			4					4			3						3																										
	<i>Ebalia intermedia</i>																																																		
	<i>Halicarcinus ovatus</i>																								3																										
	<i>Halicarcinus rostratus</i>			3				2			2																																								
	<i>Helograpsus haswellianus</i>	3	4								1																																								
	<i>Jasus edwardsii</i>																																																		
	<i>Metacarcinus novaezelandiae</i>																																																		
	<i>Naxia aurita</i>																																																		
	<i>Nectocarcinus integrifrons</i>																																																		
	<i>Notomithrax ursus</i>																																																		
	<i>Paguristes brevirostris</i>																																																		
	<i>Pagurixus handrecki</i>																																																		
	<i>Pagurus sinuatus</i>																																																		
	<i>Palaemon dolosпина</i>																																																		
	<i>Palaemon intermedius</i>																																																		

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 1-9) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	1	1	2	2	3	3	4	4	5	5	6	6	7	7	8	8	9	9	1	2
	Depth category	-1	0	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18
	Habitat	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2
Molluscs	<i>Afrallitorina praetermissa</i>																				
	<i>Agnewia tritoniformis</i>																				
	<i>Aphelodoris</i> spp.																				
	<i>Aplysia sydneynensis</i>																				
	<i>Austrocochlea constricta</i>																				
	<i>Austrocochlea porcata</i>																				
	<i>Austrolittorina unifasciata</i>																				
	<i>Austrorhynchus rostratus</i>																				
	<i>Bembicium auratum</i>																				
	<i>Bembicium melanostomum</i>																				
	<i>Bembicium nanum</i>																				
	<i>Cabestana spengleri</i>																				
	<i>Cabestana tabulata</i>																				
	<i>Cellana solida</i>																				
	<i>Charonia lampas rubicunda</i>																				
	<i>Chiton glaucus</i>																				
	<i>Chthamalus antennatus</i>																				
	<i>Cominella lineolata</i>																				
	<i>Crassostrea gigas</i>																				
	<i>Dicathais orbita</i>																				
	<i>Diloma concamerata</i>																				
	<i>Electroma georgiana</i>																				
	<i>Equichlamys bifrons</i>																				
	<i>Fusinus noveahollandiae</i>																				
	<i>Gibbomodiolia albicostus</i>																				
	<i>Haliotis rubra</i>																				
	<i>Hastula brazieri</i>																				
	<i>Ischnochiton (Heterozona) cariosus</i>																				
	<i>Ischnochiton australis</i>																				
	<i>Ischnochiton elongatus</i>																				
	<i>Ischnochiton variegatus</i>																				
	<i>Lasaea australis</i>																				
	<i>Limatula strangei</i>																				
	<i>Maoricolpus roseus</i>																				
	<i>Mitra carbonaria</i>																				
	<i>Mitra glabra</i>																				
	<i>Mytilus galloprovincialis</i>																				
	<i>Nassarius nigellus</i>																				
	<i>Notoacmea alta</i>																				
	<i>Notoacmea petterdi</i>																				
	<i>Octopus pallidus</i>																				
	<i>Octopus tetricus</i>																				
	<i>Ostrea angasi</i>																				
	<i>Oxynoe viridis</i>																				
	<i>Patella (Scutellastra) peronii</i>																				
	<i>Plaxiphora albida</i>																				
	<i>Pleurobranchaea maculata</i>																				
	<i>Polycera capensis</i>																				
	<i>Rhyssoplax calliozona</i>																				
	<i>Saccostrea glomerata</i>																				
	<i>Sassia verrucosa</i>																				
	<i>Scutus antipodes</i>																				
	<i>Siphonaria diemenensis</i>																				
	<i>Siphonaria funiculata</i>																				
	<i>Siphonaria tasmanica</i>																				
	<i>Sypharochiton pelliserpentis</i>																				
	<i>Trichomya hirsuta</i>																				
	<i>Turbo undulatus</i>																				
	<i>Xenostrobus pulex</i>																				

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 1-9) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	1				2				3				4				5					6					7					8					9						
	Depth category	-1				0				1				2				3					4					5					6					7						
	Habitat	R				R				R				R				R					R					R					R					R						
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	3	1	2	3	1	2	3	1	2	3	1	2	3	1	2	3	1	2	3	1	2	3	1	2	
Other	<i>Eunice laticeps</i>	2				1				1	2			1	2																													
	<i>Galeolaria</i> spp.																	2																										
	<i>Ikeda</i> spp.																																											
	<i>Lepidonotus melanogrammus</i>																																											
	<i>Myxicola infundibulum</i>																																											
	<i>Notoplana australis</i>																	2	2																									
	<i>Polynoid polychaete</i>																																											
	<i>Sabellastarte australiensis</i>																																											
	<i>Terebellidae</i> spp.																																											
	Unidentified flatworm																																											
	Unidentified nemertean 1																																											
	Unidentified nemertean 2																																											
	Unidentified nemertean 3																																											
	Unidentified polychaete																																											
Sessile invertebrates	<i>Actinia tenebrosa</i>																																											
	<i>Anthothoe albocincta</i>	2																																										
	<i>Ascidia sydneyensis</i>																																											
	<i>Carajoa</i> sp.																																											
	<i>Herdmania grandis</i>																																											
	<i>Oulactis mucosa</i>																																											
	<i>Phlyctenanthus australis</i>																																											
	<i>Polycarpa viridis</i>																																											
	<i>Pyura australis</i>																																											
	<i>Pyura gibbosa</i>																																											
	<i>Pyura stolonifera</i>																																											
	<i>Sycozoa pulchra</i>																																											
	Unidentified anemone																																											
	Unidentified ascidians																																											
	Unidentified Pyrua sp.																																											
Species richness (total species per depth & habitat)		8	5	8	8	11	6	6	5	10	6	11	5	7	7	11	8	8	10	6	7	7	7	10	7	7	7	10	14	9	7	8	5	5	15	10	6	5	7	6	7	8	7	7
Total species richness (total species per site)		18				18				18				19				27					28					29					13					30						

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	10	11	12	13	14	15	16	17	18
	Depth category	-1	0	2	-1	0	0	2	-1	0
	Habitat	R	R	S	R	R	R	R	R	R
	Species / Replicate search	1	2	1	1	2	1	2	1	2
Barnacles	<i>Catomerus polymerus</i>	3			2	3			3	
	<i>Chamaesipho tasmanica</i>	4			5	5			4	5
Crustaceans	<i>Cithamalus antennatus</i>				4					
	<i>Elminius modestus</i>	4			4				5	
	<i>Tetracitella purpurascens</i>									
	<i>Bellidilia undecimspinosa</i>			2						
	<i>Carcinus maenas</i>									
	<i>Cyclograpsus granulosus</i>	3	4		2	2			2	3
	<i>Ebalia intermedia</i>									
	<i>Halicarcinus ovatus</i>									
	<i>Halicarcinus rostratus</i>			3						
	<i>Helograpsus haswellianus</i>							1	1	3
	<i>Jasus edwardsii</i>					2	2	3	2	
	<i>Metacarcinus novaezelandiae</i>					1				
	<i>Naxia aurita</i>						1			
	<i>Nectocarcinus integrifrons</i>									
	<i>Notomithrax ursus</i>									
	<i>Paguristes brevirostris</i>								1	
	<i>Pagurixus handrecki</i>									
	<i>Pagurus sinuatus</i>					1				
	<i>Palaemon dolospina</i>									
	<i>Palaemon intermedius</i>									
	<i>Palaemon serenus</i>		3				2	3		
	<i>Palaemon spp.</i>									
	<i>Paraciliacaea spp.</i>									
	<i>Paragrapsus gaimardii</i>					1				
	<i>Paragrapsus quadridentatus</i>				2					
	<i>Petrocheilus australiensis</i>									
	<i>Petrolisthes elongatus</i>	5	5	5	5	5	4	5	5	5
	<i>Petrolisthes spp.</i>									
	<i>Pilumnus tomentosus</i>									
	<i>Plagusia chabrus</i>									
	<i>Strigopagurus strigimanus</i>									
	<i>Allostichaster polyplax</i>									
	<i>Asterias amurensis</i>	2	2	5	5	4	4	3	4	
	<i>Astrostele scaber</i>									
	<i>Australostichopus mollis</i>				1	2				
	<i>Comanthus tasmaniae</i>									
	<i>Comanthus trichoptera</i>									
	<i>Coscinasterias muricata</i>		1		2					
	<i>Goniocidarid spp.</i>									
	<i>Heliocidarid erythrogramma</i>		2		2	3				
	<i>Holopneustes inflatus</i>									
	<i>Meridiastra calcar</i>									
	<i>Nectria ocellata</i>									
	<i>Ophiomyxa australis</i>									
	<i>Patiriella regularis</i>		5	5	4	4	3	4		
	<i>Petricia vernicina</i>		1							
	<i>Tosia australis</i>									
	<i>Tosia magnifica</i>									
	<i>Uniophora granifera</i>									

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	10	11	12	13	14	15	16	17	18
	Depth category	-1	0	1	2	3	-1	0	1	2
	Habitat	R	R	S	S	R	R	R	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1
Molluscs	<i>Afrolittorina praetermissa</i>									
	<i>Agnewia tritoniformis</i>									
	<i>Aphelodoris</i> spp.									
	<i>Aplysia sydneynensis</i>									
	<i>Austrocochlea constricta</i>									
	<i>Austrocochlea porcata</i>									
	<i>Austrolittorina unifasciata</i>									
	<i>Austromytilus rostratus</i>									
	<i>Bembicium auratum</i>									
	<i>Bembicium melanostomum</i>									
	<i>Bembicium nanum</i>									
	<i>Cabestana spengleri</i>									
	<i>Cabestana tabulata</i>									
	<i>Cellana solida</i>									
	<i>Charonia lampas rubicunda</i>									
	<i>Chiton glaucus</i>									
	<i>Chthamalus antennatus</i>									
	<i>Cominella lineolata</i>									
	<i>Crassostrea gigas</i>									
	<i>Dicathais orbita</i>									
	<i>Diloma concamerata</i>									
	<i>Electroma georgiana</i>									
	<i>Equichlamys bifrons</i>									
	<i>Fusinus noveahollandiae</i>									
	<i>Gibbomodiolia albicostus</i>									
	<i>Haliotis rubra</i>									
	<i>Hastula brazieri</i>									
	<i>Ischnochiton (Heterozona) cariosus</i>									
	<i>Ischnochiton australis</i>									
	<i>Ischnochiton elongatus</i>									
	<i>Ischnochiton variegatus</i>									
	<i>Lasaea australis</i>									
	<i>Limatula strangei</i>									
	<i>Maoricolpus roseus</i>									
	<i>Mitra carbonaria</i>									
	<i>Mitra glabra</i>									
	<i>Mytilus galloprovincialis</i>									
	<i>Nassarius nigellus</i>									
	<i>Notoacmea alta</i>									
	<i>Notoacmea petterdi</i>									
	<i>Octopus pallidus</i>									
	<i>Octopus tetricus</i>									
	<i>Ostrea angasi</i>									
	<i>Oxynocheilus</i>									
	<i>Patella (Scutellastra) peronii</i>									
	<i>Plaxiphora albida</i>									
	<i>Pleurobranchaea maculata</i>									
	<i>Polycera capensis</i>									
	<i>Rhyssoplax calliozona</i>									
	<i>Saccostrea glomerata</i>									
	<i>Sassia verrucosa</i>									
	<i>Scutus antipodes</i>									
	<i>Siphonaria diemenensis</i>									
	<i>Siphonaria funiculata</i>									
	<i>Siphonaria tasmanica</i>									
	<i>Sypharochiton pelliserpentis</i>									
	<i>Trichomya hirsuta</i>									
	<i>Turbo undulatus</i>									
	<i>Xenostrobus pulex</i>									

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	10		11		12		13				14		15		16		17				18																																	
	Depth category	-1	0			2		-1	0	1	2	3	-1	0	1	2	3	-1	0	1	2	-1	0																																
	Habitat	R	R	S	S			R	E	S	E	S	R	R	R	S	E	R	R	R	S	E	R	R																															
	Species / Replicate search	1	2	1	2	1	2	1	2	3	1	2	1	1	1	2	1	2	1	2	3	1	1	1	2	1	2	1	2	1	2	1	2	1	2	1	2																		
Other	<i>Eunice laticeps</i>																																																						
	<i>Galeolaria</i> spp.																																																						
	<i>Ikeda</i> spp.																1		2																1	2																			
	<i>Lepidonotus melanogrammus</i>																																																						
	<i>Myxicola infundibulum</i>											1						2		1																																			
	<i>Notoplana australis</i>																																																						
	<i>Polynoid polychaete</i>																																																						
	<i>Sabellastarte australiensis</i>																																																						
	<i>Terebellidae</i> spp.																																																						
	Unidentified flatworm																																																						
	Unidentified nemertean 1																																																						
	Unidentified nemertean 2																																																						
	Unidentified nemertean 3																																																						
	Unidentified polychaete																																																						
Sessile invertebrates	<i>Actinia tenebrosa</i>	1																																																					
	<i>Anthothoe albocincta</i>																																																						
	<i>Ascidia sydneyensis</i>																																																						
	<i>Carajoa</i> sp.																																																						
	<i>Herdmania grandis</i>																																																						
	<i>Oulactis mucosa</i>																																																						
	<i>Phlyctenanthus australis</i>																																																						
	<i>Polycarpa viridis</i>																																																						
	<i>Pyura australis</i>																																																						
	<i>Pyura gibbosa</i>																																																						
	<i>Pyura stolonifera</i>																																																						
	<i>Sycozoa pulchra</i>																																																						
	Unidentified anemone																																																						
	Unidentified ascidians																																																						
	Unidentified <i>Pyura</i> sp.																																																						
Species richness (total species per depth & habitat)		13	8	6	7	6	6	14	8	23	16	6	6	8	10	10	8	11	4	4	8	4	15	16	12	6	18	6	7	7	10	11	11	5	4	6	6	5	16	18	6	10	10	10	4	14	12	10	12	11	7	8	11	10	17
Total species richness (total species per site)		22				8				17				49				5				48				30				5				47				50																	

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 10-18) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	19	0	1	2	3	20	0	1	2	3	21	22	0	1	2	3	23	24
	Depth category	-1	0	1	2	3	-1	0	1	2	3	2	-1	0	1	2	3	2	-1
	Habitat	R	R	R	R	S	R	R	R	S	E	S	R	R	R	S	E	S	R
	Species / Replicate search	1	2	1	2	3	1	2	1	2	3	1	2	1	2	3	1	1	1
Barnacles	<i>Catomerus polymerus</i>			3	3						1		5	5					4
	<i>Chamaesipho tasmanica</i>	4	5				4	5					5	5					5
Crustaceans	<i>Citharus antennatus</i>																		
	<i>Elminius modestus</i>	3					3						3						5
	<i>Tetraclitella purpurascens</i>	3	4				3	3											3
	<i>Bellidilia undecimspinoso</i>																		
	<i>Carcinus maenas</i>																		
	<i>Cyclograpsus granulosus</i>		3					3					3						
	<i>Ebalia intermedia</i>															1			
	<i>Halicarcinus ovatus</i>															1			
	<i>Halicarcinus rostratus</i>															1		3	
	<i>Helograpsus haswellianus</i>																		
	<i>Jasus edwardsii</i>				4	3			2						2	2			2
	<i>Metacarcinus novaezelandiae</i>									1									
	<i>Naxia aurita</i>																		
	<i>Nectocarcinus integrifrons</i>																		
	<i>Notomithrax ursus</i>																		
	<i>Paguristes brevirostris</i>											3						2	
	<i>Pagurixus handrecki</i>																		
	<i>Pagurus sinuatus</i>																	1	
	<i>Palaemon dolospina</i>																		
	<i>Palaemon intermedius</i>																		
	<i>Palaemon serenus</i>					2													
	<i>Palaemon spp.</i>																		
	<i>Paracilicaca spp.</i>																		
	<i>Paragrapsus gaimardii</i>																		
	<i>Paragrapsus quadridentatus</i>	3																	
	<i>Petrocheles australiensis</i>																		
	<i>Petrolisthes elongatus</i>	2	4	5			3	5		3	5			4	5	1			5
	<i>Petrolisthes spp.</i>																		
	<i>Pilumnus tomentosus</i>																		
	<i>Plagusia chabrus</i>			1															1
Echinoderms	<i>Strigopagurus strigimanus</i>					1							1						1
	<i>Allostichaster polyplax</i>																		
	<i>Asterias amurensis</i>											1	4	3			3	3	3
	<i>Astrostele scaber</i>																		1
	<i>Australostichopus mollis</i>					1													2
	<i>Comanthus tasmaniae</i>																		
	<i>Comanthus trichoptera</i>				3	1		3	3		2				1				1
	<i>Coscinasterias muricata</i>				1					3		4					2		3
	<i>Goniocidaris spp.</i>																		
	<i>Heliocidaris erythrogramma</i>		1		3	2		2	1		2	3		2	3		2		1
	<i>Holopneustes inflatus</i>																		
	<i>Meridiastra calcar</i>		1					4		4	3	2	5						2
	<i>Nectria ocellata</i>					1						2							5
	<i>Ophiomyxa australis</i>				2														
	<i>Patiriella regularis</i>																		
	<i>Petricia vernicina</i>					1													
	<i>Tosia australis</i>																		
	<i>Tosia magnifica</i>											1							1
	<i>Uniophora granifera</i>									1						1			2

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 19-24) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site Depth category Habitat	19 -1 R	0 R	1 R	2 R	3 S	20 -1 R	0 R	1 R	2 S	3 E	21 2 S	22 -1 R	0 R	1 R	2 S	3 E	23 2 S	24 -1 R	0 R	1 R	2 S	3 E	
	Species / Replicate search	1	2	1	2	3	1	2	1	2	3	1	2	1	1	1	2	1	2	1	2	1	2	
Molluscs	<i>Afrolittorina praetermissa</i>						1																	
	<i>Agnewia tritoniformis</i>																							
	<i>Aphelodoris</i> spp.																							
	<i>Aplysia sydneyensis</i>											1					3	1						
	<i>Austrocochlea constricta</i>							3					2							2				
	<i>Austrocochlea porcata</i>							1												3				
	<i>Austrolittorina unifasciata</i>	4					5	5					4	3						5	5			
	<i>Austromytilus rostratus</i>																							
	<i>Bembicium auratum</i>	1																						
	<i>Bembicium melanostomum</i>																							
	<i>Bembicium nanum</i>																							
	<i>Cabestana spengleri</i>													2		2			2	3			3	3
	<i>Cabestana tabulata</i>																						2	
	<i>Cellana solida</i>	3	4						3					1	3						3	3		
	<i>Charonia lampas rubicunda</i>																		1				2	
	<i>Chiton glaucus</i>							2															2	
	<i>Chthamalus antennatus</i>			3		4	5																	
	<i>Cominella lineolata</i>								2						5						3			
	<i>Crassostrea gigas</i>							2	4					3	2						1	2	1	
	<i>Dicathais orbita</i>				3	2	3	2													3			
	<i>Diloma concamerata</i>	3	4						3		3										2	3	4	3
	<i>Electroma georgiana</i>																							
	<i>Equichlamys bifrons</i>												3											
	<i>Fusinus noveahollandiae</i>												1	3					3		2	2		
	<i>Gibbomodiola albicostus</i>																							
	<i>Haliotis rubra</i>									3														
	<i>Hasstula brazieri</i>																							
	<i>Ischnochiton (Heterozona) cariosus</i>																							
	<i>Ischnochiton australis</i>				2				3															
	<i>Ischnochiton elongatus</i>																							
	<i>Ischnochiton variegatus</i>																							
	<i>Lasaea australis</i>																							
	<i>Limatula strangei</i>																							

Appendix 17 cont.: Rank abundances of invertebrates recorded during timed surveys (sites 19-24) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	19	0					1	2					20	0					1	2					21	22	0					2	23	24	0					1	2				
	Depth category	-1												-1												2	-1							2	-1											
	Habitat	R												R												S	R						S	E	R						R					
	Species / Replicate search	1	2	1	2	3	1	2	1	2	1	2	1	1	2	1	2	3	1	2	1	1	1	2	1	2	1	2	1	1	2	3	1	1	1	2	1	2	1	2	1	2	1	1		
Other	<i>Eunice laticeps</i>																																													
	<i>Galeolaria spp.</i>	3																																												
	<i>Ikeda spp.</i>																																													
	<i>Lepidonotus melanogrammus</i>																																													
	<i>Myxicola infundibulum</i>																																													
	<i>Notoplana australis</i>																																													
	<i>Polynoid polychaete</i>																																													
	<i>Sabellastarte australiensis</i>																																													
	<i>Terebellidae spp.</i>																																													
	Unidentified flatworm																																													
	Unidentified nemertean 1																																													
	Unidentified nemertean 2																																													
	Unidentified nemertean 3																																													
	Unidentified polychaete																																													
Sessile invertebrates	<i>Actinia tenebrosa</i>	3				2								3	4			2									3																			
	<i>Anthothoe albocincta</i>													2																																
	<i>Ascidia sydneiensis</i>																																													
	<i>Carajoa sp.</i>																																													
	<i>Herdmania grandis</i>																																													
	<i>Oulactis mucosa</i>	2																																												
	<i>Phlyctenanthus australis</i>																																													
	<i>Polycarpa viridis</i>																																													
	<i>Pyura australis</i>																																													
	<i>Pyura gibbosa</i>																																													
	<i>Pyura stolonifera</i>													1																																
	<i>Sycozoa pulchra</i>																																													
	Unidentified anemone																																													
	Unidentified ascidians																																													
	Unidentified Pyura sp.																																													
Species richness (total species per depth & habitat)		16	10	6	9	8	10	6	9	5	2		14	20	4	3	3	12	14	4	8		9	10		10	20	7	8	11	0	9	7		9	6		16	14	10	6	15	14	3	13	
Total species richness (total species per site)		41											48											14	44											12	45									

Appendix 18: Rank abundances of fish recorded during timed surveys (sites 1-9) in February-March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 18 cont.: Rank abundances of fish recorded during timed surveys (sites 1-9) in February-March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	1	2	3	4	5	6	7	8	9
	Depth category	-1	0	1	1	0	0	0	-1	-1
	Habitat	R	R	R	R	R	R	R	R	R
	Species / Replicate search	1 2	1 2	1 2	1 2	1 2	1 2	1 2	1 2	1 2
Elasmobranchs	<i>Cephaloscyllium laticeps</i>									
	<i>Dasyatis brevicaudata</i>									
	<i>Dipturus whiteleyi</i>									
	<i>Urolophus cruciatus</i>									
	<i>Urolophus paucimaculatus</i>									
Species richness (total species per depth & habitat)		0 0 1 4	0 0 3 2	0 0 2 7	0 0 1 6	1 1 2 2 0 3 3 8 4 7 0	0 0 1 6 7 1	1 0 3 4 0 3 4 3 8 3 2	1 0 6 3	1 0 2 5 5 1 8 4 3 5
Total species richness (total species per site)		5	5	7	6	16	11	14	8	18

Appendix 18 cont.: Rank abundances of fish recorded during timed surveys (sites 10-18) in February-March 2010. Depth categories: -1= intertidal; 0= 0–0.9m; 1= 1–4.9m; 2= 5–8.9 m; 3= 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	10										11										12										13										14										15										16										17										18																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																													
	Depth category	-1		0				2		S		2		S		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3		R		-1		0		1		2		E		S		3	

Appendix 18 cont.: Rank abundances of fish recorded during timed surveys (sites 10-18) in February-March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	10	11	12	13	14	15	16	17	18			
	Depth category	-1	0	2	2	-1	0	2	2	-1	0	1	2
	Habitat	R	R	S	S	R	R	R	R	R	R	R	R
	Species / Replicate search	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2	1 2 1 2
Elasmobranchs	<i>Cephaloscyllium laticeps</i>												
	<i>Dasyatis brevicaudata</i>												1
	<i>Dipturus whitleyi</i>												
	<i>Urolophus cruciatus</i>			2	1								
	<i>Urolophus paucimaculatus</i>												1 2 2
Species richness (total species per depth & habitat)		0 0 3 4	1 4	1 5	0 0 3 3 9	3 8 7 2 4 7 5	0 0 2 5 9 9 4 3 7	1 1 1 2 3	9 12	1 0 0 7 9	4 6 10 4	1 0 4 0 8 10	5 5
Total species richness (total species per site)		5	4	5	26				26	6	18	21	21

Appendix 18 cont.: Rank abundances of fish recorded during timed surveys (sites 19-24) in February-March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 18 cont.: Rank abundances of fish recorded during timed surveys (sites 19-24) in February-March 2010. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m... Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2009/2010	Site	19									20									21	22									23	24												
	Depth category	-1			0			1			2			-1			0			1			2			2			-1			0			1			2					
	Habitat	R			R			R			R			S			R			S			E			S			R			R			R			S			E		
	Species / Replicate search	1	2	1	2	3	1	2	1	1	1	2	1	2	3	1	2	1	1	1	2	1	2	3	1	2	1	2	3	1	2	1	2	3	1	2	1	1					
Elasmobranchs	<i>Cephaloscyllium laticeps</i>																																										
	<i>Dasyatis brevicaudata</i>																																										
	<i>Dipturus whitleyi</i>																																										
	<i>Urolophus cruciatus</i>																																										
	<i>Urolophus paucimaculatus</i>																																										
Species richness (total species per depth & habitat)		0	0	2	4	2	14	9	8	6	0	0	8	1	4	8	6	6	0	3	5	0	0	0	1	5	11	8	1	7	4	4	1	1	7	1	3	8	5	0	6		
Total species richness (total species per site)		20									18									7	18									7	17												

Appendix 19: Rank abundances of fish recorded during timed surveys (sites 1-9) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	1	1	2	2	3	3	4	4	5	5	6	6	7	7	8	8	9	9
	Depth category	-1	1	-1	0	-1	1	-1	1	-1	0	-1	0	-1	0	-1	0	-1	0
	Habitat	R	R	R	E	R	R	R	R	R	R	R	R	R	R	R	R	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2	1	2
Elasmobranchs	<i>Cephaloscyllium laticeps</i>																		
	<i>Dasyatis brevicaudata</i>																		
	<i>Urolophus cruciatus</i>																		
Fish	<i>Acanthaluteres spilomelanurus</i>																		
	<i>Acanthaluteres vittiger</i>																		
	<i>Alabes dorsalis</i>																		
	<i>Aldrichetta forsteri</i>																		
	<i>Aplodactylus arctidens</i>																		
	<i>Aracana aurita</i>																		
	<i>Arenigobius bifrenatus</i>																		
	<i>Arripis georgianus</i>																		
	<i>Arripis trutta</i>																		
	<i>Atherinid spp.</i>																		
	<i>Atypichthys strigatus</i>																		
	<i>Brachaluteres jacksonianus</i>																		
	<i>Brachionichthys hirsutus</i>																		
	<i>Cheilodactylus spectabilis</i>																		
	<i>Conger verreauxi</i>																		
	<i>Creocele cardinalis</i>																		
	<i>Diodon nictemerus</i>																		
	<i>Dotalabrus aurantiacus</i>																		
	<i>Favonigobius lateralis</i>																		
	<i>Forsterygion gymnotum</i>																		
	<i>Forsterygion varium</i>																		
	<i>Gymnapistes marmoratus</i>																		
	<i>Heteroclinus johnstoni</i>																		
	<i>Heteroclinus tristis</i>																		
	<i>Hippocampus abdominalis</i>																		
	<i>Hippocampus bleekeri</i>																		
	<i>Latridopsis forsteri</i>																		
	<i>Lepidotrigla papilio</i>																		
	<i>Meuschenia australis</i>																		
	<i>Meuschenia freycineti</i>																		
	<i>Neoodax balteatus</i>																		
	<i>Nesogobius hinsbyi</i>																		
	<i>Nesogobius maccullochi</i>																		
	<i>Nesogobius spp.</i>																		
	<i>Notolabrus fucicola</i>																		
	<i>Notolabrus tetricus</i>																		
	<i>Pagrus auratus</i>																		
	<i>Parablennius tasmanianus</i>																		
	<i>Pempheris multiradiata</i>																		
	<i>Pictilabrus latilavivus</i>																		
	<i>Platycephalus spp.</i>																		
	<i>Platycephalus bassensis</i>																		
	<i>Platycephalus richardsoni</i>																		
	<i>Platycephalus speculator</i>																		
	<i>Pseudocaranx georgianus</i>																		
	<i>Pseudophycis bachus</i>																		
	<i>Rhombosolea tapirina</i>																		
	<i>Scorpaena papillosa</i>																		
	<i>Scorpius lineolata</i>																		
	<i>Siphonognathus attenuatus</i>																		
	<i>Siphonognathus beddomei</i>																		
	<i>Trachinops caudimaculatus</i>																		
	<i>Trachinops taeniatus</i>																		
	<i>Trachurus declivis</i>																		
	<i>Trinorfolkia clarkei</i>																		
	<i>Upeneichthys vlamingii</i>																		
	<i>Vincentia conspersa</i>																		
Mammals	<i>Arctocephalus pusillus</i>																		
Species richness (total species per depth & habitat)		0	0	2	2	0	0	1	0	0	0	5	3	0	0	3	2	0	0
Total species richness (total species per site)		4				1				8				13		12		15	

Appendix 19 cont.: Rank abundances of fish recorded during timed surveys (sites 10-18) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1 = 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

2024	Site	10	0	11	12	13	14	15	16	17	18
	Depth category	-1	0	1	2	-1	0	1	2	-1	0
	Habitat	R	R	S	S	R	R	R	S	R	R
	Species / Replicate search	1	2	1	2	1	2	1	2	1	2
Elasmobranchs	<i>Cephaloscyllium laticeps</i>										
	<i>Dasyatis brevicaudata</i>										
Fish	<i>Urolophus cruciatus</i>			1							
	<i>Acanthaluteres spilomelanurus</i>					2	1	1	2		1
	<i>Acanthaluteres vittiger</i>		2		2	3	1	2			
	<i>Alabes dorsalis</i>							2	3		
	<i>Aldrichetta forsteri</i>										
	<i>Aplodactylus arctidens</i>									3	
	<i>Aracana aurita</i>									1	
	<i>Arenigobius bifrenatus</i>									2	
	<i>Arripis georgianus</i>							3			
	<i>Arripis trutta</i>										
	<i>Atherinid spp.</i>										
	<i>Atypichthys strigatus</i>										3
	<i>Brachaluteres jacksonianus</i>										
	<i>Brachionichthys hirsutus</i>				1				1		
	<i>Cheilodactylus spectabilis</i>										1
	<i>Conger verreauxi</i>										
	<i>Creocele cardinalis</i>										1
	<i>Diodon nichemerus</i>		1			2	1				1
	<i>Dotalabrus aurantiacus</i>		1							2	
	<i>Favonigobius lateralis</i>										1
	<i>Forsterygion gymnotum</i>					2				1	
	<i>Forsterygion varium</i>		3	2	2	2	3	3	2	3	
	<i>Gymnapistes marmoratus</i>					2	3	3	2	2	3
	<i>Heteroclinus johnstoni</i>										
	<i>Heteroclinus tristis</i>					1					
	<i>Hippocampus abdominalis</i>								1	1	
	<i>Hippocampus bleekeri</i>				1						
	<i>Laridopsis forsteri</i>										1
	<i>Lepidogrigla papilio</i>			1							
	<i>Meuschenia australis</i>		1								
	<i>Meuschenia freycineti</i>										
	<i>Neodax balteatus</i>		2		3	2					2
	<i>Nesogobius hinsbyi</i>					2	3	3	2		
	<i>Nesogobius maccullochi</i>			3	1	2	1				
	<i>Nesogobius spp.</i>							2			3
	<i>Notolabrus fucicola</i>										
	<i>Notolabrus tetricus</i>		1			1	2	2	2		
	<i>Pagrus auratus</i>									1	
	<i>Parablennius tasmanianus</i>		1	1							
	<i>Pempheris multiradiata</i>					2				2	
	<i>Pictilabrus laticlavus</i>		2			2	2				
	<i>Platycephalid spp.</i>										
	<i>Platycephalus bassensis</i>			1	2					2	
	<i>Platycephalus richardsoni</i>										
	<i>Platycephalus speculator</i>			3							3
	<i>Pseudocaranx georgianus</i>										
	<i>Pseudophycis bachus</i>							1			
	<i>Rhombosolea tapirina</i>										1
	<i>Scorpaena papillosa</i>										
	<i>Scorpius lineolata</i>										
	<i>Siphonognathus attenuatus</i>										
	<i>Siphonognathus beddomei</i>										
	<i>Trachinops caudimaculatus</i>		2			1	3				
	<i>Trachinops taeniatus</i>										
	<i>Trachurus declivis</i>										
	<i>Trinorfolkia clarkei</i>					3					
	<i>Upeneichthys vlamingii</i>										
	<i>Vincentia conspersa</i>					1					
Mammals	<i>Arctocephalus pusillus</i>							1			
Species richness (total species per depth & habitat)		0	0	9	3	5	2	4	5	0	0
Total species richness (total species per site)		10		5	6			22		14	

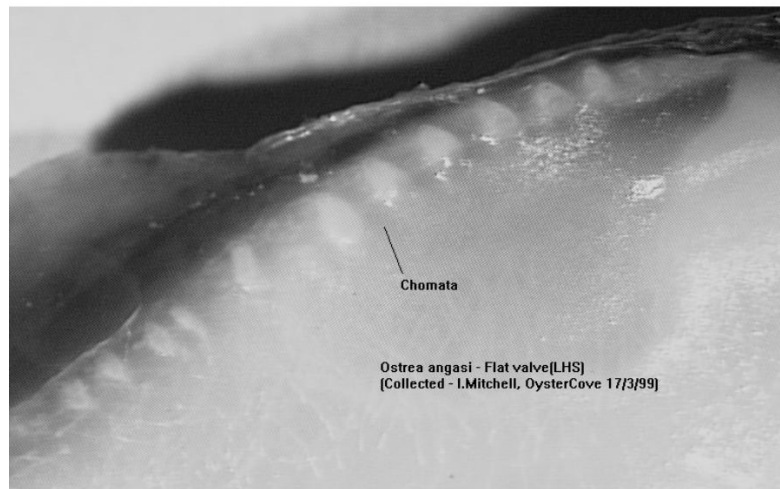
Appendix 19 cont.: Rank abundances of fish recorded during timed surveys (sites 19-24) in 2023-2024. Depth categories: -1= intertidal; 0 = 0–0.9m; 1= 1–4.9m; 2 = 5–8.9 m; 3 = 9–15m. Habitat categories: R = reef; E = ecotone (reef-sand); S = sand.

[illegible]

Appendix 20. Results of oyster samples randomly collected for taxonomic validation. Depth code is -1 = intertidal, 0 = 0-1 m depth, 1 = 1-5 m depth, 2 = 5-10 m depth.

Site	Depth code	<i>Ostrea angasi</i>	<i>Crasostrea gigas</i>	Intermediate
Cadbury Point	1	0	15	
Bedlam Walls north	0	2	15	
Bedlam Walls north	1	3	6	2
Blackman's Bay Blowhole	-1	0	7	
Blackman's Bay Blowhole	0	1	6	
Blackman's Bay Blowhole	1	6	0	
Blackman's Bay Blowhole	2	3	6	2
White Rock	-1	0	11	
White Rock	0	0	2	
White Rock	1	1	8	
White Rock	2	2	9	
Tranmere Point	-1	0	4	
Tranmere Point	0	0	4	
Tranmere Point	1	2	0	
Tranmere Point	2	NA	NA	
Crayfish Point	-1	NA	NA	
Crayfish Point	0	0	2	
Crayfish Point	1	2	3	
Crayfish Point	2	3	5	
Alum Cliffs	-1	0	7	
Alum Cliffs	0	NA	NA	
Alum Cliffs	1	1	5	
Alum Cliffs	2	0	10	
Rosny Point	-1	0	5	
Rosny Point	0	0	5	
Rosny Point	1	2	3	
Rosny Point	2	NA	NA	
Geilston	-1	0	7	
Geilston	0	0	5	
Geilston	1	NA	NA	
Geilston	2	NA	NA	
Bellerive Bluff	-1	0	6	
Bellerive Bluff	0	5	0	1
Bellerive Bluff	1	0	3	1
Bellerive Bluff	2	NA	NA	
Sum of samples		33	163	

Appendix 21. Diagnostic features for differentiating between *Crassostrea gigas* and *Ostrea angasi*. Top image shows distinct hinge crenulations visible on the flat valve of *C. gigas* adjacent to the centre of the ligament, visible internally once the shell is opened. The lower image shows distinct chomata (teeth) on the flat valve of *O. angasi*, adjacent to the ligament once the shell is opened. Image credit to Iona Mitchell.



Appendix 22. Results of *O. angasi* specific taqman qPCR test was performed in triplicate for each sample taken. Depth code is -1 = intertidal, 0 = 0-1 m depth, 1 = 1-5 m depth, 2 = 5-10 m depth.

Site	Depth	Morph ID	Notes	Angasi qPCR result	Mean cq	Stdev cq
Crayfish Point	1	Angasi	Molecular ID uncertain	Equivocal	NA	NA
Rosny Point	1	Angasi	Molecular ID uncertain	Equivocal	NA	NA
Crayfish Point	2	Angasi	Molecular ID uncertain	Equivocal	NA	NA
Bedlam Walls	2	Angasi	Morph ID doesn't match molecular ID	Negative	NA	NA
Bellerive Bluff	0	Angasi	Morph ID is same as molecular ID	Positive	27.49794	0.081604
Bedlam Walls	2	Angasi	Morph ID is same as molecular ID	Positive	27.15422	0.042087
Bellerive Bluff	0	Angasi	Morph ID is same as molecular ID	Positive	27.43168	0.218543
Rosny Point	1	Angasi	Morph ID is same as molecular ID	Positive	25.98565	0.288709
Tranmere Point	1	Angasi	Morph ID is same as molecular ID	Positive	27.07966	0.230846
White Rock	1	Angasi	Morph ID is same as molecular ID	Positive	27.78405	0.177997
White Rock	2	Angasi	Morph ID is same as molecular ID	Positive	26.70605	0.205183
Bedlam Walls	0	Angasi	Morph ID is same as molecular ID	Positive	26.80519	0.158581
Blackman's Bay Blowhole	2	Pacific	Molecular ID uncertain	Inhibited	NA	NA
Geilston	0	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Rosny Point	0	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Rosny Point	-1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bedlam Walls	2	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bedlam Walls	0	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Alum Cliffs	-1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bellerive Bluff	-1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Geilston	-1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Tranmere Point	-1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Cadbury Point	1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Cadbury Point	1	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bedlam Walls	0	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bedlam Walls	0	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Bedlam Walls	2	Pacific	Morph ID is same as molecular ID	Negative	NA	NA
Blackman's Bay Blowhole	2	Pacific	Morph ID doesn't match molecular ID	Positive	26.52947	0.279779
Bedlam Walls	2	Unknown	Morph ID uncertain	Negative	NA	NA
Bellerive Bluff	1	Unknown	Morph ID uncertain	Positive	26.52709	0.159374
Bellerive Bluff	0	Unknown	Morph ID uncertain	Positive	26.85987	0.20236
Bedlam Walls	2	Unknown	Morph ID uncertain	Positive	26.57376	0.224183